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Title: Options for producing a warm-water fish in the UK: Limits to "Green Growth"?

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Abstract: This paper explores the development of a sustainable production system for tilapia and the research implications involved with ensuring commercial viability of such a system for UK farmers. The tilapia is a warm water fish with firm texture, white flesh and mild taste quite similar to a cod or haddock. Whilst tropical in origin it is thought to be highly suitable for low cost aquaculture in temperate zones with the potential to be a more sustainable source of food with fewer environmental impacts than other substitutes. Drawing on a literature review and findings from technical trials the paper will review and compare two production systems - novel Activated Suspension Technology (AST) and conventional Recirculating Aquaculture Systems (RAS) - considering their feasibility in terms of potential and financial viability for scaling up to commercial production of tilapia and their environmental and sustainability benefits.

The review concludes that AST based only on microbial floc is currently uncompetitive with RAS in a UK context although the approach has benefits that might be incorporated in a new generation of mixed systems. Refinement of such systems needs to occur with potential adopters and could be part of diversification of mixed farms. Such development might further enhance the ethical values of fish produced in small-scale, modular RAS.

Table 1 Water and land productivity of tilapia RAS compared to selected intensive open-water production systems (after Phillips *et al.*, 1991 reported in Timmons *et al* 2002).

		Water productivity kg m ³⁻¹	Production intensity mt ha ⁻¹ yr ⁻¹	Ratio of land and water use to RAS use	
				Water	Land
RAS	Nile tilapia	10	1,340	1	1
Intensive ponds	Nile tilapia	0.05	17.4	200	77
	<i>Paneid</i> shrimp	0.05 - 0.09	4.2-11	110-200	120-320
	Channel catfish	0.2-0.3	3	400-500	446
Raceways	Rainbow trout	0.005	150	2,100	9

Table 2. Comparison of production parameters in experimental RAS and AST grow-out systems for *O. niloticus* (source: Murray et al. 2007)

Parameter	Source			
	Murray <i>et al</i> 2007	Hargreaves 2006	Rackocy 2002	Avnimelech 1999
Production system	RAS	AST	AST	AST
Indoor/ Outdoor	indoor	indoor	outdoor	outdoor
Dietary crude protein %	30.4	32	32	20
Secondary carbohydrate source	na	none	none	cellulose
Solids management (TSS mg l ⁻¹)	< 2	250-1000	898 (100-1960)	no data
Culture unit	2.8m ³ tanks	1.5 m ³ tanks	200 m ³ tank	50m ² earthen pond
Culture days	107	no data	201	30
Mean temperature (°C)	29.4	no data	28.5	-
Mean start weight (g)	19.1	41	73.6	112
Mean end weight (g)	405	134	678	218
Cumulative SGR (%) ^b	2.8	1.27	1.11 ^a	1.31 ^a
Final biomass kg m ³⁻¹	25.6	9.8	13.7	16.5
Cumulative FCR ^c	1.1	1.83	1.9	2.17
Cumulative PCR ^d	3.1	no data	-	2.18
Survival	98.2	no data	81	94.8
Water productivity kg m ³⁻¹	8.1	no data	9.7	no data

^a Extrapolated from start and end weights; ^b Specific growth rate; ^c Food Conversion ratio ^d Protein conversion ratio

1 **REVIEW: Options for producing a warm-water fish in the UK: Limits to**
2 **“Green Growth”?**

3
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9
10 *This paper explores the development of a sustainable production system for tilapia*
11 *and the research implications involved with ensuring commercial viability of such a*
12 *system for UK farmers. The tilapia is a warm water fish with firm texture, white flesh*
13 *and mild taste quite similar to a cod or haddock. Whilst tropical in origin it is thought*
14 *to be highly suitable for low cost aquaculture in temperate zones with the potential to*
15 *be a more sustainable source of food with fewer environmental impacts than other*
16 *substitutes. Drawing on a literature review and findings from technical trials the*
17 *paper will review and compare two production systems - novel Activated Suspension*
18 *Technology (AST) and conventional Recirculating Aquaculture Systems (RAS) -*
19 *considering their feasibility in terms of potential and financial viability for scaling up*
20 *to commercial production of tilapia and their environmental and sustainability*
21 *benefits.*

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23 *uncompetitive with RAS in a UK context although the approach has benefits that*
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26 *mixed farms. Such development might further enhance the ethical values of fish*
27 *produced in small-scale, modular RAS.*

28
29
30 **Introduction**

31
32 Seafood consumption in the UK is on the rise (Seafish, 2006), although in comparison
33 with other countries the amounts consumed are relatively small. In recent years a
34 high media profile has affected UK consumers' interest in and perceptions of seafood.
35 Some of the many issues making headlines range from the health benefits of including
36 fish in the diet (Britton, 2006), to concerns with the safety of consuming both wild
37 and farmed fish (Foran, Carpenter, Hamilton, Knuth & Schwager, 2005). Declining
38 wild fish stocks (Worm et al., 2006) and the quality of the marine environment (Royal
39 Commission on Environmental Pollution, 2004) are also frequently brought to the
40 public eye, creating a complex picture for the public. The diversity of contradictory
41 messages received by the public instigates confusion (Young, Grady, Little,
42 Watterson & Murray, 2006).

43
44 Most of the fish used for human consumption currently comes from wild capture
45 fisheries; however seafood from aquaculture is growing rapidly and is set to account
46 for 50% of the world's food fish in the near future (FAO, 2007a). The rapid growth in
47 aquaculture production is attributed to declining wild stocks even as these continue to
48 be exploited for use in feed for farmed carnivorous aquatic species as well as for other
49 forms of intensive livestock production. This is a major cause of controversy (White,
50 O'Neill & Tzankova, 2004)

51

52 The largely negative view of aquaculture in the UK, as a highly intensive, specialised
53 and vertically integrated business model contrasts with traditional practice elsewhere
54 around the world. In Asia where global aquaculture remains concentrated, low trophic
55 species such as carps and tilapias still dominate farmed production and much of this is
56 based on pond-based semi-intensive or extensive systems. This type of aquatic
57 farming is characterised by the high proportion of feed being produced through
58 natural food webs *in situ* (Azim & Little, 2006). Traditionally aquaculture was one
59 component of mixed farming systems and geared to meet subsistence and local
60 market needs (Beveridge & Little, 2002). But soaring demand and limitations of these
61 systems has fuelled a major scale-up in the world wide production of farmed
62 'seafood' over the last two decades both to meet local and, increasingly, international
63 markets. The shrimp boom in the mid-1980s-90s based on a limited number of species
64 (mainly *Penaeus* spp.) and more latterly Nile tilapia (*Oreochromis niloticus*) and
65 Asian river catfish (*Pangasius hypophthalmus*) have both spread and intensified,
66 particularly in developing countries where land, water and labour are abundant and
67 cheap.

68

69 Tilapias have been heralded as a seafood commodity with major potential (Josupeit,
70 2005). In contrast to shrimp production the rapid scale-up in tilapia production has
71 attracted little criticism from environmental groups and instead been portrayed as a
72 white fish alternative to species higher up the food chain (Marine Conservation
73 Society, 2006). They are being produced in a wide range of production systems and
74 countries in the Tropics and Sub-tropics unlike the Asian river catfish where
75 significant production is concentrated in one area-the Mekong Delta of Vietnam. It
76 might be argued that these factors increase the relative opportunity for sustained
77 growth of tilapia production, especially as despite the levels of growth, prices have
78 remained relatively firm (Josupeit, 2005).

79

80 Global production of tilapias has soared over the last decade (FAO, 2007b) with
81 particularly significant growth in South America for export markets and China for
82 both internal and export markets (Josupeit, 2007a, b). This tropical species that
83 originated in Africa is now the 6th most popular seafood choice in the USA (National
84 Fisheries Institute, 2005) and major aquaculture producers turn to tilapia as a new
85 species to invest in (Josupeit, 2007b). Although major centres of tilapia production
86 are in Asia, South and Central America and Africa, culture has also become
87 established in North America and Europe in the last few years. Tilapia production in
88 the UK has been mainly characterised by high profile failures to date (Bunting &
89 Little, 2005). This review assesses the technological options for tilapia production
90 within insulated agricultural buildings proposed as a potential option for rural
91 diversification (Little, 2006).

92

93

94 **Towards greener aquaculture**

95

96 In light of the contradictory messages conveyed in the media, consumer
97 understanding of the ethical and human health issues surrounding aquaculture is
98 understandably confused; however there is still a strong desire for fresh, traceable fish
99 amongst UK consumers (Young et al, 2006). The natural shoaling behaviour of many
100 fish species make the farming of fish at high density both practical and ethical (>100

101 kg m⁻³) provided that nutritionally balanced diets can be cost effectively delivered
102 and the quality of the water can be maintained (Ebeling, Timmons & Bisogni, 2006).

103

104 Most of the fish species raised intensively are top carnivores most dependent on high
105 quality feeds conventionally based on fishmeal and oils derived from wild fisheries.
106 These feed ingredients are subject to contamination with persistent organic
107 compounds and their amplification through the food chain (Worm et al, 2006). The
108 relative risk of consumption of such farmed fish compared to fish of wild origin and
109 other food stuffs for different groups is the focus of increasing consumer and
110 scientific interest (Foran, Good, Carpenter, Hamilton, Knuth & Schwager, 2005;
111 Ellingsen & Aanondsen, 2006).

112

113 The environmental costs of feed and water supply to aquaculture are also becoming a
114 major cause of criticism (Naylor et al., 2000); particularly for carnivorous species but
115 intensification of low tropic species such as tilapias and carps is also utilising
116 increasing amounts of such feeds. So called 'flow through' or 'open' intensive
117 systems, in which there is little or no water re-use, can be highly polluting on
118 receiving waters partly because the cost effective removal of dilute soluble nutrients is
119 problematic. Open systems includes raceways and cages that produce most of the
120 tilapias traded internationally (Coward & Little, 2001).

121

122 Rapid global growth in farming fish and shrimp has occurred in tandem with strong
123 commercial and environmental incentives to reduce the costs of feed and the impact
124 of effluents respectively. Any review of the short history of aquaculture will illustrate
125 that intensification is based on increasing the density of stocked animals stimulating
126 increased use of both water exchange (to maintain water quality) and higher protein
127 feeds. High water exchange aggravates nutrient loss and restricts opportunities for
128 recycling these expensive inputs; it has also been linked to poor biosecurity and
129 spread of disease in shrimp culture. This has caused a paradigm shift in recent years
130 towards use of lower protein feeds in low exchange, green water systems, initially in
131 shrimp production (McIntosh, 2000), but increasingly for other species. This
132 approach appears to be particularly attractive for systems based on low trophic species
133 such as the tilapias.

134

135 Recirculating aquaculture systems (RAS) are increasingly common land based
136 systems in temperate countries in which water is reused after removal of waste
137 nutrients and heat may be cost effectively retained in the system. The mechanisms for
138 removal of suspended and dissolved wastes to reduce solids and nitrogenous
139 compounds hazardous to fish are key parameters of RAS. There is no requirement for
140 continual discharge of effluents into the environment, as is the case with the majority
141 of conventional flow-through aquaculture systems. The 'price' to pay is the external
142 energy cost of moving water through an appropriate water treatment system,
143 temperature control, provision of adequate dissolved oxygen and need for complete
144 balanced nutrition. The commercial culture of tilapia in RAS is now established in
145 North America and parts of Europe as specialised enterprises targeting high value
146 markets. Such operations have been either based on integration with waste heat or as
147 stand-alone enterprises (Melard & Philippart, 1981; Bunting & Little, 2005). This
148 factor together with the fact that they can produce food locally with few effluents
149 suggests they meet some of the criteria of 'green' food production systems. The recent
150 history of limited RAS development in the UK suggests that production technologies

151 and markets are undeveloped; it is however established practice for value-added
152 aquaculture such as accelerated production of juveniles for on-growing in open
153 systems or ornamental production. Elsewhere in Europe they have become more
154 established for catfish and eel production supplying diverse ethnic and cultural
155 markets (Eding & Kamstra, 2002). More limited access to water for UK and European
156 aquaculture and growing regulation on effluents is likely to increase the attraction of
157 RAS. Rapid production cycles for warmwater fish are also attractive. Tilapias can
158 reach marketable size in as little as six months while 18-24 months is the norm for
159 UK farmed rainbow trout (*Oncorhynchus mykiss*) or Atlantic salmon (*Salmo salar*). The
160 extent of potential water and land productivity gains for tilapia cultured in RAS
161 compared with other intensive production systems for warm-water and temperate
162 species are highlighted in Table 1.

163
164

165 **Why tilapia, why intensive??**

166

167 Tilapias have many characteristics amenable to farming, such as its fast growth under
168 a range of conditions, resilience against disease and a flavour and texture comparable
169 with valuable marine fish (Beveridge & McAndrew, 2000). An ability to feed low in
170 the food chain in principle means that production costs can be low but also,
171 importantly, the fish can be marketed to appeal to increasingly informed consumers
172 on environmental and broader ethical grounds. The trends towards more intensive
173 practices by most commercial tilapia producers threatens these potential core
174 advantages but is a response to current commercial realities.

175

176 An ability to feed low in the food chain is matched by a high responsiveness to
177 intensification such that tilapias perform well in intensive systems based on complete,
178 but relatively low protein diets. Their tolerance of high densities (lower densities in
179 fact often trigger aggressive territorial behaviour) has meant a rapid uptake of more
180 intensive operations including more use of higher quality supplementary feeds in
181 semi-intensive ponds (Edwards, Yakupitiyage & Lin, 2000) or complete, formulated
182 feeds in intensive systems. Typically only 20-25% of fed protein is retained in the fish
183 raised in intensive systems (Avnimelech, 2006) the balance becoming pollutants that
184 must be removed. In principle if these waste nutrients could be retained in the system
185 they become substrate for protein-rich bacteria that are re-ingested and utilised by the
186 tilapia. Such nutrient recovery *in situ* occurs in conventional ponds but can be
187 operated at a higher level of intensity through use of aeration to maintain microbial
188 floc in suspension. These activated suspension ponds or technology (ASP, AST) have
189 been advocated for both tilapias and shrimp (Avnimelech, Kochva & Diab, 1994;
190 McIntosh, 2000). The nutritional value of such microbial floc to aquatic animals is
191 dependent on several factors: food preference, ability to both ingest and digest it but
192 also the density of the suspended particles (Hargreaves, 2006). Tilapias being both
193 capable of filter feeding and detritivory are ideal candidates for such systems
194 (Dempster, Baird & Beveridge, 1995; Azim, Verdegem, Mantingh, van Dam &
195 Beveridge, 2003).

196

197 Potentially the relative operational simplicity of AST can be combined with a
198 production intensity that is economically viable in the context of a diversification
199 option for mixed farms in the UK. Moreover the 'green' characteristics of the
200 approach could be favourable especially as the market for premium ethical food of all

201 types has developed rapidly but is under-supplied by local producers. The theoretical
202 basis for the AST is now considered before its application in a UK context is assessed.

203

204

205 **From feeding to floc**

206

207 Intensification of aquaculture systems imposes two major technical challenges-the
208 maintenance of dissolved oxygen and the removal of inorganic nitrogenous products.

209 The latter is critical within intensive aquaculture systems as even low levels of
210 unionized ammonia in water are toxic to most cultured species (Timmons,
211 Ebeling, Wheaton, Summerfelt & Vinci, 2002). Oxygen levels typically become the
212 limiting production factor in optimised culture-systems with adequate ammonia/
213 nitrite treatment capacity.

214

215 There are three principle nitrogen pathways to remove hazardous N species in
216 aquaculture (1) photo-autotrophic removal by algae, (2) immobilisation by
217 heterotrophic bacteria as proteinacious microbial biomass and (3) chemo-autotrophic
218 oxidation to nitrate by ‘nitrifying’ bacteria (Ebeling et al., 2006). The relative
219 importance of each varies with system type and production intensity. Hargreaves
220 (2006) distinguishes between ‘photosynthetic growth’ (PSG) and ‘mixed suspended
221 growth’ systems (MSG) based on the degree to which water quality is maintained by
222 photosynthetic and bacterial processes. Suspended particulates formed by
223 heterotrophic bacteria also provide efficient substrates for nitrifying bacteria in bio-
224 floc systems. These can be visualised as ‘green’ and ‘brown’ water systems.

225

226 Suspended-growth systems are further differentiated from ‘attached growth’ systems
227 as the waste assimilation, recycling and food production occur within the culture unit
228 as opposed to external bio-filters. Most aquaculture occurs in earthen ponds; which
229 can be considered as PSG systems. Conversely, most RAS rely primarily on chemo-
230 autotrophic bacteria attached as aerobic bio-films on filter media. These examples
231 reflect opposing management goals; in attached systems the aim is to remove nitrogen
232 from the system. In suspended systems the aim is to conserve and recycle nitrogen as
233 useful microbial biomass. Suspended growth systems have also been referred to by a
234 range of terms based on biological or containment characteristics: activated
235 suspension ponds (ASP) activated suspension technology (AST), bio-floc technology
236 (BFT), organic detrital algae soup (ODAS) etc.

237

238 Intensification of any suspended growth system requires oxygenation and good water
239 mixing to increase the rate of ammonia immobilisation, both of which can be
240 achieved simultaneously through vigorous aeration. Phytoplankton-rich systems will
241 also benefit from *in-situ* oxygen generation, but with intensification they will
242 ultimately become light limited through self shading. Thus sustained aeration and
243 mixing are essential requirements for intensification of both green and brown water
244 systems.

245

246 Although few cross-references exist in the literature these processes are also the basis
247 of the ‘activated-sludge’ sewage treatment process (Ganczarczyk, 1983; Thiel, 2002).
248 The main difference is that bio-floc accumulations in sewage treatment systems are
249 periodically settled and voided in a continuous or semi-continuous process. In closed-
250 AST the goal is to conserve bio-floc as a food source through internal nutrient

251 recycling. This mode of operation has two further beneficial features. Theoretically,
252 water exchange rates can be reduced compared to conventional RAS, which are
253 themselves conservative consumers of water (Table 1). Secondly, accumulation of
254 waste inorganic nitrogen compounds; unionised ammonia and nitrite (NH_3 and NO_2)
255 will result in growth inhibition or mortality of fish. *In-situ* heterotrophic ammonia and
256 nitrite assimilation therefore also conserves water quality in this vital respect.

257
258 These attributes provided the impetus behind two major trends in the development
259 and application of microbial bio-floc systems in aquaculture. The first has its origins
260 in attempts to optimise natural feed production in semi-intensive ponds through
261 various types of bio-manipulation. The second has its basis in the 'zero-water
262 exchange' and water quality remediation possibilities of AST in contexts where water
263 conservation is paramount. This driver had two threads. Researchers in Israel assessed
264 AST as a potential means of simultaneously intensifying yields and water productivity
265 in arid environments. Elsewhere, the same AST features, offered a means of
266 addressing bio-security and environmental concerns associated with shrimp
267 production. The development of intensive 'zero exchange' shrimp systems provided a
268 highly effective means of disease and effluent management (Burford, Thompson,
269 McIntosh, Bauman & Pearson, 2004, Hari, Kurup, Varghese, Schrama, &
270 Verdegem, 2005; Lemonnier & Faninoz, 2006; Samocha et al., 2007) with feed
271 optimisation as a secondary benefit (Burford, Thompson, McIntosh, Bauman &
272 Pearson, 2003; Wasielesky, Atwood, Stokes & Browdy, 2006). The concurrent
273 evolution of these two drivers is considered below.

274
275 The limits of natural productivity in ponds were initially explored using input: output
276 work (e.g. Schroeder, 1978) based on the premise that light-limited primary
277 productivity of conventional shallow ponds in the Tropics of $30\text{kg ha}^{-1} \text{d}^{-1}$ could be
278 further enhanced by optimising heterotrophic productivity through addition of carbon
279 rich substrates. Initially, this approach assumed that photo-autotrophic and
280 heterotrophic feed pathways were partitioned and emphasised the role of
281 heterotrophic pathways in achieving further yield gains. However, the
282 interdependence of these pathways and the mechanisms by which fish such as tilapia
283 could filter feed or harvest micro-organisms from the water column soon became
284 apparent (Colman & Edwards, 1987; Avnimelech, Mokady & Schroeder, 1989).

285
286 Concurrent work carried out in Israel on more intensive systems suggested that
287 sorghum and other energy-rich grains could be used cost effectively as supplements to
288 natural food-especially micro-algae rather than more protein-rich feeds (Hepher,
289 1988). Yields in these intensive water-limited systems, were constrained by water
290 quality limits stimulating further work aimed at enhancing AST function.
291 Theoretically, optimising ratios of C:N will enhance conversion of toxic inorganic-
292 nitrogen compounds to microbial biomass available as food for fish or shrimp while
293 further improving water quality. Goldman et al. (1987) elucidated the fundamental
294 nutrient balance principles underlying growth efficiency of marine bacteria. They
295 found C:N ratios $>10:1$ were optimal for optimising bio-floc production while
296 minimising ammonia regeneration. Many investigators (Avnimelech et al, 1989, 1994,
297 1999, Hari et al., 2004, Burford et al, 2004) then applied this principle as an approach
298 to optimising nutrient inputs and recycling within intensive bio-floc aquaculture
299 systems. The use of a carbohydrate source in addition to conventional feeds or use of

300 feeds with lower protein content was advocated on this basis for systems in which
301 bio-floc was aerated and retained in the system (Avnimelech, 1999).

302
303 The AST concept of further intensification of natural food production and use *in situ*
304 has developed from this practice and theory for species such as tilapias and shrimp
305 that are capable of utilising microbial floc as a major element of the diet and tolerating
306 water high in suspended solids. Higher intensification rates also involve a move from
307 earthen pond systems to lined-pond systems (shrimp) and tanks (tilapia). Most
308 published accounts of AST however, relate to systems which maintain algal-rich
309 water i.e. green water / PSG systems.

310
311 Generally green water systems are known to suffer inconsistent water quality, partly
312 related to algal succession that is difficult to control or influence. Bacteria dominated
313 systems tend to be more consistent (Hargreaves, 2006) but the nature and impacts of
314 succession and change within systems with minimal phytoplankton are unknown. Our
315 understanding of low-plankton systems is informed by the experience of managing
316 partitioned aquaculture systems that alternate between autotrophic and heterotrophic
317 status depending on ambient climate (Hargreaves, 2006). The principle of using
318 compartments of algal rich water to remove ammonia is complicated by mixed
319 success in controlling algal biomass.

320
321 The adaptation of these principles to a brown water / MSG system in light-limited
322 conditions in which natural feed was mainly bacterial rather than derived from
323 phytoplankton was the major objective of our research. The relative stability of
324 heterotrophic microbial populations and their independence of light conditions on
325 water quality were considered as positive factors (Avnimelech, 2006). For the
326 sunlight-limited seasonal conditions in the UK the concept of well insulated smaller
327 intensive tank-based systems located inside buildings was developed based on such a
328 'brown-water' approach. We now consider some of fundamental issues that
329 differentiate AST as researched and promoted to date with their potential for use
330 within the farming sector in the UK.

331
332

333 **Tilapia as a farm diversification strategy in the UK**

334
335

336 Intensive fish culture in the UK has been the preserve of an entrepreneurial business
337 sector and the attraction of this type of diversification for risk-adverse farmers must
338 be considered (Rosa, Kodithuwakku, Young & Little, 2007). Diversification into a
339 novel product (i.e. tilapia) based on a new technical approach (AST) is likely to
340 further increase risk. The potential benefits of using AST for tilapia production rather
341 than RAS scaled down to meet the investment profiles and potential local market
342 niches available to them need to be established.

343

344 The costs and risks of maintaining optimal temperatures for warm-water fish are an
345 initial concern to most potential adopters. The optimal temperature range for tilapia
346 production is 28-32°C, however, energy costs (heating and pumping) are
347 proportionately low (15% total direct costs; Timmons, 2005). In the past, RAS have
348 often been linked to waste heat utilisation from distilleries, power stations, factories
349 etc. Whilst an apparently green and cost-effective approach, over-reliance on third-

350 party waste energy has also contributed to failures. A source of low value heat on-
351 farm may be a motivation for diversification into warm-water fish culture. Another
352 incentive is the utilisation of disused or underutilised agricultural buildings although
353 low cost-purpose built structures such as insulated polytunnels also have potential.
354

355 Reducing the capital requirement and design complexity is an important advantage for
356 any production system. In principle AST are simpler to design and manage than RAS;
357 solids (feed and floc) are kept in suspension and dissolved oxygen levels maintained
358 through aeration. As the culture unit also acts to treat wastes, there is no requirement
359 for external biofilter, piping or pumps which results in lower capital costs and
360 theoretically, more straight forward management. The capital outlay of these
361 components for an RAS can range typically from 10- 35 % of initial fixed costs. Low
362 cost, simple AST could also be temporary or moveable structures allowing farmers to
363 take advantage of seasonal availability of space, resources and marketing
364 opportunities.
365

366 A potential incentive for producing tilapia using a microbial floc-based system rather
367 than conventional RAS is the possibility that local feeds can be used. The overall
368 reduction in feedstock quality required to raise tilapia in AST is potentially a
369 substantial saving on production costs over RAS in which feed cost typically make up
370 from 30-40% of total operating costs depending on the scale of the operation and
371 other factors (Timmons et al., 2002, Timmons, 2005). Using a feed of lower overall
372 quality feed i.e. 20% crude protein feed rather than typical formulations (28-32%CP)
373 could reduce reliance on feed ingredients such as fish meal and soybean meals.
374 Potentially it could open opportunities for growing or using feed ingredients locally or
375 on-farm in a similar manner to that practiced for intensive dairy production thus
376 reducing risk and enhancing familiarity that were important priorities for potential
377 adopters (Rosa et al., 2007).
378

379 Over-ambitious production schedules, steep technical learning curves and lack of
380 prior aquaculture experience have been inter-related causes of recurrent failure in
381 RAS. Contract farming packages which emphasise potential gains while under-
382 estimating risk has contributed to spectacular failures in other novel farm
383 diversification start-ups (e.g. ostrich, and Alpaca farming). Research indicates a
384 similar threat in the UK tilapia sector. Small-scale modular approaches hold potential
385 for limiting risks carried by new adopters with no previous aquaculture experience.
386 These adopters then have the option of scaling up to more economically efficient units
387 required to supply higher volume/ low margin commodity chains (food processors and
388 supermarkets), or continuing to produce smaller volumes of fresh product for higher
389 value niche markets. In the US, innovative tilapia production initially targeted value
390 added markets but relatively high labour costs undermined their capacity to compete
391 with imports leading them to target specialist live sales, often to ethnic minorities
392 (Serfling, 2000). Significant scale-up in production of tilapia and other species such as
393 *Pangasius* spp. in tropical countries threatens competitiveness of producers in the
394 commodity sector in the UK.
395

396 A key research question is; can such a production approach be maintained at
397 production levels that would be cost effective and attractive to farmers in the UK?
398 The use of aquatic microbial floc as the basis for tilapia production has been
399 advocated, but research on intensive indoor/ brown-water production systems is still

400 required to justify promotion of the AST approach to farmers in temperate climates
401 such as the UK.

402

403 **Comparing performance of AST and RAS**

404

405 There is a recent history of research on the operation and efficiency of AST systems,
406 most of which is based on intensively fed, green water systems in ponds or tanks
407 (reviewed by Hargreaves, 2006). Most of the commercial application appears to relate
408 to the relatively much lower-density shrimp production with relatively little published
409 information regarding higher density fish production systems (Avnimelech, 2007).
410 Unfortunately there is a dearth of data for replicated large-scale research systems and
411 most conclusions have been drawn based on either short term small-scale experiments
412 and/or observation of commercial or semi-commercial systems based on variable
413 sized fish (Table 2). Only two trials (Rakocy, Bailey, Thoman & Shultz, 2002,
414 Murray et al., 2007) report on-growing to the minimum harvest size of 400g feasible
415 for markets in the UK. Most reports have emphasised the potential for improved
416 feeding efficiency based on nutrient recycling in AST systems compared to RAS or
417 conventional pellet-fed ponds (Avnimelech et al, 1989; Milstein, Avnimelech, Zoran
418 & Joseph., 2001). Avnimelech (2006) for example cites feed: cost ratios in C:N
419 manipulated pond-AST as being almost double control systems with higher crude
420 dietary protein inclusion. However, meaningful evaluation of the commercial
421 potential of AST compared to RAS also requires knowledge of fish growth rates and
422 system carrying capacities. Unfortunately key parameters that would allow
423 interpretation of growth are often lacking (e.g. water temperature) or inadequately
424 presented. In particular crude daily weight gain rather than specific growth rates are
425 routinely used for comparisons using fish of highly variable stocking and harvest
426 weights.

427

428 Even allowing for these limitations the magnitude of difference between growth rates
429 is evident. Under controlled temperatures, stocking and feed conditions with C:N
430 manipulation and solids removal, Murray *et al* (2007) found growth rates in AST
431 were only 68% of those achievable in RAS (achieving an SGR of 2.8 % for fish
432 grown from 19g-405g; both systems fed on 30% CP diets). SGRs fell to 36% of RAS
433 levels in AST fed on 18% CP diets. When one accounts for the slower growth rate of
434 larger-fish, grow-out time to 400g is almost doubled in the fastest growing AST
435 compared to the RAS control (Murray et al., 2007).

436

437 Low carrying capacities make the commercial case for intensive AST appear still
438 more marginal. Stocking densities exceeding 100 kg fish m⁻³ are routinely achievable
439 in RAS with oxygenation and densities up to 70-80 kg fish m⁻³ with aeration
440 (Timmons et al., 2002). This compares to reported levels of only 10 - 16.5 kg fish m⁻³
441 in AST (Table 2). Murray et al. (2007) achieved levels of 28 kg fish m⁻³, but only
442 using complete feeds and solids removal. Clearly the benefits of feed and water use
443 efficiencies reported for AST need to be viewed in the context of growth inhibition
444 and reduced carrying capacity in intensive systems. Both factors have consequences
445 for overall production costs when capital and variable costs for building size, floor
446 area, insulation, labour and heating etc. are considered. The same constraints also
447 eliminate gains in water efficiency; Murray et al. (2007) and Rakocy et al. (2002)
448 measured broadly comparable optimal rates of 7.2 and 9.7 kg m⁻³ achievable in AST
449 compared to typical RAS rates of 8-10 kg m⁻³ (Tables 1 and 2).

450
451 The potential for further intensification appears to be fundamentally limited by
452 biological factors which correlate with bio-floc concentration in closed systems.
453 Hargreaves (2007) observed process instability at feeding levels above 200 g m^{-3}
454 equivalent to a stocking density of 10 kg m^{-3} (at a feed rate of $2\% \text{ bw day}^{-1}$). Rakocy
455 et al. (2002) and Murray et al. (2007) observed severe growth inhibition and increased
456 mortality at TSS levels above 850 mg l^{-1} . In practice therefore, there is a requirement
457 for solids removal in AST to maintain a level of suspended solids which will not
458 significantly retard food intake and growth or constrain economically viable stocking
459 densities. This requires some form of external clarifier (Murray et al., 2007;
460 Hargreaves, 2006; Rakocy et al., 2002). However, the variable quality (size,
461 consistency and specific gravity) of microbial floc that occurs over time complicates
462 the design and operational management of such clarifiers in indoor AST systems
463 (Murray et al., 2007).

464
465 Operation of AST incorporating solids removal also represents a partial step-back
466 towards RAS-type compartmentalisation with semi-continuous or continuous water
467 re-circulation. Solids settled in external clarifiers could be removed or managed
468 entirely independently for controlled release to the grow-out compartment.
469 Investigators found floc composition varied in closed culture-systems with
470 implications for chronic and acute event-mortalities (Murray et al., 2007; Azim, Little
471 & North., 2007; Rakocy et al, 2002). Compartmentalisation could also provide a
472 means of floc-stabilisation potentially incorporating activated-sludge techniques
473 borrowed from the water and sanitation sector where steady-state operation is a
474 critical feature. One commercial producer in the United States has already moved
475 along this route in a hybrid system; maintaining TSS within $70\text{-}130 \text{ mg l}^{-1}$ and
476 achieving net yields of $60 \text{ kg m}^{-3} \text{ year}^{-1}$ (Serfling, 2000); in other words, resorting to
477 use of bio-floc primarily as a low-cost *in-situ* water treatment process with low water
478 exchange requirements.

479
480 The fundamental theoretical benefit of AST; improved feed efficiency can also be
481 challenged. Analysis of feed and crude protein conversion and retention indicate that
482 the amounts of microbial floc in a brown water system utilised as feed over a range of
483 commercial stocking densities in fish offered feeds of a range of quality and
484 presentational form were minimal (Murray et al., 2007). This contrasts markedly with
485 the values published by Avnimelech (1999) based on observations of light-driven
486 AST systems but could reflect differences in interpretation of data. Attempts to
487 manage microbial floc production by manipulating C:N ratio, or floc levels through
488 solid removal are also highly variable in the systems described. Azim *et al* (2007) also
489 reported increased feed conversion efficiency in AST compared to RAS systems in
490 which fish were maintained at low densities and fed similar amounts of feed
491 confirming the utilisation of microbial floc by fish.

492
493 There are other important characteristics of tilapia culture in AST that deserve
494 mention. The specific conditions of AST appear to favour beneficial bacteria and
495 reduce disease incidence compared to alternative systems. The absence of disease in
496 AST systems has been related to the probiotic nature of microbial floc (Serfling,
497 2000; Murray et al 2007; Avnimelech and Bejarano, 2007).
498

499 The natural habitat of tilapias are turbid water lakes of Africa but the high levels of
500 suspended solids that characterise AST has raised issues regarding the welfare and
501 taste of the fish produced in such systems. The impacts of high levels of microbial
502 floc in AST systems on the taste and welfare of tilapias has been recently assessed.
503 Off-flavours are related to the absorption and accumulation of natural chemicals or
504 compounds (such as geosmin and MBT) through the gills, skin or gastrointestinal
505 tract of fish (Boyd & Tucker. 1998; Gautier et al., 2002). Contrary to common
506 perception culture systems rich in natural food are not necessarily more likely to
507 produce fish with off-flavour (Serfling, 2000; Eves, Turner, Yakupitiyage, Tongdee,
508 & Ponza, 1995). Bue (2005) conducted organoleptic taste trials on fish raised in both
509 RAS and AST and found no perceived differences among fish direct from tanks or
510 after standard depuration techniques in fresh or saline water
511 (Rungreungwudhikrai,1995).

512
513 Generally high levels of suspended solids are related to poor fish welfare as indicated
514 by poor growth, fusion of gill lamellae (Mettam, 2005) and susceptibility to bacterial
515 or parasite infections (Noble & Summerfelt, 1996). Lower feed intakes and
516 performance withstanding, Vincent (2006) found no indication of gill damage on fish
517 raised over extended periods within AST or RAS systems nor differences in tail
518 erosion, scale loss etc characteristic of poor welfare.

519
520

521 **Future research needs**

522

523 Assessment of the development process towards an intensive system for UK farmers to
524 produce and market an exotic food fish species has identified a number of interesting
525 issues. Prototype RAS systems now require testing with potential producers and this
526 will require an iterative action learning approach whereby insights of the adopters are
527 incorporated. Studies on the nature of entrepreneurship give some insight as to the
528 characteristics of potential adopters and whether diversification was driven by need or
529 opportunism (Rosa et al., 2007).

530

531 Clearly there are trade-offs in terms of environmental and broader ethical values of
532 fish produced by RAS and AST. Both systems have very limited effluents which
533 through virtue of their nutrient concentration are useful fertilisers (Watten, & Busch,
534 1984; McMurtry et al., 1997). Further research to quantify the potential synergisms
535 between water and nutrient use in tank based systems and associated high value
536 horticulture is required. Integration with hydroponics has particular market potential
537 as demand for closed cycle, pesticide-free fruit and vegetables increases.

538

539 Although fish produced in AST systems had few overt signs of poor welfare, the
540 lower feed intake, slower individual growth and chronic mortalities observed suggest
541 that RAS provided more consistent and optimal conditions. Further development of
542 mixed systems has been advocated in which culture units are partitioned with algae,
543 microbial floc and/or periphyton (e.g. Avnimelech, 2006; 2007; Azim and Little,
544 2006; Serfling, 2000). Optimisation of floc levels for commercial applications is a
545 research priority.

546

547 The value of microbial floc in terms of preventing fish disease problems warrant
548 scientific investigation. Probiotic approaches are now widespread in the market but

549 the relative control possible in AST and observations of the high health of fish
550 produced makes further investigation worthwhile. Designs in which the natural feed
551 component can be optimised with respect to nutritional quality and energy efficient
552 ingestion, digestion and assimilation should be prioritised. Development with
553 producers in an action research mode is most likely to result in models which are
554 management efficient and adoptable.

555
556 Intensive tilapia production is land efficient (Table 1) and may be located in
557 periurban, rather than rural locations. Benefits include the improved access to a range
558 of consumers, potentially reducing marketing costs. Controlled environments leading
559 to improved predictability of production and expected genetic and feeding gains as
560 has occurred in the broiler industry over the longer term are expected to further
561 improve competitiveness compared with other fish species and substitutes (Timmons,
562 2005)

563
564 The market context for tilapia sales in the UK is dynamic. Consumers are
565 increasingly willing to try new preparations and species of fish (Seafish, 2006b),
566 whether it be for health reasons, indulgence or environmental grounds. Potential for
567 tilapia therefore exists, not only in ethnic markets as a fresh or live alternative to
568 frozen imports, but as a locally available ‘green’ fish product possibly with eco-
569 credentials (Young, et al, 2006). Tilapia also has potential in the food service sector,
570 where novel, exciting fish products are of interest, particularly if they have amenable
571 aesthetic and preparation qualities (Seafish, 2006a). A locally available, small-scale
572 and high quality tilapia supply would therefore meet industry wide interest in fresh,
573 traceable fish supplies, however, a comparative analysis of the relative
574 competitiveness of tilapias produced locally against imports and substitutes is
575 required.

576
577

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587
588

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