

1 **Causes and consequences of sexual dimorphism in mandrill**
2 ***(Mandrillus sphinx)* feeding and spatial behavior.**

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Summary Abstract

17 Sexual dimorphism is a widespread phenomenon, the origins of which have long intrigued
18 evolutionary ecologists. Classic evolutionary theory attributes the evolution of sexual
19 dimorphism to sexual and fecundity selection, that arise due to sex differences in optimal
20 reproductive strategy. However, cross-species comparative analyses have found mixed
21 support for sexual and fecundity selection as drivers of dimorphism. Other hypotheses may
22 therefore be required to explain the full diversity of sexual dimorphisms found in nature.

23 One alternative is the evolution of sexual dimorphism via ecological causation. This thesis
24 explores the possibility of ecological causation of sexual dimorphism, with a particular focus
25 on mandrills (*Mandrillus sphinx*). Mandrills are the most sexually dimorphic primate, with
26 males over three times larger than females, possessing 45 mm canines and displaying
27 extravagant red, blue, and violet ornamentation on their faces and rumps. Males also
28 abandon social groups outside of the breeding season, foraging and travelling separately to
29 females, which could suggest an ecological component to mandrill dimorphism.

30 One of the key predictions of the ecological causation hypothesis is that greater sexual
31 dimorphism should be associated with greater ecological divergence between males and
32 females. In Chapter 2, I conducted a meta-analysis of 163 species to examine associations
33 between sexual dimorphism and ecological sex differences, measured using stable isotopes.
34 Across species, greater size dimorphism was associated with greater sex differences in trophic
35 level. Chapter 2 thus supports a key prediction of the ecological causation hypothesis and
36 suggests that the strength of relationships between sexual dimorphism and ecology may be
37 underappreciated.

38 Mandrills are challenging to study in the wild, due to the dense forests in which they are
39 found. I therefore applied research tools that avoided the need to directly observe individuals,
40 to study the socioecology of males and females and investigate the possibility of ecological
41 causation of mandrill dimorphism. In Chapter 3, I analyze the contents of 4024 mandrill fecal
42 samples, collected over an eight-year period, in conjunction with eight years of fruiting
43 phenology data. I find that mandrills are primarily frugivorous, with consumption of fruit
44 tracking phenological patterns of fruit production. The focal group also demonstrated high

45 dietary flexibility, consuming animal prey, leaves, crushed seeds, and other plant fibers when
46 fruit availability was low.

47 However, because male mandrills spend much of year solitary, the feeding niche information
48 gained from Chapter 3 largely relates to females. To investigate sex differences in feeding
49 niche, I therefore applied stable isotope analysis to mandrill hairs. In Chapter 4, I detail the
50 results of a diet-switch experiment, that quantified mandrill hair growth rates, isotopic tissue
51 turnover times and isotopic tissue-diet discrimination factors. This experiment showed that
52 stable isotope analysis of mandrill hairs can reveal dietary information on wild mandrills at a
53 temporal resolution of 5.5 days and allowed the stable isotope ratios of wild mandrills to be
54 correctly adjusted for accurate dietary inference.

55 In Chapter 5 I applied stable isotope analysis to the hairs of wild male and female mandrills. I
56 found that the diets of individual male mandrills were more consistent than individual
57 females, who instead showed more dietary variation, at the individual level. This result
58 suggests that individual male mandrills had more consistent access to preferred resources
59 than individual females, and therefore that males experience differing levels, or outcomes, of
60 resource competition. Males may therefore seasonally leave mandrill social groups to avoid
61 resource competition.

62 In Chapter 6, I used GPS collars to examine sex differences in mandrill spatial behavior. I found
63 that, during the breeding season, male habitat selection, home-range size and Brownian
64 motion variance was similar to females, but distinct from females during the non-breeding
65 season. These results indicate that collared males were not always within the social group
66 during the non-breeding season. Furthermore, during the non-breeding season, males
67 travelled shorter distances, at slower speeds, compared to the breeding season. Males
68 therefore appeared to expend less energy on travel during the non-breeding season,
69 suggesting that male mandrills may also leave groups to reduce energetic expenditure.

70 The results of Chapters 5 and 6 suggest that ecological factors compel male mandrills to
71 abandon social groups outside of the breeding season. However, the intensity of sexual
72 selection on male traits may be increased as a result of this behavior, as males must establish
73 dominance and attract unfamiliar females when they re-join social groups to breed. An

74 interaction between sexual and ecologically mediated selection may therefore offer the most
75 complete explanation of why mandrill dimorphism is so extreme.

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Declaration of authorship

78 I, Joshua T Bauld, declare that this thesis has been composed by myself and that it embodies
79 the results of my own research. Where appropriate, I have acknowledged the nature and
80 extent of the work carried out in collaboration with others.

81 **Signed**.....

82 **Date**.....

83

84

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Chapter 1: General Introduction

1.1 Background

Sexual dimorphism describes differences between males and females in traits such as size, shape, coloration, scent, and behavior (Andersson, 1994; Mori *et al.*, 2017). Sexual dimorphism is pervasive across species and incredibly diverse, for example, males are the larger sex in most mammals (Weckerly, 1998), whereas female anglerfish (*Ceratias holboelli*) are around 500 000 times heavier than males (Bertelsen, 1951). Some dimorphic traits seem clearly related to each sex's reproductive role, such as the elaborate plumage of male birds-of-paradise (Paradisaeidae), whereas others appear connected to each sex's ecology, as in the beaks of Caribbean hummingbirds (*Eulampis jugularis*) (Temeles *et al.*, 2000). Explaining this diversity has been a longstanding goal of evolutionary ecology but debate as to the origins of sexual dimorphism still persists (Shine, 1989; Wyman *et al.*, 2013).

This thesis investigates the possible ecological causes and consequences of sexual dimorphism, with a particular focus on mandrills (*Mandrillus sphinx*). Mandrills display extraordinary sexual dimorphism in traits such as body size and ornamentation (Darwin, 1871; Abernethy and White, 2013; Setchell, 2016), as well as an unusual social structure and mating system, involving groups of several hundred animals (Abernethy *et al.*, 2002; Hongo *et al.*, 2016). Why the dimorphism and social behavior of mandrills has evolved to become so extreme is, however, an open question. Mandrills are also vulnerable to extinction, primarily due to habitat loss and climate change, and improved knowledge of the species' understudied ecology is essential to improving its conservation prospects (Abernethy and Maisels, 2019).

Mandrills have historically proven difficult to study in the wild, due to the dense nature of their forest habitat and the challenge of habituating groups of several hundred animals. As a result, in this thesis I use fecal and stable isotope analysis, alongside GPS collars, to study the feeding and spatial behavior of both sexes. These techniques allow ecological data to be gathered from unhabituated animals, by sidestepping the need to directly observe their behavior. Using these data, I consider the impacts of group size on the ecology of each sex, and whether ecological factors like feeding competition and energetic costs may have contributed to the evolution of sexual dimorphism in mandrills.

30 **1.2 The evolution of sexual dimorphism.**

31 Much of the theory explaining the evolution of sexual dimorphism stems from the Darwin-
32 Bateman Paradigm (Dewsbury, 2005). Darwin (1871) observed that one sex, most often
33 males, competes for mating opportunities with the more discriminating sex in terms of mating
34 partner, most often females. The mechanism behind this sex difference in mating strategy
35 was proposed by Bateman (1948), who suggested that male *Drosophila* flies were able to
36 enhance their reproductive success through additional matings, whereas females were not.

37 Sex differences in optimal mating strategy are rooted in anisogamy, defined as sexual
38 reproduction involving the fusion of gametes that differ greatly in size. In the context of
39 anisogamy, males are defined as the sex with smaller gametes and females as the sex with
40 larger gametes (Lehtonen *et al.*, 2012). It is these sex differences in gametic investment that
41 initially impose differing constraints on the reproductive success of each sex: males invest less
42 energy to produce many small sperm, and are primarily limited by the number of eggs those
43 sperm are able to fertilize, and females invest more energy into producing fewer large eggs,
44 and are usually limited by investment into gamete production (Wade, 1979). Males may
45 therefore compete directly amongst each other for access to females (and their gametes) or
46 compete indirectly to attract receptive females (and their gametes) (Andersson and Iwasa,
47 1996; McCullough *et al.*, 2016). The selection that arises from this competition for access to
48 mates and their gametes is termed sexual selection (Shuker and Kvarnemo, 2021). To
49 compete for matings, males may evolve larger body size or exaggerated weaponry (Lindenfors
50 *et al.*, 2002; Bro-Jørgensen, 2007). In addition, males may evolve ornaments or courtship
51 displays that signal their quality to females (Emlen *et al.*, 2012; Mitoyen *et al.*, 2019). Sexual
52 selection is often considered to be the only robust theory for explaining the evolution of
53 exaggerated ornaments and weaponry (Parker and Pizzari, 2015).

54 In contrast to sexual selection, fecundity selection considers traits that affect gamete
55 production (Pincheira-Donoso and Hunt, 2017). Darwin (1871) originally formulated the
56 'fecundity advantage hypothesis' to explain female biased size dimorphism, proposing that
57 increased size may facilitate the production of greater numbers of gametes by females.
58 Fecundity selection has received a great deal of empirical support, with larger body size often
59 being associated with increased egg production in various invertebrates and ectothermic

60 vertebrates (Head, 1995; Prenter *et al.*, 1999; Cox *et al.*, 2003; Dugo-Cota *et al.*, 2022).
61 However, in mammals, larger female body size is associated with slower life histories and
62 greater energetic costs of reproduction (Lindenfors *et al.*, 2007). Thus, it has been suggested
63 that larger size may incur fecundity costs in female mammals and that fecundity selection
64 could promote sexual size dimorphism in mammals by constraining female body size (Cassini,
65 2017).

66 The general framework of sexual selection acting on males, and fecundity selection on
67 females, has proven useful in explaining the origins of sexual dimorphism, but other factors
68 are also relevant (Fairbairn *et al.*, 2007; Dugo-Cota *et al.*, 2022). Sexual selection may act on
69 females (Clutton-Brock 2009) and fecundity selection on males (Pincheira-Donoso and Hunt,
70 2017). For example, female dance flies (*Rhamphomyia longicauda*) possess sexually selected
71 ornaments designed to attract male mating partners (Murray *et al.* 2018). Moreover, in
72 polygamous *Syngnathus* pipefish, which exhibit male pregnancy, male morphology appears
73 to be influenced by fecundity selection (Wilson, 2009). Comparative analyses have also found
74 sexual and fecundity selection to be modestly and inconsistently associated with sexual size
75 dimorphism, suggesting that their effect on dimorphism may vary substantially taxa (Cox *et*
76 *al.*, 2007; Janicke and Fromonteil, 2021). Consideration of a wider range of evolutionary
77 mechanisms, including those mediated by ecology rather than reproduction, may thus be
78 required for a full understanding of how sexual dimorphism evolves (Krüger *et al.*, 2014;
79 Littleford-Colquhoun *et al.*, 2019).

80 In cases of sexual dimorphism in traits relevant to ecology, such as body size or feeding
81 morphology, selection mediated by ecology is especially likely to be relevant to species'
82 evolution (Selander, 1966; Hendrick and Temeles, 1989; De Lisle and Rowe, 2015). Such
83 ecological sexual dimorphisms were noted by Darwin (1871) as being “connected with
84 different habits of life, and not at all, or only indirectly, related to reproductive functions” (p.
85 254). Three main processes have been put forward for ecologically mediated natural selection
86 to produce sexual dimorphism. The first is the dimorphic niches hypothesis, which predicts
87 that intrinsic differences between the sexes in nutritional or energetic requirements may
88 favor the evolution of different optimal phenotype in males and females (Lande, 1980; Slatkin,
89 1984). The dimorphic niches hypothesis is supported by evidence that the fitness of males
90 and females is optimized by divergent nutritional intakes, suggesting that the sexes can differ

91 in intrinsic requirements (Hunt *et al.*, 2004; Maklakov *et al.*, 2008). The second is the bimodal
92 niches hypothesis, which suggests that the sexes may have similar ecological needs, but that
93 two distinct phenotypes may exist with the ability to meet those needs, with each sex
94 potentially evolving to either phenotype (Schoener, 1969; Slatkin, 1984). There appears to be
95 less support for the bimodal niches hypothesis in the literature (but see Schoener, (1977))
96 and evidence of differing nutritional optima between the sexes also discredits a hypothesis
97 that requires the sexes to have similar ecological needs.

98 The third proposal, and perhaps the most studied, is the ecological character displacement
99 hypothesis (Slatkin, 1984; Li and Kokko, 2021). Under this hypothesis, disruptive selection
100 resulting from frequency-dependent competition for resources is predicted to favor the
101 evolution of ecological sexual dimorphism that lowers competition for the resource (Cooper
102 *et al.*, 2011). Disruptive selection may arise when a quantitative trait affects use of a
103 continuously varying resource. For example, birds may feed on a continuum of seed sizes, the
104 consumption of which is influenced by bill size. Initially, in a normally distributed population,
105 the fitness of individuals with intermediate sized bills may be highest, if medium seeds are
106 the most common resource. However, as individuals with intermediate sized bills become
107 more frequent in the population, the availability of medium seeds will fall. The relatively high
108 availability of small and large seeds may then favor individuals with small or large bills, that
109 specialize on these resources, as these individuals will experience lower feeding competition
110 and have the highest food intake. Disruptive selection arises once the fitness of individuals
111 with medium bills falls below the fitness of individuals with small or large bills. Selection will
112 then favor the evolution of two distinct phenotypes and, if each sex evolves towards a
113 particular bill size, an ecological sexual dimorphism (Martin and Pfennig, 2009).

114 Modelling work has successfully demonstrated that ecological character displacement
115 between sexes, mediated by resource competition, can produce sexual dimorphism (Slatkin,
116 1984; Bolnick and Doebeli, 2003; Li and Kokko, 2021). De Lisle and Rowe (2015) have also
117 demonstrated the process experimentally, by showing that the fitness of individual
118 salamanders, that display sexual dimorphism in feeding morphology and resource use, is
119 highest when competing for resources against individuals of the opposite sex.

120 A limitation of models for ecological character displacement between sexes is that they do
121 not include a justification for why an ecological sexual dimorphism should evolve, instead of
122 ecological speciation or two ecologically distinct morphs containing both sexes (Bolnick and
123 Doebeli, 2003). A combination of evolutionary processes may therefore provide the most
124 likely scenario for explaining the evolution of ecological sexual dimorphism (De Lisle, 2019).
125 For example, sexual or fecundity selection may produce an initial dimorphism, say in body
126 size, and if disruptive selection on size arises, each sex will already be biased to evolve towards
127 each end of the phenotypic spectrum. Similarly, if sex differences in nutritional requirements
128 result in dimorphic niches, and frequency-dependent competition for resources arises, males
129 and females may again be expected to respond in divergent manners, if their sex-specific
130 nutritional optima impact foraging decisions. For example, De Lisle (2023) found that, after
131 only three generations, *Drosophila melanogaster* evolved greater sexual dimorphism in
132 dietary preference under high resource competition, but not under low resource competition.
133 After three generations, females and males from the high competition treatment fed on
134 greater proportions of yeast solution and sucrose solution, respectively. This result suggests
135 that inherent sex differences in nutritional requirements, related to the reproductive roles of
136 each sex, led to divergent responses by males and females to resource competition (De Lisle,
137 2023).

138 Ecological causation has often been regarded as a less parsimonious explanation for the
139 evolution of sexual dimorphism, and of minor importance, compared to sexual or fecundity
140 selection (Shine, 1989; Blanckenhorn, 2005). However, the abundance of examples of
141 ecological dimorphism found in the literature (Temeles, *et al.* 2000; Pearson, *et al.* 2002; Box
142 2, De Lisle, 2019; Bauld *et al.*, 2022, this thesis), alongside the modelling and experimental
143 work outlined above, make a compelling case for the evolution, or exaggeration, of sexual
144 dimorphism by ecological processes. The primary goal of this thesis is thus to investigate the
145 possibility that the sexual dimorphism of mandrills could have been enhanced by ecologically
146 mediated selection, not directly related to the reproductive roles of males and females.

147 **1.3 Study Species: Mandrills**

148 *1.3.1 Phylogeny and Biogeography*

149 Mandrills are among the largest cercopithecine primates and located in West Central Africa
150 (Abernethy and White, 2013). Their range extends from southern Gabon, through Equatorial
151 Guinea and southwestern Republic of Congo, to the Sanaga river in Cameroon (Abernethy
152 and Maisel, 2019); to the north of which their congener drills (*Mandrillus leucopheus*) are
153 present (Gadsby *et al.*, 2019). The eastern distribution of mandrills is largely limited by the
154 Ivindo and Ogooué rivers, though mandrills have twice been observed in northern Democratic
155 Republic of Congo, but never east of the Congo river (Fay, 1989; Allam *et al.*, 2016; Abernethy
156 and Maisel, 2019). The Ogooué river divides mandrills into two genetically distinct northern
157 and southern populations, occupying northern Gabon, Cameroon and Equatorial Guinea
158 versus southern Gabon and Democratic Republic of Congo, respectively (Telfer *et al.*, 2003).

159 The genus *Mandrillus* is placed within the African Papionines (subtribe: Papionina), that also
160 contains baboons (*Papio* spp), terrestrial mangabeys (*Cercocebus* spp), arboreal mangabeys
161 (*Lophocebus* spp), geladas (*Theropithecus* spp) and kipunjis (*Rungwecebus* spp). Body size
162 differences between genera initially led to taxonomic confusion, with the small-bodied
163 mangabeys and large bodied mandrills, drills, baboons, and geladas considered as two distinct
164 groups (Pugh and Gilbert, 2018). The result was that mandrills were initially characterized as
165 “forest baboons” (Hoshino, 1984; 1985; Feistner, 1989). However, molecular work indicated
166 that mandrills, drills, and terrestrial mangabeys were a distinct phylogenetic group, more
167 distantly related from baboons and arboreal mangabeys (Harris, 2000). A *Mandrillus-*
168 *Cercocebus* clade was further supported by morphological analyses, showing that both genera
169 possessed dental and forelimb morphology adapted to consumption of hard foods and
170 foraging in leaf litter (McGraw and Fleagle, 2006).

171

172 *1.3.2 Socioecology*

173 Knowledge of wild mandrill behavior has historically proven difficult to obtain, due to the
174 challenges of studying unhabituated primates in the dense forests of Central Africa
175 (Brockmeyer *et al.*, 2015; Hongo *et al.*, 2016). Following the early characterization of mandrills

176 as baboons, one male units, containing a single adult male and a few females, were initially
177 thought to be the basic component of mandrill societies (Jouventin, 1975; Hoshino *et al.*,
178 1984). However, subsequent investigations identified a matrilineal structure to mandrill social
179 groups, alongside female philopatry and male dispersal (Setchell, 2000; Abernethy *et al.*,
180 2002; Charpentier *et al.*, 2007; 2012; 2020). In addition, numerous studies have demonstrated
181 marked seasonal variation in the proportion of adult males present in mandrill groups, as well
182 as the existence of solitary males throughout the mandrill range, making a social structure
183 based around adult males unlikely (Abernethy *et al.*, 2002; White *et al.*, 2010, Brockmeyer *et*
184 *al.*, 2015; Hongo *et al.*, 2016). The majority of males only enter mandrill groups when the
185 number of fertile females is high, which occurs during the dry season, allowing females to
186 time most birthing, lactating and infant carrying with the subsequent wet season and high
187 food availability (Abernethy *et al.*, 2002; Hongo *et al.*, 2016; Dezeure *et al.*, 2022). Hongo *et*
188 *al.* (2014) also found that groups traveled with sub-adult and adult males close to the front,
189 with females and dependent infants closer to the rear and unaccompanied by males, further
190 discrediting the one male unit hypothesis. Mandrill social groups are therefore considered to
191 follow a multi-male multi-female structure (Abernethy *et al.*, 2002; Brockmeyer *et al.*, 2015).

192 Early studies of mandrill also underestimated the size of mandrill groups, termed 'hordes'.
193 Using observational methods, groups were estimated to be between 15-95 (Hoshino *et al.*,
194 1984) and 150 individuals (Tutin *et al.*, 1997). These low estimates were likely due to the
195 challenge of observing an entire mandrill group within forest habitats and because hordes
196 often display fission-fusion behavior, resulting in variable sub-group sizes (White *et al.*, 2010).
197 However, using films of forest gap crossings, Abernethy *et al.* (2002) counted a mean horde
198 size of 620 (range = 338- 845) individuals. More recent genetic work suggested an effective
199 population size of 1000 individuals for the same horde (Guibinga Mickala *et al.*, 2020), backed
200 up by a recent field count of 1250 individuals in the horde (David Lehmann, personal
201 communication). Mandrill hordes are therefore among the largest social groups seen among
202 non-human primates (White *et al.*, 2010).

203 Mandrills are found almost exclusively within forest habitats, though they may occasionally
204 cross savannas (White, 2007; Brockmeyer *et al.*, 2015). The sizes of mandrill home-ranges
205 were also underestimated by early observational approaches, at between 30-50 Km²
206 (Jouventin, 1975; Hoshino *et al.*, 1984; Harrison, 1988). However, using radio-telemetry,

207 White *et al.* (2010) estimated a home-range of 182 Km², containing 89 Km² of forest. Mandrills
208 thus exploit some of the largest home-ranges of any non-human primate, which is expected
209 given the extreme biomass of a mandrill horde (Clutton-Brock, 1977; White *et al.*, 2010).

210 The distances traveled by mandrill hordes are usually higher each day during periods of fruit
211 abundance than during periods of scarcity, suggesting fruit availability drives mandrill
212 movement (Hongo *et al.*, 2022). Furthermore, space use is often uneven, as would be
213 expected for a primate mainly consuming fruit resources that are clumped in trees (White, *et*
214 *al.* 2010). Space use does, however, become more even when fruit availability is low,
215 suggesting movement is influenced by resources that are distributed more ubiquitously in the
216 environment during these periods (Hongo *et al.*, 2018).

217 Seasonal changes in mandrill space use correspond with observations of mandrill food
218 selection. Fruit is the principal food source for mandrills, though invertebrate and vertebrate
219 prey, leaves, seeds, and other plant fibers are also eaten (Jouventin, 1975; Hoshino, 1985;
220 Lahm, 1986; Harrison, 1988; Rogers *et al.*, 1996; Nsi Akoué *et al.*, 2017). Fruit consumption is
221 highest during the wet seasons when it is most available, with consumption of alternative
222 food types increasing in the dry seasons, when fruit consumption falls (Tutin *et al.*, 1997;
223 Hongo *et al.*, 2018).

224

225 1.3.3 Mandrill Sexual Dimorphism

226 Mandrills are the most sexual dimorphic species among primates (Darwin, 1871; Dixson *et al.*,
227 2005). Male mass is approximately three times that of females, with males weighing 28.9 Kg
228 (range = 24.5 - 36.0 Kg) and females weighing 9.8 Kg (range = 8.0 - 13.5) (David Lehmann,
229 *unpublished data*). Much of what we know about mandrill sexual dimorphism and behavior
230 comes from a semi-free ranging colony of a few hundred animals, held in a multi-hectare,
231 naturally forested enclosure in Franceville, Gabon (Feistner *et al.*, 1992). Data on these
232 animals has proven highly informative for suggesting the possible functions and drivers of
233 mandrill dimorphism. However, some caution may be needed when translating these results
234 to wild groups, until further research is completed. For example, the behavioral function of

235 mandrill coloration may well be conserved in captivity, but the ability of dominant males to
236 dominate paternity might be exaggerated, compared to the wild.

237 In semi-free ranging groups, males with larger testes and higher plasma testosterone levels
238 are larger in mass and heavier males attain higher dominance ranks than smaller males
239 (Wickings and Dixson, 1992; Setchell *et al.*, 2008). Males also possess red sexual skin on their
240 faces and genitals (Setchell, 2005). The brightness of red coloration in males is again
241 associated with larger testes, higher testosterone levels and rank attainment (Wickings and
242 Dixson, 1992; Setchell and Dixson, 2001; Setchell *et al.*, 2006). The association between
243 brightness and testosterone suggests that red coloration may be an honest indicator of
244 androgen levels and fight readiness, thus acting as a badge of status (Setchell *et al.*, 2008).
245 Indeed, males of similar brightness are less aggressive between each other, than to duller
246 males (Setchell and Wickings, 2005). Females also display a preference for more brightly
247 colored males, meaning that red skin may also act as a signal of male condition to females
248 (Setchell, 2005).

249 The red skin on the faces of males is also accompanied by blue paranasal swellings, formed
250 from an enlarged maxillary ridge. The size of these paranasal swellings increases allometrically
251 with male body size, suggesting that they constitute an honest signal of size and fitness (Klopp
252 *et al.*, 2012). Furthermore, brighter blue swellings contrast more strongly with red sexual skin
253 and increase the contrast of male facial displays against foliage (the typical background in
254 mandrill habitat). The most dominant males are those with the highest color contrasts
255 between red skin, blue paranasal swellings, and foliage, again consistent with the hypothesis
256 that facial displays are a badge of status (Renoult *et al.*, 2011).

257 Mandrills are also sexually dimorphic in terms of canine length, with males averaging 45 mm
258 and females 10 mm canines. Males with canines shorter than 30 mm rarely breed and sires
259 have longer canines than non-sires, implying that longer canines may be a sexually selected
260 weapon in mandrills (Leigh *et al.*, 2008). Both sexes possess sternal glands for scent marking,
261 but the chemical profiles of these scents are distinguishable between the sexes and relate to
262 age and dominance rank in males only (Setchell and Dixson, 2001; Setchell *et al.*, 2010). In
263 addition, scent marking behavior is most prevalent in dominant adult males, suggesting that

264 scent marking is another way in which males communicate their dominance status (Setchell
265 *et al.*, 2010; Vaglio *et al.*, 2015).

266 Like many other cercopithecines, female mandrills display sexual swelling around estrous and
267 the shape of these swellings may contain information about female quality (Setchell and
268 Wickings, 2004; Huchard *et al.*, 2009). During the breeding season, males are observed to
269 mate-guard females with sexual swellings (Dixson *et al.*, 1993). Mate-guarding by males is
270 biased towards females with higher major histocompatibility complex (MHC) diversity and
271 MHC variants of higher quality (Setchell *et al.*, 2016). Males also preferentially mate-guard
272 high ranking or parous females over low ranking or nulliparous females (Setchell and
273 Wickings, 2006). Dominant males are the most successful in mate-guarding, resulting in a
274 positive association between mate-guarding and paternity, alongside assortative mating, in
275 which the highest ranking males mate more often with the highest quality females (Setchell
276 *et al.*, 2005; Setchell and Wickings, 2006).

277 The result of mate-guarding by dominant males, in semi-free ranging groups, is high
278 reproductive skew among males, that is not replicated among females (Wickings *et al.*, 1993;
279 Wickings, 1995). In semi-free ranging groups, reproductive skew falls as the number of
280 tumescent females increases (Charpentier *et al.*, 2005), thus reproductive skew may be lower
281 in wild groups, where dominant males probably maintain less control. However, mature
282 males also mate-guard females with sexual swellings in the wild (Hongo *et al.*, 2016) and so
283 some reproductive skew is likely to persist in wild mandrill groups.

284 The evidence described here suggests a sequence of effects, whereby the development of
285 male secondary sexual traits (larger mass, red sexual skin, blue paranasal swellings, powerful
286 scent glands, long canines) influences rank attainment and female preference, which in turn
287 facilitate successful mate guarding by large, dominant, and brightly ornamented males, who
288 then obtain a higher share of paternity. It is also notable that the sexual dimorphism of
289 mandrills develops due to bimaturation, with males investing at least three or four more years
290 than females towards growth and maturation (Setchell *et al.*, 2001; Dirks *et al.*, 2020).
291 Females also reproduce for the first time approximately four years earlier than males
292 (Wickings and Dixson, 1992; Setchell *et al.*, 2002). Higher reproductive skew and longer
293 maturation periods among males, alongside lower age at first reproduction and longer

294 reproductive lifespans for females, are all strong evidence that sexual selection has had a
295 substantial role in the evolution of mandrill dimorphism (Setchell *et al.*, 2005; Setchell, 2016).

296 Sexual selection is expected to arise when spatial aggregation of fertile females in space
297 allows individual males to monopolize paternity (Emlen and Oring, 1977; Cassini, 2020a).
298 Female primates are expected to form groups, and thus spatially aggregate, as an anti-
299 predation strategy (Sterck *et al.*, 1997). The frugivorous diet of mandrills, coupled with
300 seasonality in fruit availability throughout their geographic range (Hongo *et al.*, 2018; Bush *et*
301 *al.*, 2020), also means that fertile females are temporally clustered.

302 The seasonal timing of cyclical biological events, such as fruit production, is termed phenology
303 (Lieth, 1974). In tropical forests, the phenological cycles of fruiting trees are controlled by
304 numerous variables, including temperature, rainfall, and solar radiation (Borchert, 1983;
305 Reich and Borchert, 1984; Butt *et al.*, 2015; Potts *et al.*, 2020). Across African forests, annual
306 phenological cycles of flowering and fruiting are most frequent, leading to seasonality in fruit
307 availability (Chapman *et al.*, 1999; Bush *et al.*, 2017; Adamescu *et al.*, 2018). Seasonal changes
308 in fruit availability influence many aspects of primate behavior, such as social structure
309 (Asensio and Korstjens, 2009; Aguilar-Melo *et al.*, 2019) and reproductive phenology (Koenig
310 *et al.*, 1997; Matthews *et al.*, 2021).

311 The consequence of fruiting phenology for mandrills is that females time birthing and
312 lactation with annual peaks in fruit availability. As a result, receptive fertile females are
313 clustered in time during the longer annual dry season (Hongo *et al.*, 2016; Dezerure *et al.*,
314 2022). The large size of mandrill hordes means that the number of spatially and temporally
315 clustered fertile females is uncommonly high for a primate species (Abernethy *et al.*, 2002).
316 The potential for males to monopolize females is therefore exceptional in mandrills, which
317 should result in stronger sexual selection and has almost certainly contributed to extreme
318 dimorphism in mandrills.

319

320 *1.3.4 A role for ecological causation?*

321 However, the frugivorous diet and large group sizes of mandrills also have consequences that
322 could, theoretically, exaggerate dimorphism through ecologically mediated selection. In

323 tropical forests, tree species are often aggregated in space, with areas surrounding individual
324 trees containing above average densities of conspecifics (Hubbel, 1979; Condit *et al.*, 2000).
325 Tree species may aggregate depending on habitat conditions, if species are best adapted to
326 particular topographies, light conditions, soil types or nutrients (Itoh *et al.*, 1997; John *et al.*,
327 2017; McFadden *et al.*, 2019). Alternatively, tree distributions may be influenced by seed
328 dispersal mechanisms, leading to lower aggregation among animal dispersed species than
329 those dispersed by wind or explosively dehiscent capsules (Seidler and Plotkin, 2006). This
330 spatial aggregation of tree species results in fruit being clumped into patches within forests,
331 the distribution of which influences primate diets, habitat selection and daily path lengths,
332 among other aspects of behavior (Stevenson *et al.*, 2000; Aristizabal *et al.*, 2019).

333 The clumping of fruit resources results in within-group feeding competition, the intensity of
334 which increases with group size (Sterck *et al.*, 1997). The uneven nature of mandrill space use
335 indicates that clumping of fruit influences their movement and large horde sizes mean feeding
336 competition is likely to be high (White *et al.*, 2010; Hongo *et al.*, 2018). Heavier male mandrills
337 might be expected to easily exclude females from feeding patches, as is seen in gorillas
338 (Watts, 1985). However, travel by mandrills within their home-range tends to be along the
339 ground and feeding often takes place in trees (Hoshino *et al.*, 1984; Lahm, 1986). Males could
340 therefore be disadvantaged during feeding competition because climbing should be more
341 costly for heavier animals (Pontzer, 2016). For example, the large size of male gorillas appears
342 to hinder them in reaching arboreal feeding sites (Remis, 1995). Thus, large group size likely
343 produces high feeding competition in mandrills, which may be especially costly for males.

344 Larger groups of frugivorous primates are also expected to travel further each day between
345 patches of clumped fruit resources (Chapman *et al.*, 1995). Indeed, mandrills travel further
346 during periods of fruit abundance (Hongo *et al.*, 2022), which also coincide with lower
347 numbers of tumescent females (Abernethy *et al.*, 2002; Dezeure *et al.*, 2022). Thus, the
348 periods of greatest energetic expenditure on travel for males, if they are within a horde,
349 coincides with the period when reproductive opportunities are most infrequent. The high
350 body mass of males may also make traveling alongside the horde more challenging than for
351 females (Hongo *et al.*, 2016). This challenge could arise because of trade-offs between the
352 energy an organism expends on body maintenance, which is higher for larger animals, against
353 energy that may be spent on locomotion (Boratyński, 2020). Furthermore, during the

354 breeding season males are mate-guarding, fighting and produce grunting displays for
355 approximately 12 hours per day (Abernethy *et al.*, 2002; Hongo *et al.*, 2016). Probably as a
356 result of these behaviors, the breeding season is associated with elevated fecal glucocorticoid
357 metabolites, parasite load, oxidative and nutritional stress in males (Setchell *et al.*, 2010;
358 Beaulieu *et al.*, 2014; Oelze *et al.*, 2020).

359 High feeding competition, in which males may be disadvantaged, coupled with greater costs
360 of travel and mating competition may well explain observations that male mandrills leave
361 hordes outside of the breeding season (Abernethy *et al.*, 2002; Hongo *et al.*, 2016). However,
362 the resulting seasonal influx of males during the breeding season has two consequences for
363 the mandrill mating system. Firstly, the hierarchy among males is unstable, meaning males
364 must fight to establish dominance when they enter a horde to breed (Smit *et al.*, 2022). This
365 increase in intra-sexual mating competition could explain why mandrill size and canine (i.e.
366 weaponry) dimorphism is so extreme. Secondly, males must obtain mating opportunities in
367 the absence of long-term social relationships with females (Abernethy *et al.*, 2002). The need
368 to court unfamiliar females in a dense habitat implies high inter-sexual mating competition,
369 which could be responsible for the intensity of male ornamentation and grunting
370 vocalizations. Thus, the temporary nature of male horde membership could result in stronger
371 sexual selection and extreme dimorphism in mandrills, compared to primate species with
372 more stable group structures. If resource competition and travel costs are responsible for
373 males leaving mandrill hordes, ecologically mediated selection may be said to have acted to
374 exaggerate mandrill dimorphism, by selecting for male social behavior that subsequently
375 increases sexual selection.

376 To demonstrate that resource competition and travel costs compel males to leave social
377 groups, sex differences in feeding and spatial behavior must be quantified. However, as the
378 above discussion of mandrill socioecology demonstrates, obtaining accurate data on wild
379 mandrill behavior via observational methods is extremely challenging. The challenge is
380 especially great for solitary male mandrills, which are rarely observed beyond short sightings,
381 and so their behavior is essentially unknown. One solution to the challenge of observing
382 unhabituated wild animals is to study their ecology using remote tracking devices and stable
383 isotope analysis (Cooke *et al.*, 2004; Crawford *et al.*, 2008; Handcock *et al.*, 2009).

384 Stable isotope analysis has been successfully used to investigate numerous features of
385 primate ecology, including feeding competition (Dammhahn and Kappeler, 2010; Flores-
386 Escobar *et al.*, 2020), nutritional stress (Oelze *et al.*, 2020), habitat segregation (Schoeninger
387 *et al.*, 1997; Carter and Bradbury, 2016) and meat consumption (Oelze *et al.*, 2011). Stable
388 isotope analysis works by quantifying the ratios of heavy to light stable isotopes in animal
389 tissues, which reflect the ratios of the foods they have consumed (Crawford *et al.*, 2008). For
390 example, nitrogen isotope ratios ($^{15}\text{N}/^{14}\text{N}$) reflect the trophic level at which an animal feeds
391 (Hobson and Welch, 1992). Conversely, carbon isotope ratios ($^{13}\text{C}/^{12}\text{C}$) in terrestrial
392 ecosystems reflect the plant species that an animal, or its prey, have eaten (Codron *et al.*,
393 2007).

394 For primates that are difficult to observe, stable isotope analysis of hairs is particularly useful.
395 Because keratin is metabolically inert, the isotope ratios of hairs remain unchanged after
396 growth (McHuron *et al.*, 2016). The linear nature of hair growth in turn means that variation,
397 or a lack thereof, in isotope ratios along the length of a hair record temporal changes in diet
398 (Cherel *et al.*, 2009). A single sample of hair thus contains long-term information about the
399 feeding niche of an individual, overcoming the need for repeated observations of feeding
400 behavior (Oelze, 2016). Analyzing the stable isotope ratios of mandrills hairs will thus allow
401 the feeding niches of males and females to be contrasted and the impact of feeding
402 competition on each sex to be explored.

403 GPS collars allow the movement of primates to be tracked in challenging conditions that
404 prevent direct focal follows (Dore *et al.*, 2020). By repeatedly recording the locations of an
405 individual, aspects of spatial behavior such as home-range size and habitat selection can be
406 studied (Pebsworth *et al.*, 2012; Anderson *et al.*, 2017). Investigating the locations, home-
407 ranges and habitat choices of male and female mandrills will reveal whether males are more
408 likely to be found outside of the horde during the breeding season, as is expected based on
409 observational studies. Furthermore, fitting GPS collars allows the travel distances and speeds
410 of tracked primates to be quantified (Klegarth *et al.*, 2016; Bracken *et al.*, 2022). These
411 measures will uncover whether males expend less energy on travel during the non-breeding
412 season, when they are expected to be solitary.

413

414 **1.4 Thesis Aims**

415 In this thesis I am to: 1) Explore whether ecologically mediated selection may have a more
416 general role in the evolution of sexual dimorphism than is currently appreciated; 2) Expand
417 our knowledge of mandrill socioecology, by examining the feeding and spatial behavior of a
418 wild horde living in a relatively intact habitat, largely free of human disturbance over the past
419 40 years; improved knowledge of mandrill socioecology will also inform ongoing conservation
420 efforts throughout mandrill range; 3) Assess the possibility that ecologically mediated
421 selection is partially responsible for the exaggeration of sexual dimorphism in mandrills.

422 **Chapter 2: Can diet niche partitioning enhance sexual dimorphism?**

423 A key expectation of hypotheses predicting that sexual dimorphism can evolve via ecologically
424 mediated selection is that greater sexual dimorphism should be associated with larger
425 ecological sex differences. In Chapter 2 I therefore conduct a meta-analysis investigating the
426 strength of associations between sexual size dimorphism and isotopic (ecological) differences
427 between males and females. I also consider the impact of species mean size, dietary class,
428 and gape limitation on the strength of associations between dimorphism and isotopic sex
429 differences, as each of these factors may be expected to influence the ecological relevance of
430 size dimorphism. I then discuss whether the meta-analytic results imply a strong or a weak
431 role for ecologically mediated selection in the evolution of sexual dimorphism.

432 **Chapter 3: Rare long-term data reveal the seasonal dietary plasticity of mandrills (*Mandrillus* 433 *sphinx*) in response to fruiting tree phenology.**

434 Studies of primate feeding strategies are optimized by long-term data that allow the impact
435 of between year variation in food availability to be accounted for. In Chapter 3, I use eight
436 years of data on mandrill diets and tree phenology, to study the impact of fruit availability on
437 mandrill food selection. Using generalized additive models, I explore seasonal patterns in the
438 consumption of major food types and which foods may constitute important fallback foods. I
439 then quantify which fruit genera are consumed most frequently as a function of availability
440 and discuss why certain fruits may be preferred over others. These results are then discussed
441 in terms of why mandrill feeding strategies and preferences may have arisen.

442 **Chapter 4:** Investigating stable isotope turnover and discrimination in mandrill (*Mandrillus*
443 *sphinx*) hairs with a diet-switch experiment.

444 Stable isotope analysis is best applied using species-specific knowledge of the relationship
445 between diet and tissue stable isotope ratios. In Chapter 4, I quantify the most important
446 variables for applying stable isotope analysis to mandrill hair samples. Knowledge of hair
447 growth rates and tissue turnover times allow isotope data obtained from hairs to be placed
448 in time, increasing their ecological relevance. Tissue-diet discrimination factors allow the
449 typical enrichment between the diet and tissue isotope ratios of an animal to be accounted
450 for, which is essential for estimating diet composition from tissue stable isotope ratios. I then
451 consider the validity of the diet-switch results within the context of other controlled feeding
452 studies of primates and mammals.

453 **Chapter 5:** Stable isotope data suggest an association between extreme sexual dimorphism
454 and resource competition in mandrills (*Mandrillus sphinx*).

455 Niche divergence between sexually dimorphic males and females raises the possibility that
456 ecologically mediated selection has influenced the evolution of sexual dimorphism. In Chapter
457 5, I analyze the stable isotope ratios of mandrill hair samples, to examine sex differences in
458 isotopic niche at the levels of sex and individual. Analyzing sex-level differences in isotopic
459 niche reveals whether mandrill dimorphism is associated with dietary divergence between
460 males and females. Examining individual-level niches facilitates assessment of the impact of
461 resource competition on the diets of individual males and females. Findings are discussed by
462 considering whether resource competition may push male mandrills to leave hordes outside
463 of the breeding season.

464 **Chapter 6:** Sexual dimorphism in mandrill (*Mandrillus sphinx*) spatial behavior. A role for
465 travel costs in the evolution of sexual dimorphism?

466 Sexual dimorphism in primates is usually assumed to result from sexual selection. However,
467 males and females that differ markedly in body size may experience differing energetic
468 constraints on space use. In Chapter 6, I use data obtained from direct observational follows
469 and GPS collars to quantify sex differences in seasonal patterns of home-range use, daily path
470 length and travel speeds. More similar home-range use by each sex during the breeding

471 season, and divergence during the non-breeding season, would suggest that collared males
472 left the focal horde outside of the breeding season. If this is accompanied by shorter daily
473 path lengths and slower speeds for males during the non-breeding season, male mandrills
474 may leave social groups to reduce the energy they expend on travel when reproductive
475 opportunities are infrequent. I discuss the results within the context of Chapter 5 and other
476 recent data on the body condition of adult male mandrills, to consider whether ecologically
477 mediated selection has impacted the evolution of extreme dimorphism in mandrills.

478

479 **Chapter 2: Can diet niche partitioning enhance sexual dimorphism?**

480

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485 JB, KA, JN, DL, IJ, and LB conceived the research question and study design. JB conducted the
486 data collection. JB performed analyses and wrote the manuscript drafts with guidance from
487 KA, JN, DL, IJ, and LB. All authors contributed critically to each draft and the published
488 version formatted for this thesis.

489 **2.1 Abstract**

490 Classic evolutionary theory suggests that sexual dimorphism evolves primarily via sexual and
491 fecundity selection. However, theory and evidence are beginning to accumulate suggesting
492 that resource competition can drive the evolution of sexual dimorphism, via ecological
493 character displacement between sexes. A key prediction of this hypothesis is that the extent
494 of ecological divergence between sexes will be associated with the extent of sexual
495 dimorphism. As the stable isotope ratios of animal tissues provide a quantitative measure of
496 various aspects of ecology, we carried out a meta-analysis examining associations between
497 the extent of isotopic divergence between sexes and the extent of body size dimorphism. Our
498 models demonstrate that large amounts of between-study variation in isotopic (ecological)
499 divergence between sexes is nonrandom and may be associated with the traits of study
500 subjects. We, therefore, completed meta-regressions to examine whether the extent of
501 isotopic divergence between sexes is associated with the extent of sexual size dimorphism.
502 We found modest but significantly positive associations across species between size
503 dimorphism and ecological differences between sexes, that increased in strength when the
504 ecological opportunity for dietary divergence between sexes was greatest. Our results,
505 therefore, provide further evidence that ecologically mediated selection, not directly related
506 to reproduction, can contribute to the evolution of sexual dimorphism.

507 **2.2 Introduction**

508 *2.2.1 Sexual dimorphism*

509 Classic evolutionary theory posits that anisogamy, or variation between males and females in
510 gametic investment, causes sex differences in optimum life history and reproductive strategy
511 (Andersson, 1994; Lehtonen *et al.*, 2016). Evolution toward these sex-specific optima
512 ultimately produces phenotypic differences between males and females, known as sexual
513 dimorphism (Parker and Pizzari, 2015). One frequently observed sex difference is sexual size
514 dimorphism in which the mean body size of one sex exceeds that of the other (Fairbairn *et al.*,
515 2007). Size dimorphism is typically predicted to arise via sexual selection when the
516 reproductive success of one sex is limited by mating opportunities, and large size allows
517 individuals of that sex to increase their matings. For example, the biggest male southern
518 elephant seals (*Mirounga angustirostris*) are the best able to monopolize females at breeding
519 colonies (Le Boeuf *et al.*, 2019). Alternatively, or additionally, size dimorphism may be favored
520 by fecundity selection, when one sex's reproductive success is limited by gamete production
521 and gamete production relates to body size, such as in emydid turtles (Emydidae), in which
522 larger females produce more eggs and are therefore more fecund (Stephens and Wiens,
523 2009). It is this traditional view that size dimorphism is primarily attributable to reproductive
524 differences, which is best evidenced and dominates scientific literature and consensus
525 (Blanckenhorn, 2005).

526 However, as far back as Darwin's discussions of their "habits of life" (Darwin, 1871, p.254),
527 authors have noted that sexual dimorphisms can relate not only to each sex's reproductive
528 success but also to their respective ecologies (Shine and Goiran, 2021; Temeles *et al.*, 2000;
529 Wasiljew *et al.*, 2021). Divergence of the sexes along various biotic and abiotic niche axes
530 (e.g., temperature, diet, habitat) should facilitate intersex niche partitioning, which may
531 covary with sexual dimorphism in traits relevant to ecology (Butler *et al.*, 2007; Herrel *et al.*,
532 1999). For example, in the seabird *Sula nebouxii*, larger females dive deeper and feed on
533 larger prey than their male counterparts (Zavalaga *et al.*, 2007).

534 Such ecological sexual dimorphisms could arise as an indirect, ecological consequence of
535 dimorphism due to reproductive differences, or directly via ecological causation. An example
536 of ecological dimorphism arising as an apparent consequence of reproductive differences can

537 be found in mouthbrooding cichlid fishes. In species in which only one sex carries out
538 mouthbrooding, sex differences in diet also arise as a consequence of distinct reproductive
539 roles, as the mouthparts involved also influence foraging (Ronco *et al.*, 2019). Ecological
540 sexual dimorphisms arising as a consequence of reproductive differences are often
541 considered the most parsimonious explanation for their existence (De Lisle, 2019; Shine,
542 1989).

543

544 2.2.2 Ecological character displacement between sexes

545 Theoretical work also suggests, however, that ecological sexual dimorphisms can arise solely
546 from ecologically mediated natural selection if frequency-dependent competition for a
547 resource produces disruptive selection. Under these circumstances, the sexes could evolve
548 toward distinct phenotypes, which maximize resource acquisition, by facilitating divergence
549 along niche axes (Slatkin, 1984). This process of ecological character displacement between
550 sexes is analogous to that between nascent species and offers an alternative evolutionary
551 outcome to frequency-dependent resource competition (Bolnick and Doebeli, 2003).

552 One issue with theoretical models of ecological character displacement, however, is the lack
553 of an a priori justification for character displacement between sexes, as opposed to random
554 subsets of populations (Bolnick and Doebeli, 2003). As a result, an integrated view of
555 reproductive competition and ecological character displacement provides the best model for
556 the evolution of ecological sexual dimorphisms (De Lisle, 2019, 2021). From this perspective,
557 initial phenotypic divergence and/or differing ecological optima between the sexes may
558 usually arise due to anisogamy and reproductive differences (Maklakov *et al.*, 2008; Schärer
559 *et al.*, 2012). For instance, reproductive differences may create sex differences in nutritional
560 optima, leading to divergent foraging decisions and resource allocation (Morehouse *et al.*,
561 2020; Raubenheimer and Simpson, 2018). These initial differences may then be acted upon
562 by disruptive natural selection, such that the observed differences between sexes emerge
563 through both sexual and ecologically mediated selection. For example, intra-specific reversals
564 in the direction of python (*Morelia spilota*) size dimorphism track interpopulation differences
565 in mating systems, but the degree of size dimorphism relates to available prey sizes (Pearson
566 *et al.*, 2002).

567 The ecological character displacement hypothesis for sexual dimorphism is perhaps best
568 evidenced by a series of experiments by De Lisle and Rowe (2015) in which male and female
569 salamanders (*Notophthalmus viridescens*) were placed in semi-natural mesocosms, at various
570 competitor densities. In this species, sexual dimorphism in body size and feeding morphology
571 corresponds to sex differences in diet and microhabitat; meaning females may compete more
572 strongly with other females and males with other males. In De Lisle and Rowe's study, growth
573 rates were lower in high-density mesocosms, suggesting that competition impacts fitness for
574 which growth rate is a proxy. Furthermore, females grew faster in mesocosms with a male-
575 biased sex ratio. Females, therefore, had higher fitness when alongside a greater proportion
576 of male competitors, with which they should compete for less, suggesting that sexual
577 dimorphism somewhat reduced competition. Given such evidence, it is conceivable that
578 ecologically mediated selection can play a role in the evolution of ecological sexual
579 dimorphisms.

580

581 *2.2.3 How important is ecological character displacement?*

582 An outstanding and unresolved question concerns the importance, across taxa, of ecologically
583 mediated selection for creating diversity in sexual dimorphisms. Here we address this
584 question by investigating general associations, across species, between size dimorphism, and
585 ecological divergence in trophic level and basal carbon resources. A central tenet of the
586 ecological character displacement hypothesis is that the degree of dimorphism should scale
587 positively with the degree of ecological divergence between sexes (De Lisle and Rowe, 2015).
588 It, therefore, follows that more sexually dimorphic species should generally show greater
589 ecological divergence than less dimorphic species. Of course, associations between
590 dimorphism and ecological sex differences would not necessarily indicate causation because
591 dimorphism in any individual species may have arisen via ecological causation or as an indirect
592 consequence of reproductive differences. Associations would, however, be consistent with
593 predicted outcomes of ecological character displacement working in isolation or reproductive
594 differences creating the opportunity for disruptive ecologically mediated selection and would
595 suggest a stronger relationship between sexual dimorphism and ecology than previously
596 appreciated.

597 Ecological divergence related to size dimorphism could be exhibited by sex differences in
598 ecological mean if, for example, dimorphism impacts the prey available to each sex, leading
599 them to feed, on average, at different trophic levels (Mills *et al.*, 2021). Furthermore, sexual
600 size dimorphism could lead to sex differences in ecological variation via numerous
601 mechanisms. For example, the larger sex may be more variable if large size confers access to
602 a greater range of resources (Voigt *et al.*, 2018) or the smaller sex more variable if they are
603 competitively subordinate (Wan *et al.*, 2013). We therefore examine associations between
604 size dimorphism and ecological differences in foraging between males and females, measured
605 using stable isotope analysis. Stable isotope analysis is a common technique for analyzing
606 foraging ecology and we outline our rationale for why it is appropriate for quantifying sex
607 differences in feeding below. Our investigation uses a meta-analytic approach, synthesizing
608 previously published stable isotope data on vertebrates, with a global geographical scope.

609

610 *2.2.4 Stable isotope ecology*

611 Over recent decades, stable isotope analysis has become an effective tool used for
612 investigating animal ecology (Hobson, 1999; Hobson and Welch, 1992; Swan *et al.*, 2020).
613 Because the ratios of naturally occurring stable isotopes vary in the foods animals consume,
614 and these ratios are incorporated into animal tissues during formation, much can be revealed
615 about an animal's ecology by analyzing stable isotope ratios in its different tissues (Ponsard
616 and Ardit, 2000). Different isotopic systems provide alternative information about the animal
617 from which they are sampled (Newton, 2016), such that ratios of nitrogen stable isotopes
618 ($\delta^{15}\text{N}$, see Methods for an explanation of δ notation) vary with trophic levels (Caut *et al.*, 2009)
619 and ratios of carbon-stable isotopes ($\delta^{13}\text{C}$) vary with food chain basal resource (Farquhar *et al.*,
620 1989; Yoneyama *et al.*, 2010). For example, relative ^{15}N -enrichment of polar bears
621 compared with seals indicate polar bears occupy a higher trophic level (Hobson *et al.*, 2002)
622 and $\delta^{13}\text{C}$ can distinguish the diets of zebras and giraffes that feed on C4 and C3 plants,
623 respectively (Codron *et al.*, 2006). Combined stable isotope ratios of animal tissues thus allow
624 inferences about the individual niche, meaning ecological differences can be quantified at
625 various levels, including niche differences between males and females, with the greater sex

626 differences in stable isotope ratios taken to indicate more ecological divergence (Foote *et al.*,
627 2012; Lehmann *et al.*, 2015).

628

629 *2.2.5 Is size dimorphism associated with isotopic sex differences?*

630 Because stable isotope data tend to be reported reasonably consistently across taxa,
631 compared with other measures of diet, the considerable stable isotope ecology literature
632 provides an opportunity to investigate cross-species associations between sexual dimorphism
633 and ecological (isotopic) sex differences. Our study achieves three main aims. First, using
634 meta-analytic models, we quantify between-study variation in isotopic sex differences in the
635 stable isotope literature and the fraction of this variation constituting heterogeneity (I^2). In
636 the context of meta-analysis, heterogeneity describes the amount of observed between-
637 study variation in effect size that is due to nonrandom variation in true effect size, as opposed
638 to random sampling variation (Borenstein *et al.*, 2017). As a consequence, heterogeneity also
639 indicates the fraction of between-study variation that may be explained by predictor
640 variables, such as the traits of study subjects.

641 Second, having discovered substantial heterogeneity among studies, we next investigate how
642 much heterogeneity in isotopic sex differences can be explained by size dimorphism. We use
643 meta-regression models including size dimorphism as a predictor variable, to examine
644 associations with isotopic sex differences, and interpret the strength of associations as an
645 indicator of the amount of heterogeneity in ecological sex differences that are explained by
646 sexual dimorphism.

647 We use size dimorphism as a predictor variable in our analyses despite our interest in its
648 response to certain ecological contexts. Our choice is primarily pragmatic: the diversity of the
649 stable isotope literature means we can readily compute effect sizes using means, errors, and
650 sample sizes for isotopic data of both sexes in many species. In contrast, body size data are
651 most commonly available as mean values, and therefore more suitable as a predictor variable.
652 Our choices also moderate the sensitivity of our meta-analyses. Stable isotopes may not
653 capture all ecological differences between sexes; for example, male and female birds may
654 feed on different seeds, which would not manifest as trophic level differences. Similarly, the

655 sexes may differ in trophic structures and feed on different diets, while being the same body
656 mass, which would be missed by our measure of size dimorphism. However, our sacrifice of
657 some of this detail allowed us to maximize the taxonomic scope and therefore the generality
658 of our results. It also means that our meta-analyses are conservative in nature and that
659 associations between sexual dimorphism and ecological sex differences may be stronger than
660 we detect here.

661

662 *2.2.6 How important is the ecological context?*

663 Our final aim was to examine whether associations between sexual dimorphism and
664 ecological sex differences are modified by the ecological context. We first test whether
665 species' dietary class and/or mean species size modify relationships between size dimorphism
666 and isotopic sex differences. Dietary class may modify the impact of size dimorphism on
667 isotopic sex differences because species consuming different diets vary in their ecological
668 flexibility. For example, omnivores by definition feed at more trophic levels than herbivores,
669 which could create more opportunity for size dimorphism to exert an influence on the trophic
670 level of each sex. Mean species size (defined here as the mean of males and females) may
671 influence the effect of size dimorphism by causing between-species differences in resource
672 access, which may then impact how size dimorphism affects resource use within species. For
673 example, if size affects the maximum prey size available to each sex, size dimorphism may
674 have a greater impact on smaller species that are already more limited concerning the size of
675 their prey. Conversely, the greater absolute size of larger species may mean proportional size
676 differences between males and females have more impact on their respective interactions
677 with other food web members. We quantify the potential influence of species' dietary class
678 and mean size by including them as additional predictor variables, alongside size dimorphism,
679 in meta-regression models.

680 Another possibility is that size dimorphism has the greatest ecological impact on carnivores
681 that are gape-limited, meaning they can only consume prey smaller than themselves (Shine,
682 1991; Shine *et al.*, 2003). For example, in an aquatic food chain formed of gape-limited fish,
683 each species can consume all species smaller than itself, but no species the same size or
684 larger. A trophic level should therefore closely track body size, with the largest fish at the

685 highest trophic level. If a fish species in such a food chain were size dimorphic, the larger sex
686 would have greater access to larger, higher trophic level prey, than the smaller sex, resulting
687 in a difference between males and females in the maximum possible trophic level. As optimal
688 foraging theory predicts that predators often feed preferentially on larger prey, due to greater
689 energetic returns per prey item (Dodrill *et al.*, 2021; Stephens and Krebs, 1986), the larger sex
690 in a dimorphic gape-limited fish would be predicted to feed at a higher trophic level.
691 Conversely, non-gape-limited predators and scavengers, such as cats, can consume prey
692 orders of magnitude larger than themselves, which may minimize any impact of size
693 dimorphism on the trophic levels of each sex. Therefore, we investigate whether gape
694 limitation strengthens associations between size dimorphism and isotopic sex differences.
695 We do this using a meta-regression on a data set constrained to fish and snake species, which
696 are presumed to be able to feed solely by swallowing whole prey and thus considered gape-
697 limited. The predictions from this model are then compared with those from a model
698 containing all other carnivores in our data set, to assess whether the effect of size dimorphism
699 on trophic sex differences is greater in gape-limited carnivores.

700

701 **2.3 Methods**

702 Our meta-analytic approach and reporting were completed with reference to the guidelines
703 laid out by O'Dea *et al.* (2021).

704

705 *2.3.1 Data collection*

706 We collated peer-reviewed literature available in the Web of Science Core Collection. The
707 stable isotope literature is large, with the search term “stable isotope” returning ~76,500
708 studies at the time of writing. To constrain the search, we combined the following specific
709 terms, using the default publication year range of 1900–2020, on 10/11/2020: Isotop* Nich;
710 Isotop Nich* Male; Isotop* Nich* Female; Isotop* Nich* Male Female; Isotop* Nich* Sex
711 Diff*; Isotop Nich* Dimorph; Isotop Dimorph*.

712 Our searches returned 3489 studies, which we placed into a spreadsheet to highlight
713 duplicates for manual removal. Removing duplicates resulted in 2807 studies for the title and
714 abstract screening. At this stage, we made the decision to constrain our analysis to the
715 nitrogen and carbon stable isotope systems, due to the relatively small number of studies
716 using other systems that were returned by our search terms. We also rejected studies during
717 the title and abstract screening if they did not use bulk stable isotope analysis, used samples
718 of human, museum, archeological or palaeontological origin, were review, comment, or
719 method papers, or if the animals sampled were not wild, not adults, not vertebrates or if data
720 were not available for both sexes.

721 We then searched the remaining 1279 studies using the ctrl + F search function and,
722 separately, the terms “sex”, “male” and “female”, excluding studies if they contained none of
723 these terms, under the assumption that they did not contain stable isotope ratios for each
724 sex and, if at least one term was present, checking for the presence of the required data.
725 Additional reasons for exclusion were if the full text was inaccessible without purchase or
726 contacting authors, presented incomplete data (mean, error, or sample size missing), was not
727 in English, or was a paper correction.

728 We then attempted to extract data from the remaining 210 studies. Additional reasons for
729 exclusion at this stage were if raw data were presented as images with >50 rows, if data were
730 from an earlier study already included or if data extraction from figures was not possible. We
731 extracted data from figures using a mouse pointer to individually select data points from an
732 image of the figure, with the image calibrated to the axis values from the original figure;
733 therefore, too much point overlap made this process inaccurate, because not all points could
734 be selected for inclusion. The entire process provided 173 studies in which mean, standard
735 deviation, and sample sizes for each sex were presented in the manuscript, or could be
736 calculated from raw data, or could be taken from model outputs, or extracted from figures
737 (Figure 2.1). We collected data for any vertebrate species, from any global location and, if
738 stable isotope ratios for each sex were presented for more than one tissue type, we entered
739 each tissue as a separate row in our database.

740

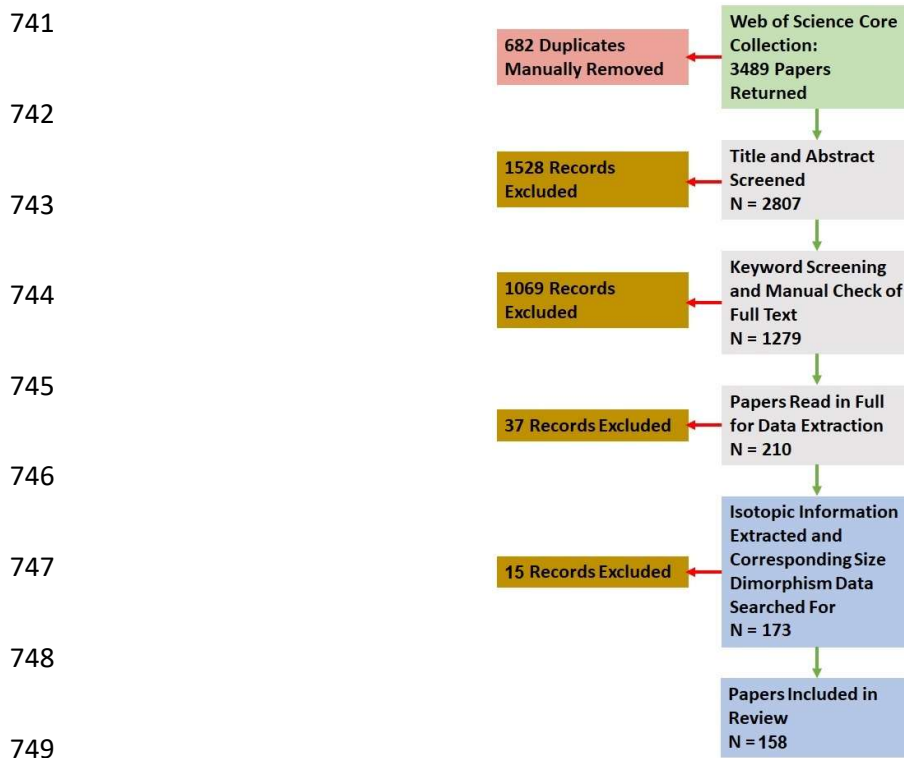


Figure 2.1. Our sequence of study collation, screening, and data extraction, alongside the number of studies excluded at each stage and included in the final analysis.

750

751 2.3.2 Effect size choice and calculation

752 All the stable isotope data we collected were presented in permil units (‰). Permils units
 753 describe enrichment or depletion of heavy isotopes, relative to international standards, which
 754 exist for nitrogen, carbon, and all other commonly used systems in stable isotope analysis.
 755 For example, the standard for nitrogen stable isotope ratios is atmospheric nitrogen (air). The
 756 relative enrichment or depletion is expressed using delta (δ) notation, such that

757

$$\delta = \frac{(R_{sample} - R_{standard})}{R_{standard}}$$

758 where R_{sample} equals the isotope ratio of the sample and $R_{standard}$ equals the isotope ratio of
 759 the standard (McKinney *et al.*, 1950) where R = heavy isotope/light isotope, for example,
 760 $^{15}\text{N}/^{14}\text{N}$. Thus, a positive δ value indicates enrichment in the heavier isotope (in this case ^{15}N)
 761 and a negative value indicates isotopic depletion of the sample, relative to the international

762 standard. When comparing two sampling units, such as sex, a more positive (or less negative)
763 value for one sex indicates enrichment in the heavy isotope relative to both the standard and
764 the other sex. It is this difference between sexes in isotopic enrichment that we have used to
765 calculate the effect sizes in our meta-analysis.

766 All studies from which we extracted data expressed stable isotope data in permil units,
767 therefore further standardization of effects sizes was not necessary (Nakagawa and Cuthill,
768 2007; Nakagawa and Santos, 2012). We calculated mean differences between male and
769 female stable isotope ratios as the raw mean difference between isotopic means of each sex,
770 as found in each study. We calculated these with a positive mean difference indicating that
771 males were ¹⁵N- or ¹³C-enriched compared with females and a negative mean difference
772 indicating females were isotopically enriched compared with males. For example, a positive
773 $\delta^{15}\text{N}$ mean difference indicates that males feed, on average, at a higher trophic level.

774 Regarding meta-analyses of variation, two effect size measures are often recommended in
775 the fields of ecology and evolution, which are the log Variability Ratio (lnVR) and the log
776 Coefficient of Variation Ratio (lnCVR; Senior *et al.*, 2020). lnCVR has the advantage of allowing
777 for mean–variance relationships in effect sizes (i.e., an increase in variance with mean value;
778 Senior *et al.* 2016): our data did not show any evidence of such relationships (Appendix 1). In
779 addition, because lnCVR accounts for variation in mean value by expressing absolute variation
780 as a proportion of group mean, sex differences in isotopic variation could actually be
781 misrepresented through this standardization. For example, if one sex is twice as enriched
782 relative to the international standard as the other and shows twice as much variation, lnCVR
783 would express this as an equal variation. Conversely, as lnVR is calculated using the raw
784 variation values for each group, with no accounting for mean values, the more enriched sex
785 would also be shown to be twice as a variable, more accurately representing each sex's
786 ecology. We, therefore, selected lnVR as our effect size metric for sex differences in variation.

787 Finally, we selected mass dimorphism (kg) as our measure of size dimorphism, as this measure
788 was most readily available for the highest number of vertebrates, allowing the taxonomic
789 scope of our analysis to be as wide as possible. Body masses for each sex were established
790 using web searches, prioritizing data from peer-reviewed scientific studies, followed by
791 published books and, if necessary, taxon-specific websites. As reliable body mass data could

792 not be obtained for all species, the number of studies in the analysis was reduced to 158. A
793 list of data sources used in our meta-analyses is provided in the Data Sources section.

794

795 *2.3.3 Model choice and structure*

796 All models used complete case analysis, meaning rows containing missing data for predictor
797 variables or effect sizes (dependent variables) were removed from the analysis. We used
798 multilevel meta-analytic models to quantify the amount of between-study variation that
799 exists for each isotope and effect size and how much of this variation constitutes
800 heterogeneity, as opposed to random sampling variation. We then used multilevel meta-
801 regression models to investigate whether the heterogeneity found could be explained by our
802 predictor variables. To investigate the relationship between sexual dimorphism and ecology,
803 we examined the strength of associations between size dimorphism and sex differences in
804 isotopic mean and variance for carbon and nitrogen.

805 To examine whether ecological context can modify the relationship between dimorphism and
806 ecology, we ran meta-regressions using sex differences in mean nitrogen isotope as the
807 response variable, with additional predictor variables, dietary class, and mean species size,
808 included alongside size dimorphism. Dietary class constituted a categorical variable with three
809 levels: carnivore, omnivore, and herbivore. Mean size was a continuous variable, calculated
810 by averaging the male and female mass data used for calculating size dimorphism. We ran
811 models including dietary class and mean size separately and together, with and without all
812 combinations of two- and three-way interactions. Three-way interactions were theoretically
813 justified because, if mean species size could modify the effect of size dimorphism on the
814 trophic level, this modification may be more apparent in species with more inherent trophic
815 flexibility, such as omnivores, than those with less, such as herbivores. In addition, the sample
816 size was large enough that the number of parameters to be estimated with three-way
817 interactions did not prevent models from converging. The best models were identified using
818 AICc scores, with lower scores taken to indicate better models (Arnold, 2010).

819 Initially, we only considered a modifying effect of dietary class on nitrogen mean sex
820 differences because carbon sex differences did not exhibit a significant association with size

821 dimorphism across our entire data set. However, this could be expected to exclude
822 herbivores, which likely feed at only one trophic level, from showing an association. We,
823 therefore, conducted an additional model examining whether dietary class modified the
824 association between size dimorphism and sex differences in mean carbon stable isotope ratio.
825 This model balances our analyses because carbon isotopes in terrestrial systems are primarily
826 influenced by plant photosynthetic mechanisms and therefore represent a niche axis along
827 which the sexes could more easily diverge in herbivores.

828 As a final test of the association between size dimorphism and ecological sex differences, we
829 quantified the effect of gape limitation by running two meta-regressions including only size
830 dimorphism as a predictor variable and limiting the data sets to gape-limited and non-gape-
831 limited carnivores, in which size dimorphism may have differing impacts on sex differences in
832 trophic level.

833 Residuals of all models were approximately normally distributed; thus, no data
834 transformations were used.

835

836 *2.3.4 Random effects*

837 In all the above multilevel models, we included study identity and species as random factors,
838 to account for random sampling variation at both these levels and to adequately account for
839 pseudoreplication, since we potentially considered measures for several tissue samples from
840 the same specimens. We also included phylogeny as a random factor, to account for
841 relatedness between the species included in our data set, following the method of Sanchez-
842 Tojar *et al.* (2020).

843

844 *2.3.5 Publication bias and sensitivity analysis*

845 Scientific literature may be subject to publication bias, whereby favorable results are
846 preferentially published, thus skewing the results of meta-analyses. We produced funnel plots
847 to identify such biases, by visualizing the distribution of published effect sizes and determining
848 whether there are missing observations that might be expected in the literature (based on

849 variation in effect sizes). In addition, biases may arise when research builds upon influential
850 results from poor quality or low power studies, leading to reduced effect sizes through time,
851 as the true effect is quantified with repetition or higher quality studies. To test for such
852 publication bias in isotopic sex differences, we ran meta-regressions using sex differences in
853 isotopic mean and variance, for carbon and nitrogen, as the dependent variable and
854 publication year as the only predictor variable.

855 The results of meta-analyses may also be sensitive to decisions about the weights assigned to
856 individual studies as well as to high-influence data points (Koricheva *et al.*, 2013). Meta-
857 analytic models usually account for both within-study variance and between-study variance
858 when assigning weights to individual study results. However, when between-study variance
859 is high, within-study variance can be masked when weighting studies, potentially impacting
860 model results. We, therefore, ran additional models using only the inverse of within-study
861 variance to assign study weights, to determine the influence of our choice of weighting
862 parameter. Finally, to analyze the sensitivity of our models to high-influence data points, we
863 completed a leave-one-out analysis, to calculate Cook's distances for each data point and ran
864 additional models with high-influence data points removed. The results of our tests of
865 publication bias, alternate study weighting and sensitivity analysis, alongside justifications for
866 final data inclusion and model choices can be found in the Appendix.

867

868 2.3.6 Software

869 All data processing, analyses, and plotting were completed using R v4.0.2. We used the R
870 package “metaDigitise” v1.0.1 (Pick, Nakagawa and Noble, 2019) for all data extraction from
871 figures and the package “metafor” v.2.4-0 (Viechtbauer, 2010) to calculate all effect sizes and
872 to run all meta-analytic and meta-regression models. Our phylogeny was constructed using
873 the “rotl” v3.0.12 (Michonneau *et al.*, 2016) and “ape” v5.6.2 (Paradis and Schliep, 2019)
874 packages and we calculated the phylogenetic signal with “phylosignal” v1.3 (Keck *et al.*, 2016)
875 and “phylobase” v0.8.1 (Hackathon, 2020). We created all plots using the R package “ggplot2”
876 v3.3.3 (Wickham, 2016) and tables using “flextable” v0.7.3 (Gohel *et al.*, 2020).

877 **2.4 Results**

878 *2.4.1 Dataset*

879 Our final database contained isotopic information from 158 studies, covering 163 species.
880 Mammals were the most common taxa (n = 68), followed by birds (n = 60), fish (n = 18),
881 reptiles (n = 17), and a single amphibian. The species with the greatest female-biased
882 dimorphism was the northern map turtle (*Graptemys geographica*), in which females are 10×
883 the mass of males and the species with the greatest male-biased dimorphism was the
884 elephant seal (*Mirounga leonina*), with males seven times larger than females. The number
885 of effect sizes used in the analyses was highest for $\delta^{15}\text{N}$ mean sex differences (n = 282),
886 followed by $\delta^{13}\text{C}$ mean differences (n = 276), $\delta^{15}\text{N}$ lnVR (n = 272), and $\delta^{13}\text{C}$ lnVR (n = 266).

887 *2.4.2 Quantifying heterogeneity in between-sex isotopic differences*

888 Between-study variation was found for sex differences in mean $\delta^{15}\text{N}$ (trophic level) and $\delta^{13}\text{C}$
889 (food chain basal carbon resource) and sex differences in $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ variation (Figure 2.2).
890 The amount of heterogeneity (I^2) was 90.57% and 94.38% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ mean sex
891 differences, respectively. Such high heterogeneity indicates that almost all between-study
892 variation in effect size is nonrandom and has the potential to be explained by predictor
893 variables. Regarding sex differences in isotopic variation, heterogeneity was 64.2% and
894 72.83% for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, respectively, indicating that the majority of between-study
895 variation in between-sex differences in an isotopic variation also has the potential to be
896 explained by predictor variables. In the case of nitrogen, one sex was at least twice as variable
897 as the other in 8.5% of cases and for carbon, in 13.1% of cases (Figure 2.2).

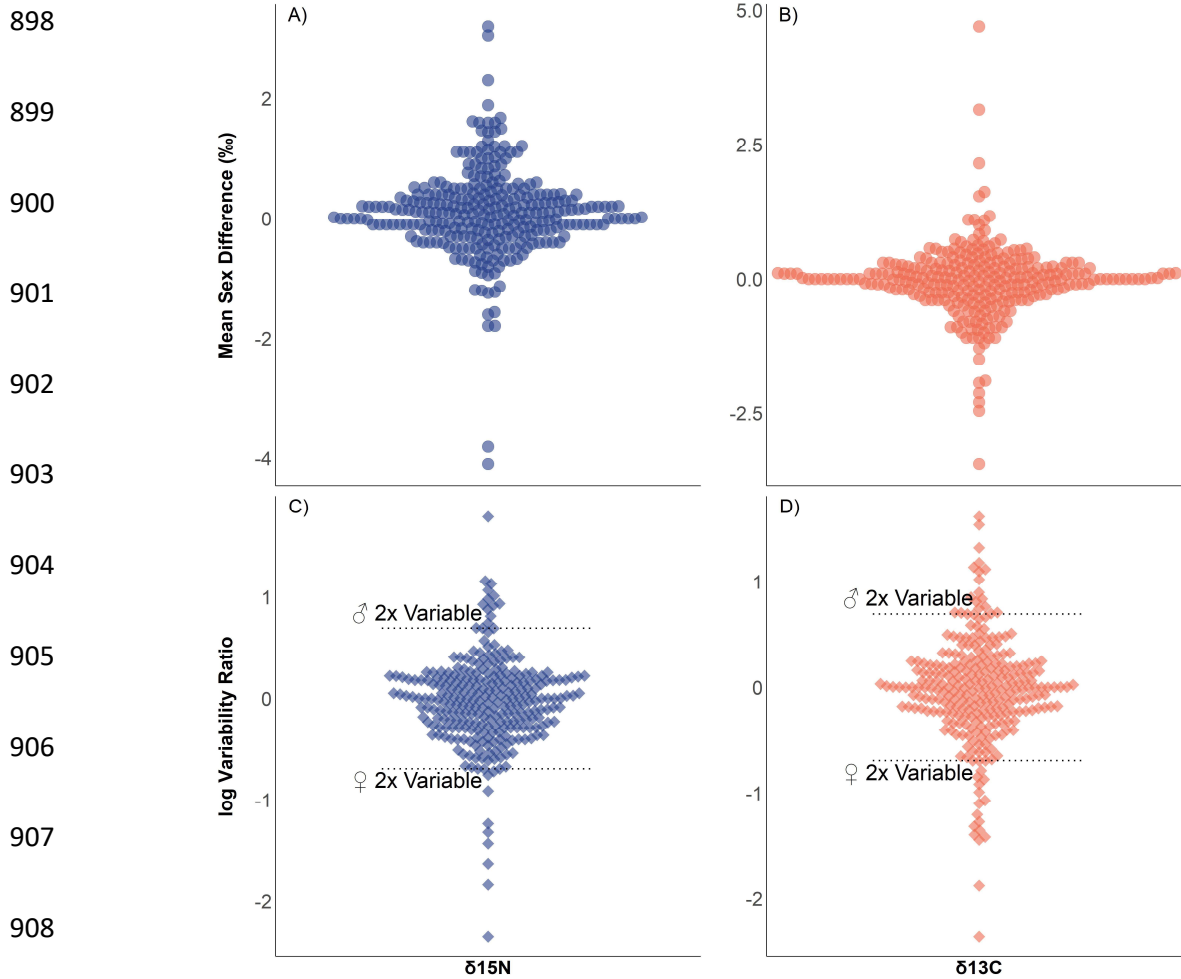


Figure 2.2 Published stable isotope ratio sex differences for nitrogen mean (a), carbon mean (b), nitrogen variation (c), and carbon variation (d). In (a, b), each point represents the raw difference between male and female mean stable isotope ratio, of one tissue of one species. Positive values indicate higher values in males, whereas negative values indicate higher values in females. In (c, d), each point represents the log male:female variability ratio of one tissue, in one species. Positive values indicate males showed more isotopic variation and those above the dotted line indicate that males were more than twice as variable as females. Negative values indicate females showed more isotopic variation and those below the dotted line indicate that females were more than twice as variable as males. X-axes constitute one category, with jitter added to better visualize overlapping observations.

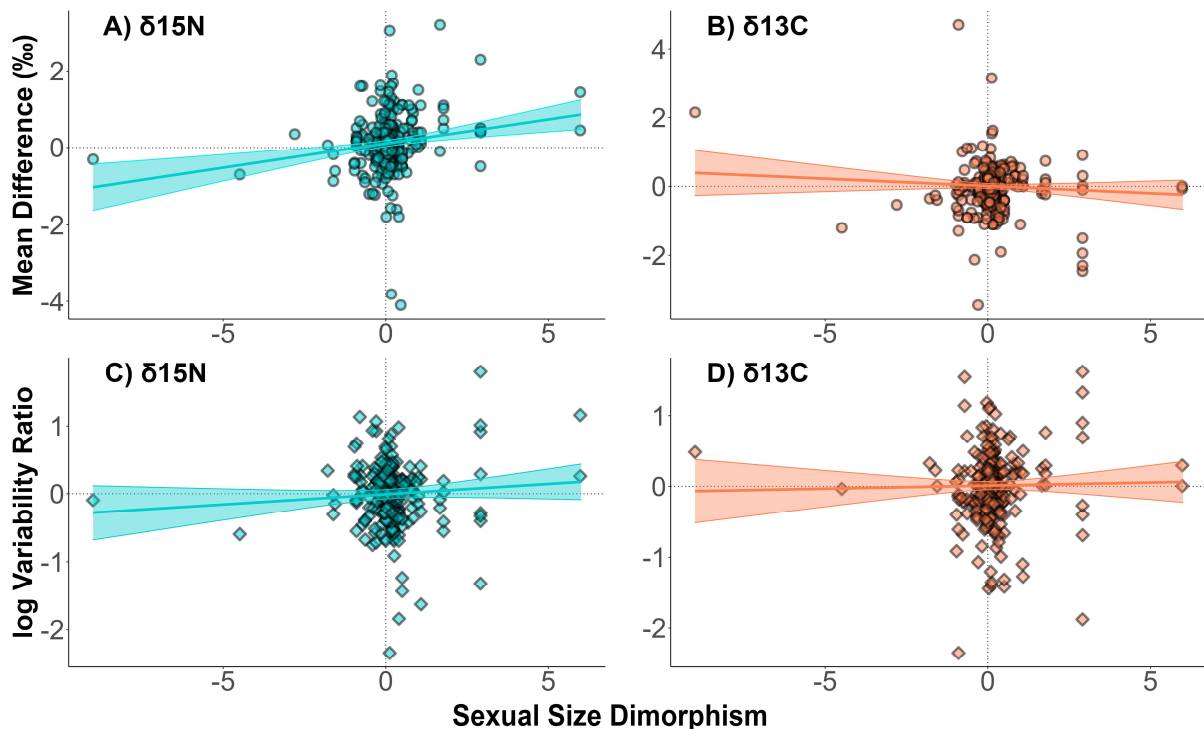
909

910

911 2.4.3 Associations between isotopic sex difference and size dimorphism

912 Associations between isotopic sex difference and size dimorphism To examine the possibility
913 that size dimorphism relates to ecology, we carried out meta-regressions containing size
914 dimorphism as the sole predictor variable and found modest or nonexistent associations with
915 isotopic sex differences. The estimated effect size of size dimorphism on $\delta^{15}\text{N}$ mean difference
916 was significantly positive (mean = 0.126, 95% CI: 0.06–0.19, $p < .001$, indicating that a size
917 dimorphism of 100% led to a $\delta^{15}\text{N}$ increase of 0.126‰, on average. This effect was modest
918 compared with the variation in isotopic sex differences in our data set (range $\delta^{15}\text{N}$ sex
919 difference: -4.1 to 3.2 ‰). Models of relationships between size dimorphism and $\delta^{13}\text{C}$ mean
920 differences, $\delta^{15}\text{N}$ variation, and $\delta^{13}\text{C}$ variation produced estimate confidence bands that
921 overlapped zero, indicating no significant associations between size dimorphism and these
922 measures of isotopic sex differences. The predictions from these models, alongside their
923 underlying raw data, are visualized in Figure 2.3.

924



925 *Figure 2.3. The relationship between sexual size dimorphism and stable isotope sex differences*
926 *in nitrogen mean (a), carbon mean (b), nitrogen variation (c), and carbon variation (d). Lines*
927 *and 95% confidence intervals are based on meta-regression predictions. Data points are raw*
928 *data, overlaid to visually assess how well size dimorphism explains isotopic sex differences.*

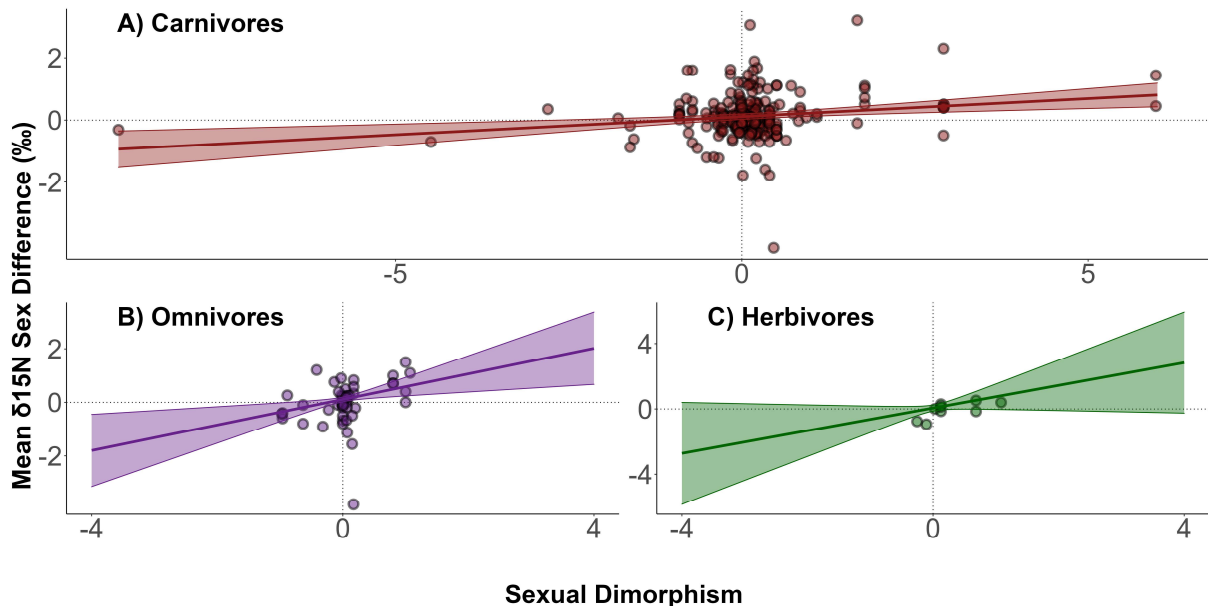
929 2.4.4 Quantifying the effects of ecological context

930 To test the possibility that associations between size dimorphism and feeding vary in strength
 931 among ecological contexts, we used meta-regressions to test whether dietary class, mean
 932 size, or gape limitation modified the effect of size dimorphism on isotopic sex differences.
 933 Model selection using AICc scores indicated that a model containing species mean size and
 934 dietary class as predictors of $\delta^{15}\text{N}$ sex differences, with an interaction between size
 935 dimorphism and dietary class, improved model fit (AICc = 512), compared with the size
 936 dimorphism-only model above (AICc = 518; Table 2.1). The best model contained an
 937 interaction between size dimorphism and dietary class, such that the association between
 938 size dimorphism and sex differences in nitrogen mean was statistically nonsignificant in
 939 herbivores, significant and moderate in carnivores (0.17, 95% CI: 0.053–0.18, $p < .001$), and
 940 significant and strongest in omnivores (0.36, 95% CI: 0.019–0.7, $p = .038$; Figure 2.4). The
 941 model also contained a significant effect of mean species size on sex differences in nitrogen
 942 mean (0.0000036, CI: 0.00000023–0.0000069, $p = .036$).

943 *Table 2.1. AIC scores for models examining associations between size dimorphism and $\delta^{15}\text{N}$*
 944 *mean sex differences in different ecological contexts.*

Model Formula	AIC
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism * Dietary Class + Species Mean Size	512.0546
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism * Species Mean Size + Dietary Class	513.2787
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism * Species Mean Size	513.8847
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism * Dietary Class * Species Mean Size	513.8942
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism * Dietary Class	514.3551
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism + Dietary Class * Species Mean Size	515.1587
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism + Dietary Class + Species Mean Size	515.3208
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism + Species Mean Size	515.9071
$\delta^{15}\text{N}$ Mean Sex Difference ~ Size Dimorphism + Dietary Class	517.6389

945



947 *Figure 2.4. The relationship between sexual size dimorphism and mean nitrogen stable isotope*
948 *ratio sex differences in carnivores (a), omnivores (b), and herbivores (c). Lines and 95%*
949 *confidence intervals are based on meta-regression predictions. Data points are raw data,*
950 *overlaid to visually assess how well size dimorphism explains trophic sex differences in each*
951 *dietary class.*

952

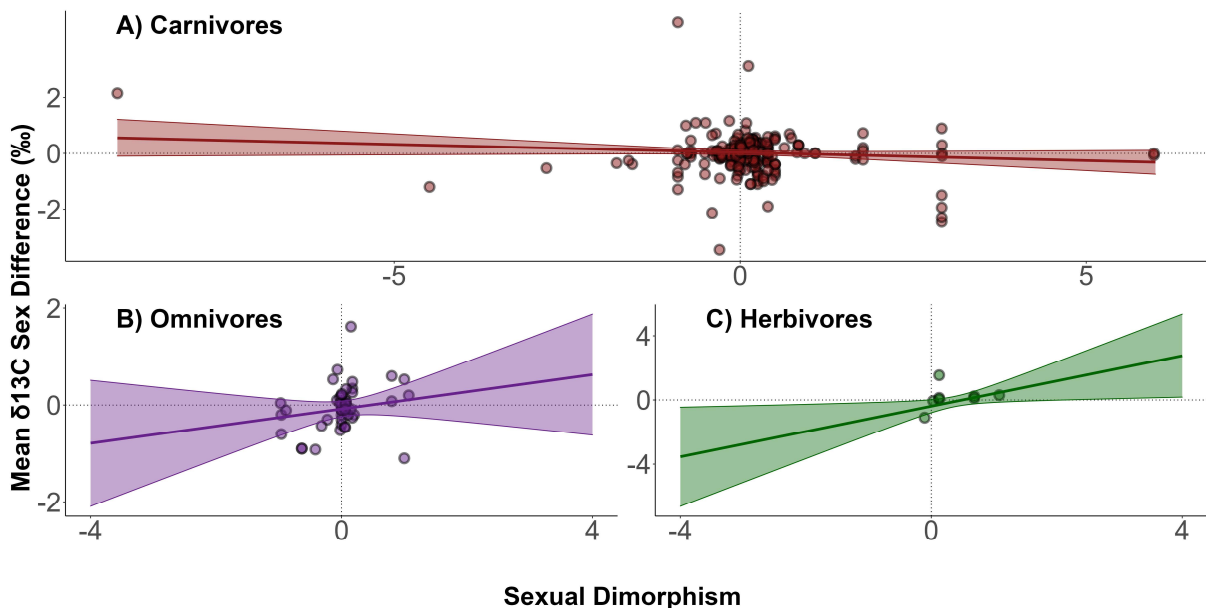
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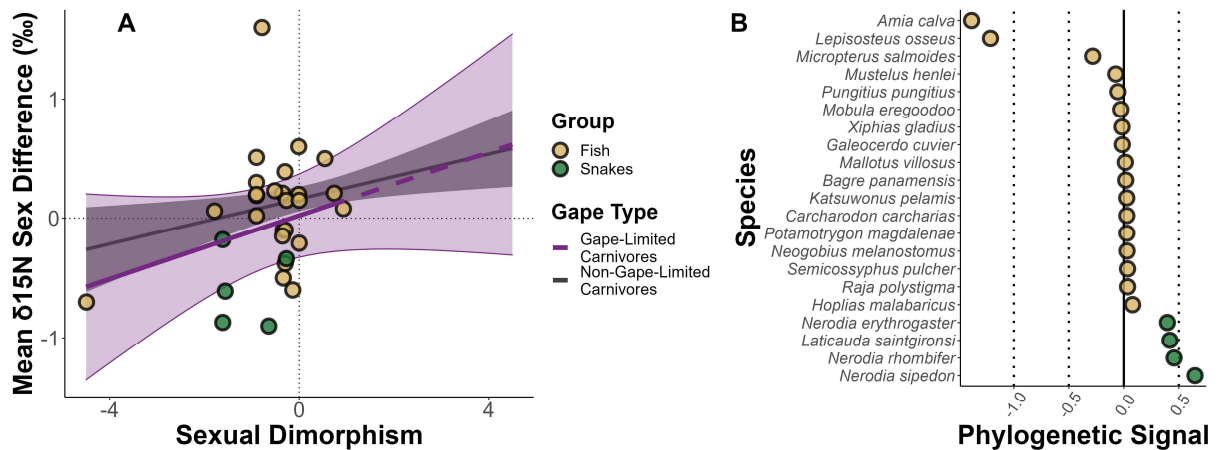
956

957 Regarding sex differences in $\delta^{13}\text{C}$, including diet alongside size dimorphism produced results
 958 that contrasted with $\delta^{15}\text{N}$. In the case of carbon, carnivores and omnivores instead exhibited
 959 nonsignificant associations between size dimorphism and isotopic sex differences, whereas
 960 herbivores exhibited a significant positive association (0.847, 95% CI: 0.139–1.555, $p = .02$;
 961 Figure 2.5).



962 *Figure 2.5. The relationship between sexual size dimorphism and mean carbon stable isotope*
 963 *ratio sex differences in carnivores (a), omnivores (b), and herbivores (c). Lines and 95%*
 964 *confidence intervals are based on meta-regression predictions. Data points are raw data,*
 965 *overlaid to visually assess how well size dimorphism explains trophic sex differences in each*
 966 *dietary class.*

967 Limiting the data set to only gape-limited predators, in which trophic level is predicted to
 968 relate more closely to body size, resulting in a 41% increase in the estimated effect of size
 969 dimorphism on $\delta^{15}\text{N}$ mean difference. However, the effect was nonsignificant because of the
 970 more modest sample size (0.133, 95% CI: -0.0412 to 0.306, $p = .135$), relative to non-gape-
 971 limited (0.094, 95% CI: 0.023–0.17, $p = .01$; Figure 2.6a). Our data set contained local
 972 phylogenetic signals for $\delta^{15}\text{N}$ sex differences in gape-limited predators, with positive
 973 phylogenetic signal in snake species (Figure 2.6b), controlling for which weakened the
 974 association between size dimorphism and trophic sex differences in gape-limited predators.
 975 A nonphylogenetic meta-regression showing a stronger association may be found in Appendix
 976 1.



977 Figure 2.6. (a) The relationship between sexual size dimorphism and mean nitrogen stable
 978 isotope ratio sex difference in gape-limited and non-gape-limited carnivores. Lines and 95%
 979 confidence intervals are based on meta-regression predictions. Data points are raw data for
 980 gape-limited carnivores, overlaid to assess how well size dimorphism explains trophic sex
 981 differences in gape-limited carnivores. Predictions based on gape-limited species are shown in
 982 purple (the prediction line is dashed outside the raw data range) and for the non-gape-limited
 983 carnivores in gray. (b) Local indicators of phylogenetic association (phylogenetic signal) for
 984 mean nitrogen sex differences in gape-limited carnivores.

985

986 2.5 Discussion

987 Using meta-analytical methods, we examined cross-species relationships between sexual size
 988 dimorphism and ecological divergence between sexes, measured using stable isotopes. We
 989 found that across 166 globally distributed species, size dimorphism was weakly associated
 990 with sex differences in mean trophic level ($\delta^{15}\text{N}$), but not mean food chain basal resource
 991 ($\delta^{13}\text{C}$), or variation in either isotope. We also found associations between size dimorphism
 992 and isotopic sex differences to be modified by ecological context. The effect of size
 993 dimorphism on sex differences in mean trophic level was strongest in omnivores, more
 994 modest in carnivores, absent in herbivores, and very modestly affected by species mean size.
 995 Dietary class influenced the effect of size dimorphism on sex differences in food chain basal
 996 resource in an inverse manner to trophic level, as an association was found in herbivores, but
 997 not in omnivores or carnivores. Finally, we found partial evidence that sex differences in

998 trophic level could be more strongly associated with size dimorphism in gape-limited than in
999 non-gape-limited carnivores, as predicted by theory.

1000

1001 *2.5.1 Heterogeneity exists in isotopic sex differences*

1002 We found heterogeneity in between-sex differences in isotopic mean and variance, for
1003 nitrogen and carbon isotopes. Heterogeneity in sex differences for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ indicates
1004 nonrandom between-study variation in the extent to which males and females feed at
1005 different trophic levels and in different food chains, respectively. Our meta-analytic models,
1006 therefore, indicated that large amounts of between-study variation in ecological sex
1007 differences may be explained by study-level variables. Though we found some isotopic sex
1008 differences to be associated with size dimorphism, which we discuss presently, a large
1009 amount of variation was unexplained. Though we found some isotopic sex differences to be
1010 associated with size dimorphism, which we discuss presently, a large amount of variation was
1011 unexplained. This unexplained variation remains open to explanation by further analyses,
1012 some suggestions for which we also outline below. We hope our database provides a useful
1013 starting point for further investigations of sexual dimorphism and ecological differences
1014 between males and females (Dryad: <https://doi.org/10.5061/dryad.k98sf7m99>).

1015

1016 *2.5.2 Across all species, size dimorphism is associated with sex differences in trophic level, but* 1017 *not food chain basal carbon resources or ecological variability*

1018 We found evidence that sexual size dimorphism does, in some instances, scale positively with
1019 ecological sex differences. Size dimorphism exhibited a moderate positive association with
1020 sex differences in mean $\delta^{15}\text{N}$ and thus explained some variation in trophic level differences
1021 between males and females. Previous cross-species investigations of the relationship
1022 between size dimorphism and $\delta^{15}\text{N}$ sex differences have often found inconsistencies, with
1023 size dimorphism sometimes relating to trophic differences and sometimes not (Mancini *et al.*,
1024 2013; Phillips *et al.*, 2011). These inconsistencies are likely because the effect is probably
1025 modest and may be modified by a wide range of factors, as evidenced by the amount of
1026 unexplained variation in our data set. Thus, the size and scope of our analysis is likely the

1027 factor that has allowed us to find a clear but moderate effect of size dimorphism on trophic
1028 level differences between sexes. No relationships existed between size dimorphism and
1029 variation in $\delta^{13}\text{C}$ mean sex differences across all species or sex differences in variation in either
1030 isotope. Size dimorphism thus does not appear to have a general cross-species influence on
1031 sex differences in food chain basal carbon resources or trophic variability. These results
1032 conflict somewhat with the conclusions of studies on individual species (Calado *et al.*, 2020;
1033 Voigt *et al.*, 2018). It may therefore be the case that size dimorphism is related to sex
1034 differences in basal carbon resources and trophic variability in particular species, populations,
1035 or contexts, but this association is not sufficiently consistent to covary with sex differences
1036 across species.

1037

1038 *2.5.3 Associations between size dimorphism and isotopic sex differences are modified by* 1039 *dietary class and gape limitation, but not mean size*

1040 We found that ecological context influenced associations between size dimorphism and
1041 ecological sex differences. Dietary class modified the effect of size dimorphism on trophic
1042 level ($\delta^{15}\text{N}$) differences between males and females, with no effect of dimorphism in
1043 herbivores, a moderate effect in carnivores, and the strongest effect in omnivores. In
1044 contrast, we found an association between size dimorphism and sex differences in food chain
1045 basal resource ($\delta^{13}\text{C}$) in herbivores, but not in omnivores or carnivores. These results are
1046 consistent with our prediction that inherent differences in ecological flexibility might alter the
1047 association between size dimorphism and ecology and make sense given the respective
1048 feeding niches of the three dietary classes.

1049 As herbivores would be expected to feed exclusively on plants, they should only occupy the
1050 position of the primary consumer, leaving little scope for size dimorphism to influence trophic
1051 level. Instead, herbivores can more easily diverge along the niche axis related to plant
1052 consumption. Because most herbivores in our data set occupy terrestrial habitats, plant
1053 consumption is reflected in carbon isotopes in their tissues. Thus, if size dimorphism relates
1054 to dietary divergence in herbivores, this divergence seems to arise as sex differences in plant
1055 consumption, as opposed to trophic level. For example, stable isotopes suggest that male
1056 African elephants (*Loxodonta africana*) consume more grass than females, which may be

1057 because larger body size increases food digestive efficiency and food ingestion per mouthful
1058 (Shannon *et al.*, 2013).

1059 Carnivores may occupy any position from secondary consumer upwards, meaning each sex
1060 could take prey from one or multiple trophic levels, creating some opportunity in carnivores
1061 for trophic differences between the sexes to be influenced by size dimorphism. In contrast to
1062 the other two groups, omnivores would be expected to consume foods from a minimum of
1063 two trophic levels, leading to a greater probability of trophic differences between sexes and
1064 the largest opportunity for size dimorphism to influence this difference. This might be why
1065 size dimorphism was most strongly associated with trophic sex differences in omnivores in
1066 our data set. Sex differences in carnivore and omnivore basal resources may have covaried
1067 less with size dimorphism because the basal resources of nonprimary consumers relate not
1068 only to their diet but also to the diets of their prey (Codron *et al.*, 2018). Therefore, unless
1069 dimorphism produces sex differences in prey access that corresponds to different basal
1070 resources, for example, one sex exclusively hunting browsers and the other hunting grazers,
1071 a strong association between dimorphism and basal resource sex differences is unlikely. The
1072 indirect nature of the relationship between size dimorphism and basal resource in carnivores
1073 and omnivores could mean that divergence along this niche axis is less likely to be driven by
1074 dimorphism than in herbivores. Together, these results suggest that the association between
1075 size dimorphism and ecological sex differences is dependent on the ecological opportunity
1076 for feeding differences between males and females.

1077 Our model also contained an independent effect of species mean size on sex differences in
1078 trophic levels. However, this effect was multiple orders of magnitude lower than the
1079 analytical reproducibility of nitrogen stable isotope analysis. Thus, even though the model
1080 estimate was technically positive, we conclude that absolute species size has no meaningful
1081 effect on trophic-level sex differences.

1082 Limiting the data set to gape-limited carnivores led to a 41% increase in the effect of size
1083 dimorphism on trophic sex differences, compared with non-gape-limited carnivores, though
1084 the gape-limited model was nonsignificant. There is substantial evidence available that larger
1085 gape increases maximum ingestible prey size and trophic level in fish and snakes (Barnes *et al.*
1086 *al.*, 2021; Nilsson and Brönmark, 2000; Persson *et al.*, 1996; Webb and Shine, 1993).

1087 Consequently, the larger effect of size dimorphism on trophic sex differences in the gape-
1088 limited predators we analyzed could indicate that gape limitation increases the ecological
1089 relevance of dimorphism, producing a stronger relationship between dimorphism and
1090 ecology. More research is required, however, due to the non-significance of our model. The
1091 wide confidence intervals in the gape-limited model may have been due to the small sample
1092 size ($n = 30$) and any potential interaction between gape limitation, size dimorphism, and
1093 ecological sex differences should therefore be examined with a larger data set.

1094 An additional possibility is that gape limitation is important to the relationship between size
1095 dimorphism and ecological sex differences only when certain conditions are met. There was
1096 a phylogenetic signal for trophic-level sex differences in our gape-limited data (Figure 2.6b)
1097 and in nonphylogenetic models, the effect of size dimorphism in gape-limited predators was
1098 significant and substantially greater than when controlling for phylogeny (Appendix). It could
1099 be the case that the phylogenetic random effect captured differences between species in the
1100 niches or food webs to which they have adapted. If gape limitation is to produce trophic sex
1101 differences in a species, then body size must relate to the trophic level in the prey of that
1102 species, in order for higher trophic level prey to be accessible to only the larger sex. This
1103 condition may not hold true in all taxa or food webs, and this wider context may need to be
1104 considered when investigating the influence of gape limitation on associations between
1105 sexual dimorphism and trophic sex differences.

1106 In summary, we found the extent of size dimorphism to be associated with the extent of
1107 ecological sex differences, an association that increased in strength when the ecological
1108 opportunity for trophic variation was greatest and potentially when size dimorphism was
1109 more ecologically relevant. Our results are consistent with both a potential role for ecological
1110 character displacement in enhancing sexual dimorphism or for ecological sex differences to
1111 arise due to reproductive differences. However, as the associations we found were generally
1112 modest, they support previous predictions that the role of ecological character displacement
1113 is relatively minor (De Lisle and Rowe, 2015; Fairbairn, 1997) and suggest that size
1114 dimorphism produced via reproductive differences is not a powerful driver of feeding
1115 differences between males and females.

1116 2.5.4 Future Research Directions

1117 Several questions arise from our meta-analysis that should form the subject of future work.
1118 First, the results of our analysis on gape limitation were uncertain and the apparent influence
1119 of phylogeny on our results could suggest that the relevance of gape limitation is species or
1120 niche specific. Future analyses could therefore examine the impact of gape limitation on
1121 ecological sex differences with a greater sample size and taxonomic scope than used
1122 presently. For example, a recent analysis found a positive relationship between gape size and
1123 fruit size consumed in frugivorous birds (McFadden *et al.*, 2022), and including similar data in
1124 future analyses could allow the importance of gape limitation to ecological sex differences to
1125 be generalized across a wide range of species and niches.

1126 As our results suggest a relationship between sexual dimorphism and ecology but do not
1127 distinguish between reproductive differences and ecological character displacement as
1128 driving that relationship, the next major question concerns their relative importance as
1129 evolutionary mechanisms. A weak role for ecologically mediated selection implies that sexual
1130 and fecundity selection are the main drivers of sexual dimorphism. However, a recent analysis
1131 found that size dimorphism was only weakly associated with sexual selection across species,
1132 leading the authors to suggest that “alternative mechanisms such as ecological character
1133 displacement may be crucial to understand the full diversity of [size dimorphism] in animals”
1134 (Janicke and Fromonteil, 2021). Considering our own results, alongside their conclusion, we
1135 suggest that a future priority should be the incorporation of sexual, fecundity, and ecologically
1136 mediated selection into single cross-species analyses, to quantify their relative importance to
1137 the evolution of size dimorphism.

1138 Our analyses support an association but suggest the cross-species patterns may be modest.
1139 Why would ecological character displacement be a weaker selective force than other drivers
1140 of sexual dimorphism? One possibility is that the frequency-dependent nature of resource
1141 competition means that the strength of competition falls as the sexes phenotypically diverge
1142 (De Lisle and Rowe, 2015). Alternatively, divergence from the species mean phenotype, while
1143 alleviating resource competition, may itself entail fitness costs that eventually exceed those
1144 of competition for resources (Bolnick and Doebeli, 2003; Slatkin, 1984). Either possible
1145 scenario may place an upper limit on the extent to which ecologically mediated selection can

1146 drive character displacement between sexes. Therefore, establishing the mechanistic
1147 limitations on ecological character displacement between sexes should also become the focus
1148 of future investigations, most likely via modeling and experiment.

1149 A final question is what additional variables could be included in future analyses, to explain
1150 the considerable variation in ecological sex differences? Our analyses have highlighted that
1151 high amounts of between-study variation in our data remain unexplained, providing
1152 opportunities to use our database to investigate additional drivers of ecological differences
1153 between males and females. Importantly, isotopic values for an animal's tissues may be
1154 affected by many factors, such as body size, body condition, diet quality, and ontogenetic
1155 growth (Carleton and Martinez del Rio, 2010; Lecomte *et al.*, 2011; Wolf *et al.*, 2009). Sex
1156 differences in any of these variables could potentially influence sex differences in isotopic
1157 signals and influence cross-species isotopic comparisons. However, their impact is often
1158 species-specific, so a comparative synthesis of the sort we have conducted would require
1159 species-level data to become widely available across many taxa.

1160 In relation to why ecological sex differences evolve, nutritional requirements are one
1161 potential avenue of investigation. Males and females may target distinct sets of resources in
1162 order to meet sex specific nutritional needs, such as lactation or sexually selected signals
1163 (Harrison *et al.*, 2017; Thompson, 2013). These differences may influence foraging and other
1164 aspects of behavioral ecology (Morehouse *et al.*, 2020). As stable isotopes in animal tissues
1165 vary with the foods animals consume, the sex differences in isotope ratio we have observed
1166 may illustrate how males and females target distinct resources, to fulfill their own sex-specific
1167 nutritional requirements. Future investigations could therefore seek to quantify the strength
1168 of associations between-sex differences in nutritional requirements and stable isotope
1169 values, which could contribute greatly to our understanding of ecological differences between
1170 males and females.

1171 **2.6 Acknowledgements**

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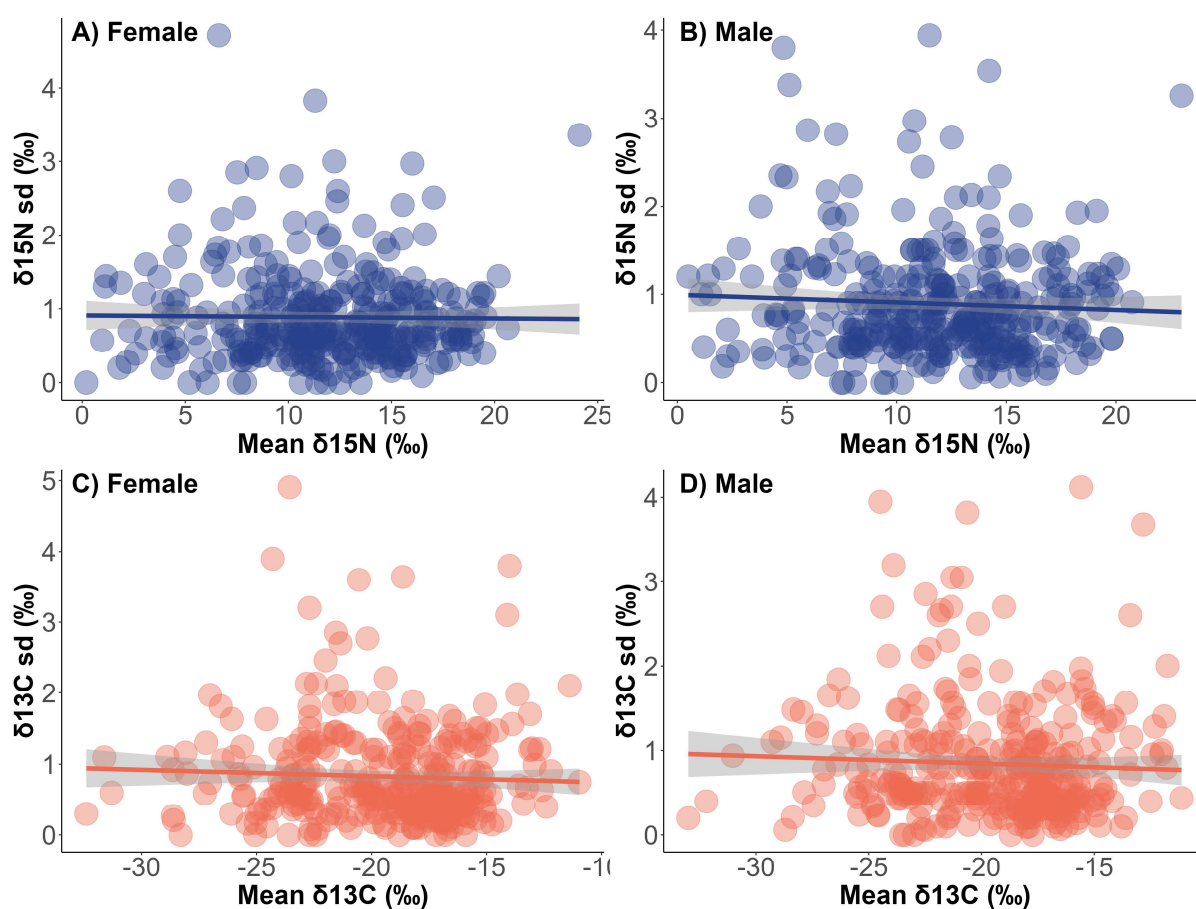
1174 [NERC IAPETUS DTP (NE/L002590/1)]. IJ was supported by a UKRI Future Leaders Fellowship
1175 (MR/T019018/1).

1176

1177 2.7 Appendix

1178 2.7.1 Mean-Variance Relationships

1179 Calculating effect sizes that quantify variation differences between groups, such as sex, may
1180 be impacted by mean–variance relationships. These describe an increase in variance with an
1181 increase in mean and may adversely influence the outcome of meta-analyses. We, therefore,
1182 plotted the relationship between raw isotopic mean and standard deviation, alongside linear
1183 regressions, to investigate the presence of mean–variance relationships in our data. We found
1184 no evidence for mean–variance relationships in female nitrogen (Figure A2.1a), male nitrogen
1185 (Figure A2.1b), female carbon (Figure A2.1c), or male carbon (Figure A2.1d).



1186 *Figure A2.1. Mean–variance relationships in female nitrogen(a), male nitrogen (b), female*
1187 *carbon (c), and male carbon (d).*

1188 2.7.2 Publication bias

1189 Publication Year

1190 As one possible identifier of publication bias is a reduction of effect sizes through time, we
 1191 completed meta-regressions with publication year as the sole predictor variable. We found
 1192 no effect of publication year on the magnitude of published sex differences in isotopic mean
 1193 or variation, for either nitrogen or carbon isotopes (Table A2.1).

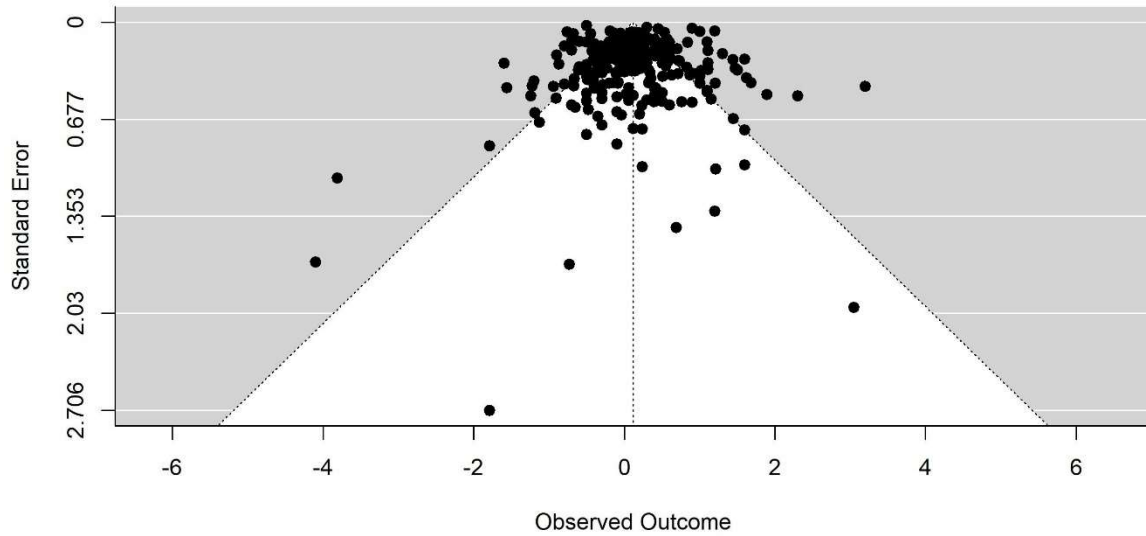
1194 *Table A2.1. Effect of publication year on sex difference in mean and variation, for nitrogen*
 1195 *and carbon.*

Isotope	Measure	Effect of Publication Year	Standard Error	Confidence Limit Lower Bound	Confidence Limit Upper Bound	<i>p</i>
Nitrogen	Mean Difference	-0.0010	0.0100	-0.0221	0.0202	0.9270
Nitrogen	Variation	-0.0035	0.0066	-0.0164	0.0095	0.6010
Carbon	Mean Difference	0.0066	0.0081	-0.0094	0.0225	0.4195
Carbon	Variation	-0.0094	0.0073	-0.0237	0.0049	0.1979

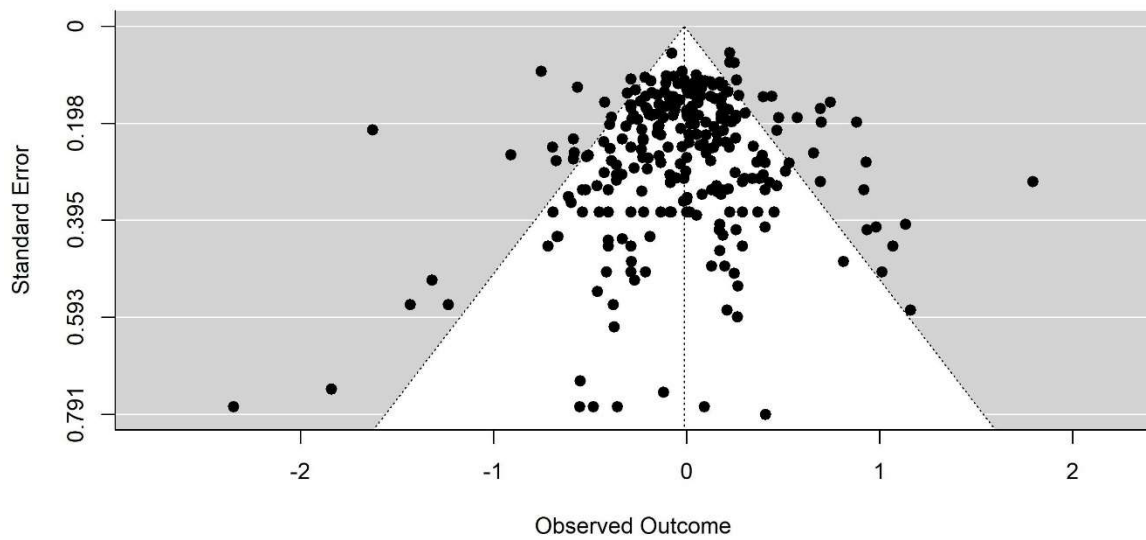
1196 Funnel plots

1197 Funnel plots can be used to investigate possible publication bias by illustrating asymmetries
 1198 in published effect sizes, which would suggest particular results are favorably published. Such
 1199 biases in published literature would influence the outcome of meta-analyses, by skewing
 1200 summary effect size estimates toward the favored outcome. We, therefore, produced funnel
 1201 plots displaying published effect sizes for sex differences in nitrogen mean (Figure A2.2),
 1202 nitrogen variation (Figure A2.3), carbon mean (Figure A2.4), and carbon variation (Figure

1203 A2.5). In all four cases, our plots displayed a fairly even distribution in study outcomes,
1204 suggesting that publication bias is not prominent in the literature we have examined.



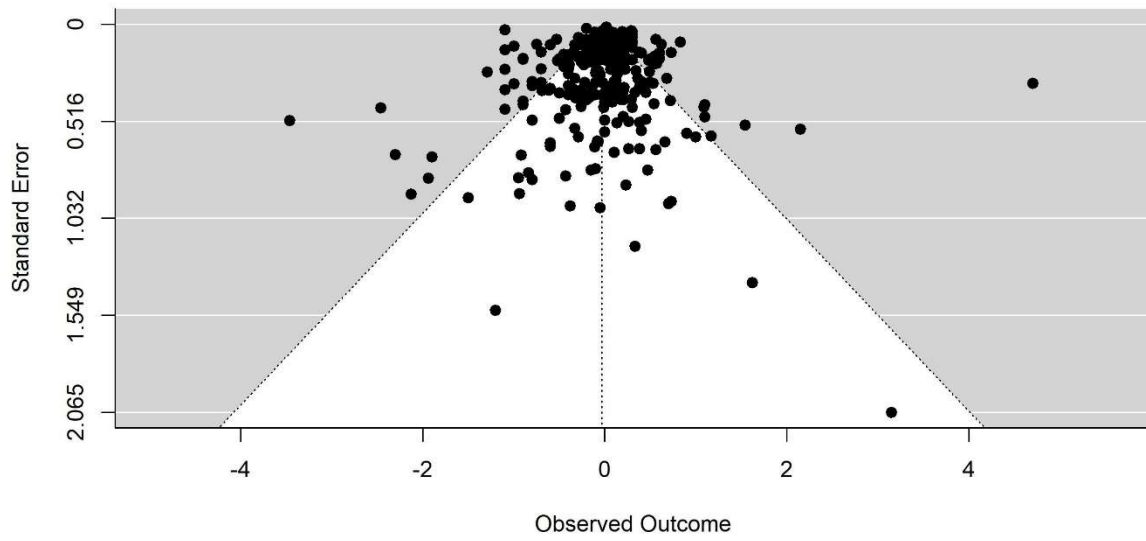
1205 *Figure A2.2. Distribution of published sex differences in nitrogen isotope means.*



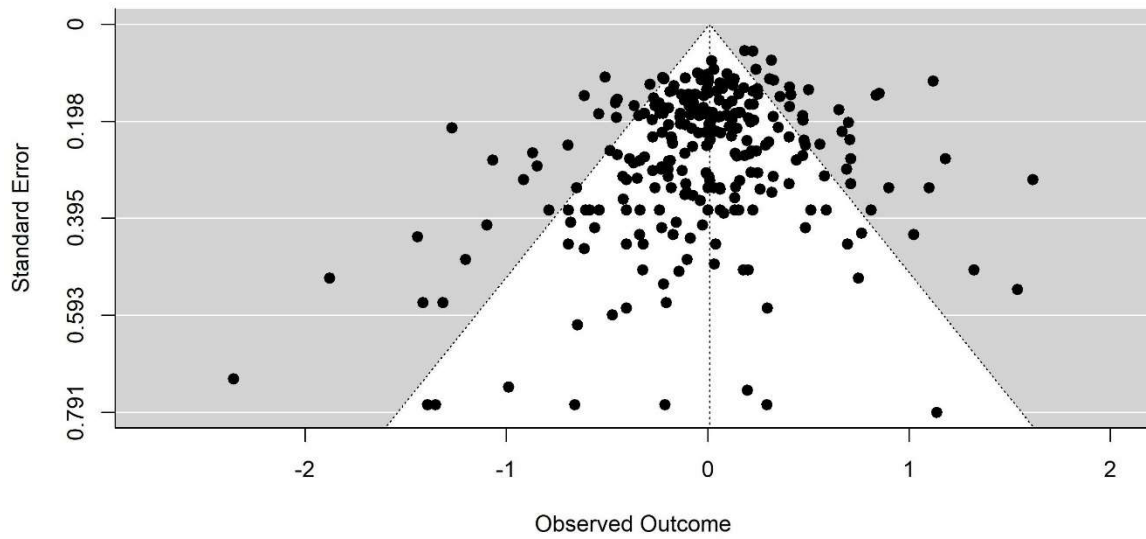
1206 *Figure A2.3. Distribution of published sex differences in nitrogen isotope variation.*

1207

1208



1209 *Figure A2.4. Distribution of published sex differences in carbon isotope means.*



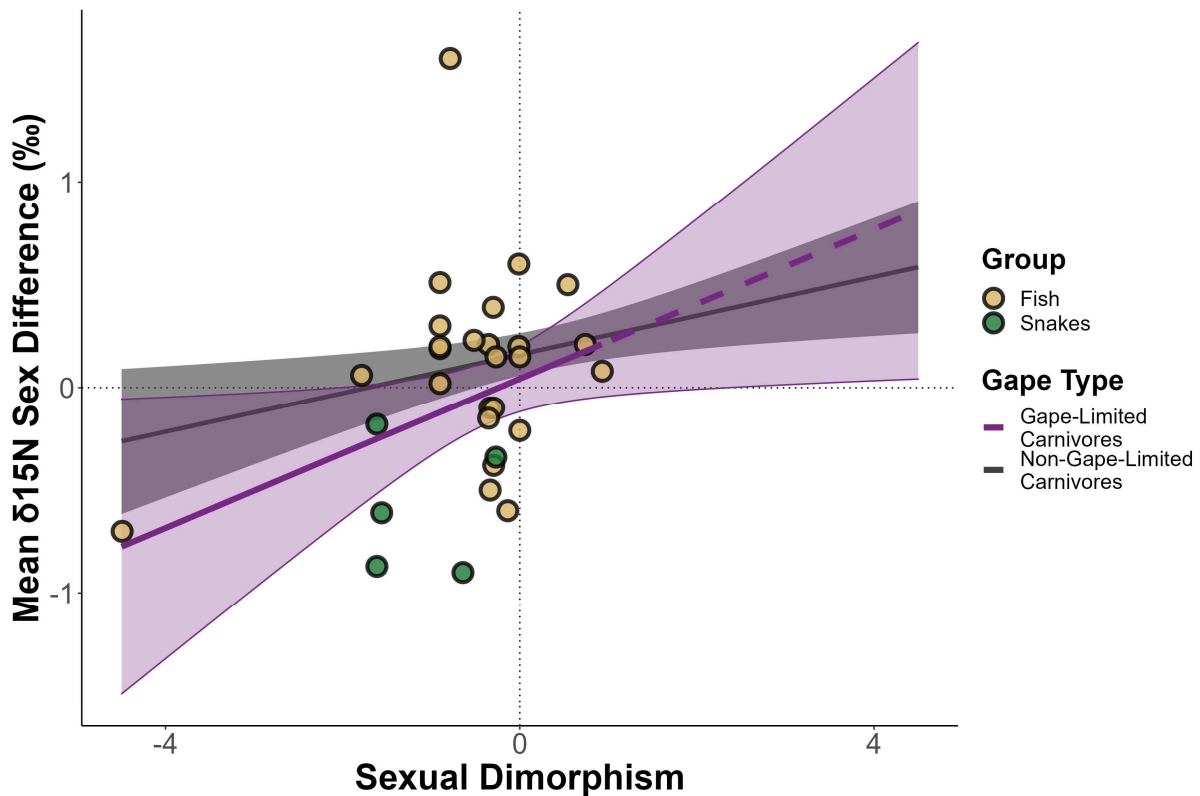
1210 *Figure A2.5. Distribution of published sex differences in carbon isotope variation.*

1211

1212

1213 2.7.3 Gape limitation: non-phylogenetic models

1214 When not controlling for phylogeny the effect of size dimorphism in gape-limited predators
1215 was statistically significant and almost double that of non-gape-limited carnivores (Figure
1216 A2.6).



1217 Figure A2.6. The effect of size dimorphism in gape-limited predators (purple) and non-gape-
1218 limited carnivores (gray) when not controlling for phylogenetic relatedness between species.

1219

1220 2.7.4 Effect of study weighting method

1221 The weight given to individual effect sizes can alter the outcome of meta-analyses. As our
1222 analysis consisted of meta-regressions, including the random factors “paper number,”
1223 “species,” and “phylogeny,” each effect size used as a response in our analysis was weighted
1224 accounting for within-study variance, heterogeneity between studies, species, and
1225 phylogenetic relatedness, and covariance between these random factors. Thus, our models
1226 assumed differences between studies and species in the true isotopic difference between
1227 sexes. However, high heterogeneity, which was present in our data, can mask within-study
1228 variance. It is therefore recommended to also conduct models weighting studies solely by the

1229 inverse of within-study variance, to examine the impact of weighting method on model
1230 predictions and, therefore, the conclusions of the meta-analysis.

1231 Regarding the relationship between-sex differences in nitrogen mean (trophic level) and size
1232 dimorphism, our results were not robust to changing the weighting method, as the confidence
1233 interval for the estimated effect of size dimorphism on trophic sex differences overlapped
1234 zero (Table A2.2). Our qualitative conclusion would therefore have changed with the alternate
1235 weighting, to state that size dimorphism is unrelated to trophic differences between sexes.
1236 However, as this weighting method does not account for between-study and between-species
1237 heterogeneity in trophic sex differences, we believe it to be inappropriate. The diversity of
1238 species investigated by the studies we have meta-analyzed, and the consequent diversity of
1239 our analysis, mean assuming a universal difference in sex differences in trophic level is clearly
1240 erroneous. We are therefore skeptical of the conclusion this weighting method produces and
1241 are more confident in the original model included in the main text.

1242 When weighting by the inverse of within-study variance in the model examining the effect of
1243 size dimorphism on sex difference in nitrogen variation (Table A2.3), carbon mean (Table
1244 A2.4), and carbon variation (Table A2.5), our qualitative conclusions remained the same.

1245 When including dietary class and mean size alongside size dimorphism, as predictors of sex
1246 differences in nitrogen mean, the effect of mean size (which was effectively zero) was absent
1247 and the effect of size dimorphism was absent in all dietary classes when weighting by the
1248 inverse of within study variance (Table A2.6). Our results in models examining gape-limited
1249 (Table A2.7) and non-gape-limited carnivores (Table A2.8) were also not robust to alternate
1250 weighting. Finally, the effect of size dimorphism on sex differences in carbon mean sex
1251 differences was also statistically nonsignificant when using the alternate weighting method
1252 (Table A2.9). However, for the same reasons outlined above, we are more confident in the
1253 original models included in the main text.

1254

1255

1256 *Table A2.2. Output of fixed effects model examining the effect of size dimorphism on sex*
 1257 *differences in nitrogen mean, weighting studies only by the inverse of within-study variance.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.046	0.091	0.51	0.61
Size Dimorphism	summary	0.1	0.062	1.63	0.1

1258 *Table A2.3. Output of fixed effects model examining the effect of size dimorphism on sex*
 1259 *differences in nitrogen variation, weighting studies only by the inverse of within-study*
 1260 *variance.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	-0.0071	0.041	-0.17	0.86
Size Dimorphism	summary	0.039	0.03	1.35	0.18

1261 *Table A2.4. Output of fixed effects model examining the effect of size dimorphism on sex*
 1262 *differences in carbon mean, weighting studies only by the inverse of within-study variance.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	-0.051	0.11	-0.58	0.62
Size Dimorphism	summary	0.14	0.08	1.84	0.066

1263 *Table A2.5. Output of fixed effects model examining the effect of size dimorphism on sex*
 1264 *differences in carbon variation, weighting studies only by the inverse of within-study*
 1265 *variance.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.044	0.039	1.13	0.26
Size Dimorphism	summary	0.030	0.033	0.91	0.36

1266 *Table A2.6. Output of fixed effects model examining the effect of size dimorphism, dietary*
 1267 *class and species mean size on sex differences in nitrogen mean, weighting studies only by*
 1268 *the inverse of within-study variance.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.21	0.076	2.76	0.0059

Size Dimorphism	summary	0.059	0.06	0.99	0.32
Herbivore	summary	-0.71	0.3	-2.38	0.017
Omnivore	summary	-0.53	0.28	-1.89	0.059
Mean Size	summary	0.00000097	0.0000027	0.37	0.71
SSD:Herbivore	summary	0.65	0.55	1.19	0.23
SSD:Omnivore	summary	0.38	0.27	1.43	0.15

1269 *Table A2.7. Output of fixed effects model examining the effect of size dimorphism on sex*
1270 *differences in nitrogen mean, in gape-limited carnivores, weighting studies only by the*
1271 *inverse of within-study variance.*

Term	Type	Estimate	Standard Error	Statistic	<i>p</i>
Intercept	summary	0.11	0.2	0.55	0.59
Size Dimorphism	summary	0.11	0.16	0.73	0.47

1272

1273 *Table A2.8. Output of fixed effects model examining the effect of size dimorphism on sex*
1274 *differences in nitrogen mean, in non-gape-limited carnivores, weighting studies only by the*
1275 *inverse of within-study variance.*

Term	Type	Estimate	Standard Error	Statistic	<i>p</i>
Intercept	summary	0.23	0.087	2.6	0.0094
Size Dimorphism	summary	0.054	0.066	0.82	0.41

1276 *Table A2.9. Output of fixed effects model examining the effect of size dimorphism on sex*
1277 *differences in nitrogen mean, in non-gape-limited carnivores, weighting studies only by the*
1278 *inverse of within-study variance.*

Term	Type	Estimate	Standard Error	Statistic	<i>p</i>
Intercept	summary	0.030	0.07	0.43	0.67
Size Dimorphism	summary	0.087	0.074	1.14	0.26
Herbivore	summary	-0.89	0.3	-3.01	0.0026
Omnivore	summary	-0.086	0.28	-0.31	0.76
Size Dimorphism:Herbivore	summary	1.3	0.47	2.8	0.0051
Size Dimorphism:Omnivore	summary	0.28	0.21	1.33	0.18

1279 *2.7.5 High leverage data points*

1280 The outcome of meta-analyses may also be adversely impacted by outliers/high leverage data
 1281 points that skew model estimates. We therefore used Cook's leave-one-out analysis to
 1282 identify high leverage data points that may have an unduly large effect on our models. Several
 1283 approaches are possible for identifying data points as high leverage based on Cook's scores,
 1284 and we chose to assign those with a Cook's score over three times the mean score, for data
 1285 points in a given model, as potentially high leverage. We found this approach to be the most
 1286 conservative, by identifying the highest number of points as possibly high leverage. We then
 1287 removed these data from the models for which they may be high leverage and re-ran each
 1288 model.

1289 In models examining only the effect of size dimorphism on sex differences in nitrogen mean
 1290 (Table A2.10), nitrogen variation (Table A2.11), carbon mean (Table A2.12), and carbon
 1291 variation (Table A2.13), removing high leverage data points did alter estimated effect sizes
 1292 but did not change the qualitative conclusions we could draw from the models. When
 1293 including dietary class and species mean size as predictors, alongside size dimorphism, the
 1294 effect sizes changed (Table A2.14), but our qualitative conclusions did not. In models
 1295 analyzing the impact of size dimorphism on sex differences in nitrogen mean in gape-limited
 1296 (Table A2.15) and non-gape-limited carnivores (Table A2.16), removing high leverage data
 1297 points also did not change our qualitative conclusions. Finally, when removing high leverage
 1298 data points from the model examining the effect of size dimorphism and diet on sex
 1299 differences in carbon mean sex differences, the effect of size dimorphism in herbivores was
 1300 no longer significant (Table A2.17). This is likely because of the small sample size of herbivores
 1301 in our data set, rather than an indication that any points should be removed.

1302 *Table A2.10. Output of meta-regression model examining the effect of size dimorphism on*
 1303 *sex differences in nitrogen mean, with high leverage data points removed.*

Term	Type	Estimate	Standard Error	Statistic	<i>p</i>
Intercept	summary	0.076	0.039	1.95	0.051
Size Dimorphism	summary	0.2	0.053	3.57	0.00036

1304 *Table A2.11. Output of meta-regression model examining the effect of size dimorphism on*
 1305 *sex differences in nitrogen variation, with high leverage data points removed.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	-0.012	0.029	-0.43	0.67
Size Dimorphism	summary	-0.0048	0.033	-0.15	0.88

1306 *Table A2.12. Output of meta-regression model examining the effect of size dimorphism on*
 1307 *sex differences in carbon mean, with high leverage data points removed.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.022	0.03	0.74	0.46
Size Dimorphism	summary	0.063	0.039	1.63	0.1

1308 *Table A2.13. Output of meta-regression model examining the effect of size dimorphism on*
 1309 *sex differences in carbon variation, with high leverage data points removed.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	-0.0027	0.027	-0.1	0.92
Size Dimorphism	summary	0.031	0.028	1.11	0.27

1310 *Table A2.14. Output of meta-regression model examining the effect of size dimorphism and*
 1311 *dietary class and species mean size on sex differences in nitrogen mean, with high leverage*
 1312 *data points removed.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.1	0.044	2.35	0.019
Size Dimorphism	summary	0.18	0.051	3.45	0.00055
Herbivore	summary	0.032	0.46	0.069	0.95
Omnivore	summary	-0.097	0.11	-0.9	0.37
Species Mean Size	summary	0.0000064	0.0000025	2.62	0.0088
Size Dimorphism: Herbivore	summary	0.15	0.64	0.24	0.81
Size Dimorphism: Omnivore	summary	0.44	0.23	1.93	0.054

1313 *Table A2.15. Output of meta-regression model examining the effect of size dimorphism on*
 1314 *sex differences in nitrogen mean, in gape-limited carnivores, with high leverage data points*
 1315 *removed.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.066	0.15	0.45	0.66
Size Dimorphism	summary	0.025	0.096	0.26	0.79

1316 *Table A2.16. Output of meta-regression model examining the effect of size dimorphism on*
 1317 *sex differences in nitrogen mean, in non-gape-limited carnivores, with high leverage data*
 1318 *points removed.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.12	0.054	2.19	0.029
Size Dimorphism	summary	0.19	0.077	2.53	0.011

1319 *Table A2.17. Output of meta-regression model examining the effect of size dimorphism and*
 1320 *dietary class on sex differences in carbon mean, with high leverage data points removed.*
 1321 *Output of meta-regression model examining the effect of size dimorphism and dietary class*
 1322 *on sex differences in carbon mean, with high leverage data points removed.*

Term	Type	Estimate	Standard Error	Statistic	p
Intercept	summary	0.014	0.036	0.38	0.7
Size Dimorphism	summary	0.028	0.043	0.66	0.51
Herbivore	summary	0.45	0.49	0.91	0.36
Omnivore	summary	-0.11	0.086	-1.28	0.2
Size Dimorphism: Herbivore	summary	-0.41	0.9	-0.46	0.65
Size Dimorphism: Omnivore	summary	0.29	0.19	1.51	0.13

1323

1324 **Chapter 3 Rare long-term data reveal the seasonal dietary plasticity**
1325 **of mandrills (*Mandrillus sphinx*) in response to fruiting tree**
1326 **phenology.**

1327 The research was conceived by Joshua Bauld, David Lehmann, Katharine Abernethy, Luc Bussière,
1328 Jason Newton and Isabel Jones. Data collection was conducted by Katharine Abernethy and the SEGC
1329 field scientists. Joshua Bauld carried out data analysis, with guidance from Luc Bussiere and David
1330 Lehmann. Joshua Bauld wrote the chapter with guidance from David Lehmann, Katharine Abernethy,
1331 Jason Newton, Luc Bussière and Isabel Jones.

1332 **3.1 Abstract**

1333 Understanding primate dietary plasticity may give insights into trait evolution and resilience
1334 to environmental change. Here we investigate the feeding ecology of mandrills (*Mandrillus*
1335 *sphinx*), a species that forms groups of close to 1000 individuals, which presumably impacts
1336 feeding ecology by creating exceptionally high feeding competition. Mandrills are also
1337 threatened by habitat loss and climate change and a full understanding of their dietary
1338 plasticity is essential to ongoing conservation efforts. Evidence suggests that mandrills are
1339 highly generalist feeders, able to consume a wide range of resources to make up shortfalls in
1340 fruit availability. However, a lack of long-term data on fruit production within the mandrill
1341 geographic range means that it is unknown whether the flexible feeding strategies observed
1342 previously are stable over multiple years. We combined two rare datasets comprising eight
1343 years of fecal collections and fruit availability observations to assess the dietary flexibility of
1344 mandrills in response to seasonal variations in fruit production in Lopé National Park, Gabon.
1345 We found fruit to be the most frequently consumed resource and that fruit consumption
1346 covaried positively with fruit availability, peaking during periods of fruit abundance. Mandrill
1347 dietary diversity increased during periods of fruit scarcity, through greater consumption of
1348 animal prey, leaves, seeds, and other plant fibers. These results demonstrate that mandrills
1349 are primarily frugivorous, but that they are also highly flexible feeders, able to respond to
1350 temporal variation in fruit production over several annual cycles. In addition, we found
1351 mandrills to vary in the extent to which they preferred different fruit taxa. Lipid-rich oil palm
1352 (*Elaeis guineensis*) fruits were by far the most frequently consumed resource and may

1353 constitute a keystone resource to mandrills in the study site. Our results may be used to
1354 initiate future study of mandrill nutritional ecology, as well as inform ongoing conservation
1355 efforts throughout mandrill range countries.

1356

1357 **3.2 Introduction**

1358 Diet has long been a focus of primatological research (Lambert and Rothman, 2015), because
1359 of its relevance to topics such as species coexistence (Houle *et al.*, 2006), space use (Hanya *et*
1360 *al.*, 2004; Zhang *et al.*, 2021), life histories (Borries *et al.*, 2011) and morphological trait
1361 evolution (Regan *et al.*, 2001; Onstein *et al.*, 2020). Furthermore, habitat destruction and
1362 climate change are intense conservation challenges for primates, many species of which are
1363 at high risk of extinction (Pacifici *et al.*, 2017; Bernard and Marshall, 2020). Investigating
1364 dietary flexibility could provide insights about primates' resilience to change and identify the
1365 most (and least) effective conservation strategies (Harcourt *et al.*, 2002; Nowak and Lee,
1366 2013). A greater understanding of wild primate diets may also inform the ongoing debate
1367 about optimal human nutrition (Milton, 2000; 2003; Alt *et al.*, 2022).

1368 Long term behavioral and habitat data are key to a full understanding of primate ecology and
1369 evolution (Chapman *et al.*, 2017; Melin *et al.*, 2020). In the case of diet, long-term data
1370 facilitate inferences about feeding ecology that account for inter-annual changes in variables
1371 such as food availability (Chapman *et al.*, 2002; Zhou *et al.*, 2009; Erhart *et al.*, 2018). In this
1372 study, we use rare multi-year datasets on diet and fruit production to investigate the feeding
1373 ecology of mandrills (*Mandrillus sphinx*), by analyzing seasonal changes in diet across eight
1374 annual cycles. We focus on mandrills because their large social groups (Abernethy *et al.*, 2002)
1375 and extreme sexual dimorphism (Setchell, 2016) may greatly impact dietary plasticity at the
1376 individual and group level.

1377

1378 *3.2.1 The Influence of Phenology*

1379 One of the major influences on primate diets is seasonal variation in food availability (van
1380 Schaik *et al.*, 1993; Peres, 1994; Tuyisingize *et al.*, 2022). The seasonal timing of biological

1381 events, such as fruit production, is termed phenology (Lieth, 1974), which in tropical trees is
1382 influenced by temperature and rainfall (Reich, 1995; Mendoza *et al.*, 2017; Potts *et al.*, 2020).
1383 Frugivorous primates rely on fruit as their principal food source and, as a result, typically
1384 exhibit seasonal dietary variation in response to phenological cycles of fruit production (Guo
1385 *et al.*, 2007; Chancellor *et al.*, 2012; Butt *et al.*, 2015; DeLuycker, 2021). The taxonomic
1386 diversity and abundance of fruit may, however, vary between years because the fruiting
1387 phenology of different tree species can follow annual, sub- or supra-annual cycles (Bush *et*
1388 *al.*, 2017; Adamescu *et al.*, 2018). As a consequence, our primary goal was to describe average
1389 seasonal changes in mandrill food selection across multiple years, that may have varied in
1390 exactly which fruit species were available.

1391 Fluctuations in fruit abundance result in periods of scarcity through the year, and during these
1392 times, frugivorous primates are expected to exhibit dietary and behavioral flexibility to make
1393 up any nutritional shortfalls (Tutin *et al.*, 1991). Possibilities include switching to other food
1394 types, such as leaves or invertebrates, feeding on more consistently available, but lower
1395 quality fruits (Hill, 1997; McConkey *et al.*, 2002; Clink *et al.*, 2017), foraging over larger
1396 areas (Nagy-Reis and Setz, 2017), or reducing group or party size, to exploit smaller food
1397 patches (Tutin and Fernandez, 1993). ‘Preferred’ food types may therefore be identified as
1398 those for which consumption covaries positively with availability (Leighton, 1993). In contrast,
1399 ‘fallback’ foods can be considered as alternative food types, the consumption of which
1400 covaries negatively with the consumption of preferred resources (Wrangham and Marshall,
1401 2007). This switching to alternative resources may furthermore be accompanied by an
1402 increase in dietary diversity, as primates attempt to make up the nutritional shortfalls of
1403 alternative food sources (Lambert and Rothman, 2015).

1404 In addition to assessing which fallback foods are consumed, knowledge of which fruits are
1405 selected during periods of abundance is also necessary for understanding frugivore feeding
1406 strategies (Leighton, 1993; Doran-Sheehy *et al.*, 2009). Fruits which are consumed in greater
1407 amounts as a function of availability may be considered ‘preferred’ and those consumed less
1408 as a function of availability may be classified as ‘avoided’ (Russo *et al.*, 2005). A robust
1409 understanding of how primate diets vary in response to phenology is an important first step
1410 to understanding their feeding strategies (Chapman *et al.*, 2002). For example, some foods

1411 may be targeted because of their macronutrient contents (Conklin-Brittain *et al.*, 1998) and
1412 some may be avoided due to containing plant defensive compounds (Masette *et al.*, 2015).

1413

1414 3.2.2 *The Influence of Resource Competition*

1415 Primate food selection may also be affected by within- or between-species resource
1416 competition. The socioecological model for primate group formation suggests that females
1417 form groups as an anti-predation strategy, but that this aggregation of animals in space
1418 produces feeding competition (Wrangham, 1980; van Schaik, 1989). The intensity of feeding
1419 competition is then in turn influenced by the distribution of resources and group size, and can
1420 occur both within and between groups of conspecifics (Sterck *et al.*, 1997). Scramble
1421 competition, in which individuals (or groups) compete to be first to arrive at clumped
1422 resources, may arise once individuals form groups and increase in strength as group size
1423 increases (Miller *et al.*, 2020). Feeding competition can alternatively take the form of contest
1424 competition, whereby dominant animals (or groups) exclude others from a resource (Scarry
1425 *et al.*, 2013). Within- and between-group feeding competition in turn means that even when
1426 resource availability is high, preferred resources may still be less accessible to some
1427 individuals (or groups), impacting their food selection. For example, high-ranking
1428 chimpanzees have been observed to monopolize the most nutritious fruits at the tops of
1429 trees, at the expense of subordinate individuals (Houle and Wrangham, 2021).

1430 Feeding competition may also occur between different species, impacting the food selection
1431 of primates living in multi-species assemblages. For example, chimpanzees have been
1432 observed to exclude subordinate species from fruit trees (Houle *et al.*, 2010). Species may
1433 also avoid competition by segregating habitats in time or space. For example, pitheciine
1434 monkeys are thought to have evolved to feed on immature fruit, to access seeds earlier than
1435 other species (Ledogar *et al.*, 2013). Similarly, primate assemblages may split habitats by
1436 height (Sushma and Singh, 2006) or feed in separate habitats during periods of resource
1437 scarcity (Wahungu, 1998). The result of between-species competition is that sympatric
1438 frugivorous may exhibit dietary distinction even during periods of fruit abundance and
1439 consume different fallback foods during periods of scarcity. Consideration must therefore be
1440 given to the fact that frugivorous primates usually live in groups and alongside other

1441 frugivorous species, leading to competition for high value resources, when investigating how
1442 food selection varies in response to phenology.

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1444 3.2.3 Mandrill Ecology

1445 Mandrills are a particularly interesting species in which to study the relationship between tree
1446 phenology and diet, because they exhibit several exceptional traits relevant to food selection.
1447 Though classified as frugivores, mandrills consume a remarkably wide range of foods,
1448 including vertebrate and freshwater prey (Jouventin, 1975; Hoshino, 1985; Harrison, 1988;
1449 Norris, 1988; Rogers *et al.*, 1996). Mandrills also live in the largest known social groups,
1450 termed 'hordes', of any non-human primate (620+/-166, range 340-845; Abernethy *et al.*,
1451 2002). Such large group sizes likely result in high levels of within-group feeding competition
1452 and rapid rates of patch depletion. Consequently, mandrills also occupy the largest total
1453 home-ranges documented in any wild primate (182 km²) and travel costs are likely
1454 exceptionally high for individuals living in mandrill hordes (White *et al.*, 2010). These
1455 socioecological traits are coupled with the most extreme size dimorphism seen in primates
1456 and extravagant facial adornments on male animals (Darwin, 1876; Setchell, 2016).
1457 Furthermore, the geographic range of mandrills overlaps with those of numerous other
1458 primate species and high biomass frugivores, such as forest elephants (*Loxodonta cyclotis*)
1459 and red river hogs (*Potamochoerus porcus*) (Tutin *et al.*, 1997). Individual mandrills must
1460 therefore cope with high levels of feeding competition, to consume a diet that provides
1461 sufficient energy to sustain extensive travel, with nutritional requirements also potentially
1462 differing between sexes, due to extreme dimorphism.

1463 Periods of fruit scarcity may hamper the ability of individual mandrills to consume sufficient
1464 resources. Previous investigations have documented that when fruit availability is low,
1465 mandrill feed on a variety of fallback foods, with a consequent increase in dietary diversity
1466 (Nsi Akoué *et al.*, 2017; Hongo *et al.*, 2018). Tree phenological cycles and fruit availability may
1467 vary between years, however, and so whether this feeding strategy remains consistent over
1468 several annual cycles is an open question, unanswered primarily due to a lack of long-term
1469 data on either tree phenology or mandrill diets (White, 2007; Hongo *et al.*, 2018).

1470 In this study, we investigate mandrill feeding strategies using a rare eight year dataset on
1471 mandrill diets, obtained from fecal samples gathered over an eight-year period in Lopé
1472 National Park, Gabon. We analyze these data in conjunction with a long-term (1986-present)
1473 tree phenology dataset, one of only 12 available in Africa (Adamescu *et al.*, 2018) and the only
1474 one within the known geographic range of mandrills. The Lopé phenology dataset indicates
1475 that fruit abundance varies seasonally, with fruit production peaking in the two wet seasons
1476 (Feb-May and September-November) and falling in the two dry seasons (June-August and
1477 December-January) (Bush *et al.*, 2017). This fluctuation in fruit availability has previously been
1478 observed to influence the diets of primates living within the park (White *et al.*, 1994; Tutin *et*
1479 *al.*, 1997). However, the exact timing and duration of each season, as well as the total amount
1480 of fruit produced, may vary between years (Tutin *et al.*, 1991). By combining the fecal and
1481 phenology datasets we are therefore able to build upon prior investigations with an analysis
1482 of mandrill feeding strategies that account for long-term variation in tree phenology and fruit
1483 availability.

1484 To investigate mandrill diets we analyzed 4024 fecal samples, collected between September
1485 1996 and October 2004, from which we recorded the major food types consumed and
1486 identified plant tissues to species level, where possible. Using these data, we first set out to
1487 describe our study horde's diet in terms of preferred food types, use of fallback foods and
1488 seasonal changes in food type diversity. Mandrills also appear to prefer some fruits over
1489 others during periods of abundance (White, 2007, Nsi Akoué *et al.*, 2017). We therefore also
1490 examined whether particular fruits were consumed more frequently as a function of
1491 availability, suggesting they are preferred foods, and explored why preferences may exist in
1492 terms of maximum tree height and nutritional contents. We were interested in the impact of
1493 maximum tree height on the frequency at which fruit genera were consumed, because a
1494 negative association could suggest that mandrills avoid foraging in tall trees or are excluded
1495 from the tops of trees by other frugivorous species. Similarly, we wanted to examine the
1496 relationship between nutritional contents and consumption frequency to try and elucidate
1497 the mechanisms driving mandrill food selection.

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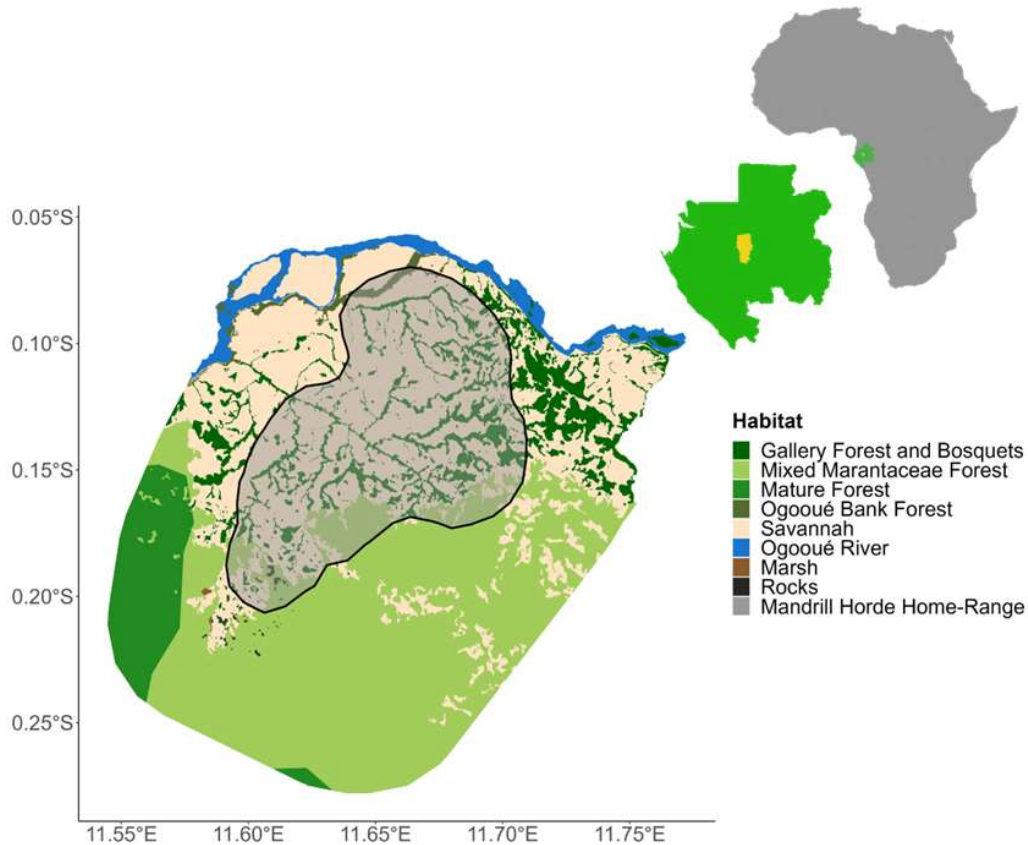
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1500 **3.3 Methods**

1501 *3.3.1 Study site*

1502 Our study was carried out in Lopé National Park, located in Gabon (-0.2N, 11.6E), Central
1503 Africa (Figure 3.1). Lopé National Park covers an area of 4964 Km² which is mostly old growth
1504 Guineo-Congolian evergreen tropical rainforest, but contains a dynamic forest-savanna
1505 mosaic landscape covering approximately 10 % of its northern area (Figure 3.1). This mix of
1506 savanna and gallery forests is bordered to the north by the Ogooué (the second largest river
1507 in the Congo-Ogooué basin). Our 182 km² study area, based on the area used by mandrills in
1508 Abernethy *et al.* (2002), is situated within the savanna forest matrix and adjacent continuous
1509 forest (Figure 3.1). The dominant type of vegetation at the continuous forest edge in this area
1510 is ‘Marantaceae forest’, which is a young secondary forest with a dense herbaceous
1511 understory, dominated by plants of the families Marantaceae and Zingiberaceae. Gallery
1512 forest fragments that extend out from the main continuous forest-block into the savanna
1513 typically form along small riverbeds and have sparse ground vegetation of lianas and small
1514 shrubs. Plant species assemblages and compositions differ between gallery forests, forest
1515 fragments and continuous forest (see White, 1994; White and Abernethy, 1997, Tutin *et al.*,
1516 1997, Léal, 2004; Ukizintambara *et al.*, 2007; White, 2007). Notably, the gallery forests have
1517 a history of anthropization over several hundred years, that included the planting of oil palms
1518 (*Elaeis guineensis*) (Maley and Chepstow-Lusty, 2001; Bostoen *et al.*, 2013). Lopé National
1519 Park receives on average 1466+/- 201mm of rain per year; which falls within two distinct time-
1520 windows; the long rainy season runs from February to May, and the shorter rainy season from
1521 September to November. The short dry season takes place from December to January while
1522 the longer dry season occurs from June to August (White 1994, Bush *et al.* 2020).

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1532 *Figure 3.1. The location of Lopé National Park in Gabon, Central Africa, and the distribution of*
1533 *habitats in the study site, with the home-range (90% kernel density isopleth) of the focal horde*
1534 *overlaid (data from White et al., 2010).*

1535
1536

3.3.2 Study Population

1537 Our study population consisted of a mandrill horde, usually numbering between 600-800
1538 individuals, though temporary extreme counts of as many as 1350 individuals have been
1539 recorded in the past, when two hordes meet (Abernethy *et al.*, 1997). The group exhibits
1540 fission-fusion dynamics during periods of low food availability (White, 2007), meaning that
1541 the number of individuals varied considerably throughout our eight-year study period, with a
1542 recorded range of 340-845 for subgroups (Abernethy *et al.*, 2002). Consequently, the level of
1543 within-group feeding competition may have varied throughout the sampling period,
1544 potentially impacting food selection. In addition, adult males often leave the horde outside
1545 of the breeding season, but even when present, the proportion of adult and sub-adult males
1546 has never been observed to exceed 12% (Abernethy *et al.*, 2002) and so our fecal samples

1547 were mostly sourced from adult females and juveniles of both sexes. Our analysis of mandrill
1548 diets thus largely corresponds to permanent horde members and fewer adult males.

1549

1550 3.3.3 Fecal Collection and Analysis

1551 To investigate seasonal changes in mandrill food selection, we collected 4024 fecal samples
1552 over 8 years and 2 months, between September 1996 and October 2004. For all feces
1553 collections, the mandrill group being sampled was identified as being all or part of the two
1554 hordes known to use the study area, using radiotelemetry to identify the presence or absence
1555 of collared individuals (see Abernethy *et al.*, 2002; White, 2007). Only fresh dung (10 minutes
1556 to 5 hours old) was collected, to minimize loss of material to fecal predators. When a dung
1557 pile was found, the entire fecal pellet was collected into a plastic ziplock bag and conserved
1558 at ambient temperature until analysis. Collections were made weekly on a predetermined day
1559 to prevent bias to collections in more accessible habitats and to ensure an even
1560 representation throughout the year. When logistical constraints prevented collections being
1561 made on the determined day of the week, efforts were continued to contact the group and
1562 samples were obtained as soon as possible after this date.

1563 Identification of dietary items was made within 3 days of collection to ensure that constituent
1564 parts were recognizable and had not been degraded. Analysis of constituent parts followed
1565 the protocol established at SEGC by Tutin and Fernandez (1993) for analysis of ape dung. We
1566 used 1 mm mesh brass soil sieves to separate the matrix with running water and retain
1567 identifiable undigested components. Cleaned dung components were identified to plant part
1568 and species as far as possible, using the herbarium and reference collections of seeds curated
1569 at SEGC (White and Abernethy, 1997). Where species level identification was not possible,
1570 plant genus, plant family or finally 'component type', for example 'unknown stems' or 'plant
1571 fiber', classifications were made. Animal and insect remains could rarely be classified beyond
1572 Order level (e.g., mammal, reptile, insect, mollusk), however where possible a more precise
1573 taxon was noted.

1574

1575 3.3.4 Descriptive Analyses of Food Types and Plant Taxa

1576 To analyze seasonal changes in consumption of major food types, we classified identified
1577 foods into one of seven categories. The whole seeds, pulp, skins and fibers of identified fruit
1578 species were pooled as 'Fruit'; leaves from monocotyledonous plant species were clustered
1579 as 'Monocot Leaves'; dicotyledonous leaves were categorized as 'Dicot Leaves'; chitin from
1580 insects, hair, bone and other vertebrate or invertebrate remains were grouped as 'Animal
1581 Parts'; pieces of mushroom were categorized as 'Fungus'; seeds that showed clear evidence
1582 of seed predation (rather than fruit consumption and seed dispersal) were marked as
1583 'Crushed Seeds'; and unidentified fibers or plant material, pieces of bark, twigs and pieces of
1584 wood were termed 'Other Fiber'.

1585 Alongside seasonal changes in consumption of each food type, we also aimed to analyze
1586 seasonal changes in diet breadth. Here, we define 'diet breadth' as the count of different non-
1587 fruit categories, plus the count of distinct fruit genera present in a fecal sample. We then
1588 define 'fruit breadth' as the count of distinct fruit genera present in a fecal sample. We can
1589 therefore compare changes in diet breadth and fruit breadth throughout the year, to describe
1590 temporal variation in the contribution of fallback (i.e. non-fruit) foods to mandrill diet
1591 breadth. To classify different plant taxa within the fruit breadth variable, we grouped different
1592 tree species to the Genus or Family level, because species level identifications were not
1593 always possible. For example, for *Uapaca* sp., there are three species, *U. heudeleotti*, *U.*
1594 *paludosa* and *U. guineensis* present in the study area and their seeds are difficult to reliably
1595 tell apart, especially when damaged. This grouping resulted in a loss of some dietary
1596 resolution, but allowed us to maximize the sample size for characterizing the phenology of
1597 food items in mandrill dung.

1598 Throughout our analyses of seasonal variation in food selection and diet breadth we focus on
1599 the presence versus absence of food items, rather than quantifying the amount of each in
1600 fecal samples. This is partly because comparisons of quantities across items is difficult (e.g.,
1601 for fruits with dramatically different seed sizes), but also because fecal pellets represent only
1602 the undigested fraction of food and could therefore easily misrepresent food selection or
1603 underestimate unidentifiable remains (Tutin and Fernandez, 1993).

1604 3.3.5 Fruit Availability

1605 To facilitate comparisons between fruit consumption and fruit availability, and thus identify
1606 preferred or avoided fruit taxa, we calculated fruit availability using data on tree fruiting
1607 phenology, tree abundance and tree size. Tree phenology has been monitored in Lopé
1608 National Park since 1986 to present. Field researchers use binoculars to record canopy cover
1609 of immature and mature fruits, rating coverage on a 9 nine point scale from 0, for no coverage,
1610 to 4, for complete coverage (Tutin *et al.*, 2017; Bush *et al.*, 2018). Stem density (stems ha⁻¹)
1611 and mean diameter at breast height (cm) (converted to radius) measures were taken from
1612 existing census data for each of the four major forest types in the study area (White 1994;
1613 Cardoso *et al.* 2020). Phenology data were available for 30 species, in 23 of the 53 genera
1614 found in mandrill feces and we obtained a subset of the data available for these species; that
1615 matched the date range of the fecal sampling. We then calculated monthly fruit availability
1616 scores (FAS) for each species, following the method of Cardoso *et al.* (2020):

$$1617 \quad \text{FAS} = p \times c \times d \times r^2$$

1618 where p is the proportion of a species bearing fruit in any given month; c is the mean
1619 proportion of the canopy of each species covered by mature fruit; d is the mean stem density
1620 of a species across all census plots in a habitat type; and r is the mean radius at breast height
1621 of a species across all census plots in a habitat type. Monthly FAS were calculated for each
1622 species in each habitat type and then multiplied by the total area of each habitat within our
1623 study area. FAS for each species, in each habitat were then summed, to produce a measure
1624 of the total monthly availability of each fruit species, across our study site. Thus, we produced
1625 measures of fruit availability that could be compared directly to the frequency at which
1626 fruiting taxa appeared in mandrill fecal samples, which were also collected on a monthly
1627 basis. In cases where FAS were calculated for multiple species in the same genus, we summed
1628 the FAS for all species in a genus, so that our measure of fruit availability matched the
1629 taxonomic resolution of our fecal dataset.

1630

1631 3.3.6 Statistical Analyses

1632 Fecal sampling events occurred across all four seasons (long wet, long dry, short wet, short
1633 dry), but to raise temporal resolution we focused instead on the day of year on which each
1634 sample was collected (e.g. January 1st = Day 1). This allowed us to estimate changes in
1635 prevalence at finer temporal scales than those available when considering seasons as discrete
1636 periods. To quantify how the probability of occurrence of major food types varied through
1637 time we fitted a generalized additive mixed model (gamm) with a logit-link function and
1638 binomial error structure using the 'mgcv' package v1.3.89 (Wood, 2017) in R v4.1.3 (R Core
1639 Team, 2022). Gamm models make no assumptions about the functional form of a curve, but
1640 rather allow us to estimate the curvature best supported by the data. We included a
1641 smoothing term for day of year as a fixed factor and food type as a categorical fixed factor, to
1642 describe the effect of day of year on binomial presence of each food type in the diet; without
1643 fitting a global effect for day of year across food types. We also allowed the curve for each
1644 food type to have its own wiggleness, meaning the curve shape could vary between food
1645 types, to accurately describe temporal patterns of dietary presence for individual food types
1646 (equivalent to Model I in Pedersen *et al.*, (2019)). Day of year was specified as a cyclic cubic
1647 regression term to ensure that the intercepts for early January and late December aligned.
1648 We also included random intercepts for year as an additional smoother, so that the curves
1649 describing the frequency at which each food type appeared in mandrill diets accounted for
1650 differences between years in mandrill feeding. Our model structure therefore tends to
1651 penalize differences between sampling years, in proportion to the difference between each
1652 year and all other years. Smoothing parameter estimation was conducted using restricted
1653 maximum likelihood (REML).

1654 To analyze whether mandrills compensate for low fruit availability by consuming alternate
1655 resources, we modeled temporal changes in diet breadth (count of distinct food types and
1656 fruit genera in a fecal sample) and fruit breadth (count of different fruit genera in a fecal
1657 sample) using Poisson gamms, with a log-link function, including day of year as a fixed effect
1658 and random intercepts for year. To test whether consumption of fallback foods was
1659 statistically associated with fruit consumption, we compared the proportion of fecal samples
1660 containing non-fruit food types when fruit was present or absent, using chi-squared tests.

1661 To assess evidence of preference by mandrills for certain fruit genera, we used a binomial
1662 generalized linear mixed model, with consumption of fruit as the dependent variable and
1663 scaled fruit availability (without mean centering) as the independent variable. A positive
1664 relationship between fruit availability and consumption would suggest that mandrills
1665 consume more fruit when it is most available, whereas no association between fruit
1666 consumption and availability would imply that fruit is not a preferred resource of mandrills.
1667 Similarly, relative differences in the strength of relationships between availability and
1668 consumption of different fruit genera would also imply differences in the extent to which they
1669 are preferred or avoided by mandrills. Therefore, we also included a random slopes argument
1670 in this model, allowing the relationship between availability and consumption to differ
1671 between fruiting genera, interpreting steeper positive slopes as an indication of greater
1672 preference by mandrills. The model containing random slopes for each genus was compared
1673 to one containing random intercepts and a fixed effects only model using AIC and the model
1674 with the lowest AIC taken as the preferred model (Burnham *et al.*, 2011).

1675 We next wanted to examine whether maximum tree height and fruit nutritional contents
1676 (protein, carbohydrates, fat, water, fiber, tannins, or phenols) were associated with the
1677 frequency at which different fruits were consumed. To do so, we ran separate generalized
1678 linear mixed models, with tree height or each nutritional trait as a continuous independent
1679 variable, random intercepts for each fruit genus and the binomial presence of each genus in
1680 mandrill feces as the dependent variable. We took maximum tree height data from the
1681 primary vegetation guide for Lopé National Park (White and Abernethy, 1997) or online
1682 databases of plant traits and nutritional data from Rogers *et al.* (1990).

1683 Finally, we aimed to test whether the seed sizes of fruit influenced the rate at which they
1684 appeared in mandrill feces. We categorized seed sizes as “very small”, “small”, “medium”,
1685 “large” and “very large” based on scaled illustrations in White and Abernethy (1997) and
1686 analyzed differences in the frequency at which fruit with different seed sizes were consumed,
1687 using a generalized linear mixed model. Seed size was included as a categorical independent
1688 variable, fruit genus as a random effect and the binomial frequency at which fruit genera
1689 appeared in feces as the dependent variable.

1690 For all models, we visually inspected diagnostic plots to ensure good model fit and adherence
1691 to model assumptions. We also explored the sensitivity of our gamms to the number of
1692 smoother knots, k , using the *gam.check* function. For a few models, the default number of
1693 smoothers suggested the possibility of underfitting. However, increasing k did not reveal
1694 patterns that altered our interpretation and so, for simplicity and consistency, we present
1695 fitted values for curves from models with $k = 10$ for fixed effects and $k = 9$ for the random
1696 effect of year.

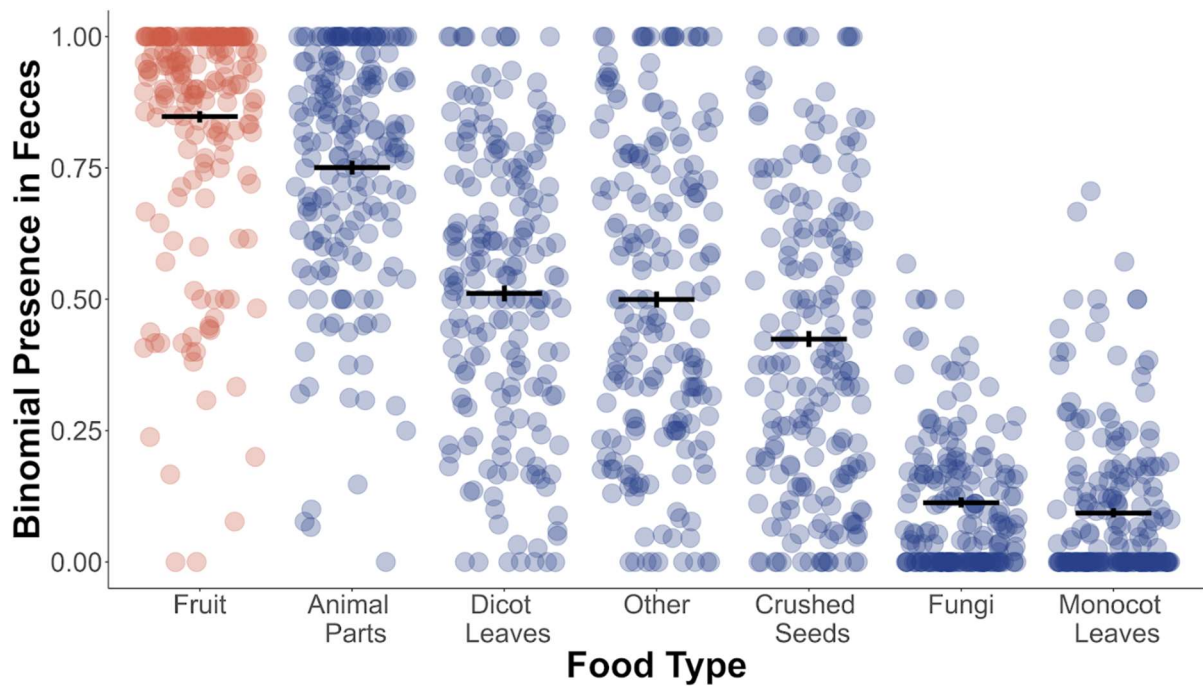
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1698 **3.4 Results**

1699 *3.4.1 Major food types consumed*

1700 We grouped the foods that mandrills consumed into 7 categories: fruit, animal parts
1701 (including vertebrates and invertebrates), dicotyledonous leaves, monocotyledonous leaves,
1702 crushed seeds, fungi, other fibers (including non-fruit fiber and wood). The mandrills
1703 consumed a minimum of 67 different plant species, which is a conservative estimate, as not
1704 all samples could be identified to species level and many of the families and genera identified
1705 have multiple representatives within the horde's home-range (Table A3.1).

1706 Fruit was the most frequently consumed food type, based on binomial presence versus
1707 absence, present in 84.8% (95% CI: 83.7-85.9) of fecal samples. The second most frequently
1708 consumed food type was animal parts (75.1%, 73.7-76.4), followed by dicotyledonous leaves
1709 (51.1%, 49.6-52.7), other fibers (50%, 48.4-51.5), crushed seeds (42.4%, 40.9-44), fungi
1710 (11.3%, 10.3-12.3) and monocotyledonous leaves (9.3%, 8.5-10.2) (Figure 2.2).



1711 *Figure 3.2. Consumption of major food types by the focal horde, across all fecal samples (n =*
 1712 *4024). Data points represent an individual day of the year (n = 192) on which feces were*
 1713 *sampled. The position of points on the y axis indicates the proportion of individuals to have*
 1714 *consumed each resource on a given day of the year, averaged across all sampling years, and*
 1715 *based on binomial presence versus absence. Horizontal lines indicate the proportion of all fecal*
 1716 *samples containing a given food type and vertical black lines indicate the 95% binomial*
 1717 *confidence intervals around those proportions.*

1718 3.4.2 Seasonal patterns of food type consumption

1719 A generalized additive model indicated that consumption frequency of all major mandrill food
 1720 types was associated with Day of Year (Figure 3.3; Table A3.2). Fruit consumption peaked
 1721 during the two wet seasons (Figure 3.3: Fr), whereas some alternative food types showed
 1722 inverse trends compared to fruit, with crushed seeds, dicot leaves and other fibers all
 1723 exhibiting two peaks in the dry seasons, when fruit consumption was lowest (Figure 3.3: DL,
 1724 O, CS). Consumption of animal parts appeared somewhat independent of fruit consumption.
 1725 Peaks in animal consumption were present at the end of the long wet season, when fruit
 1726 consumption began to decline, but also at the beginning of the long wet season, when fruit
 1727 consumption was increasing, as well as during the short wet season, when fruit consumption

1728 was relatively high (Figure 3.3: AP). Consumption of monocot leaves and fungi had two small
1729 peaks in the wet seasons (Figure 3.3: Fu, ML).

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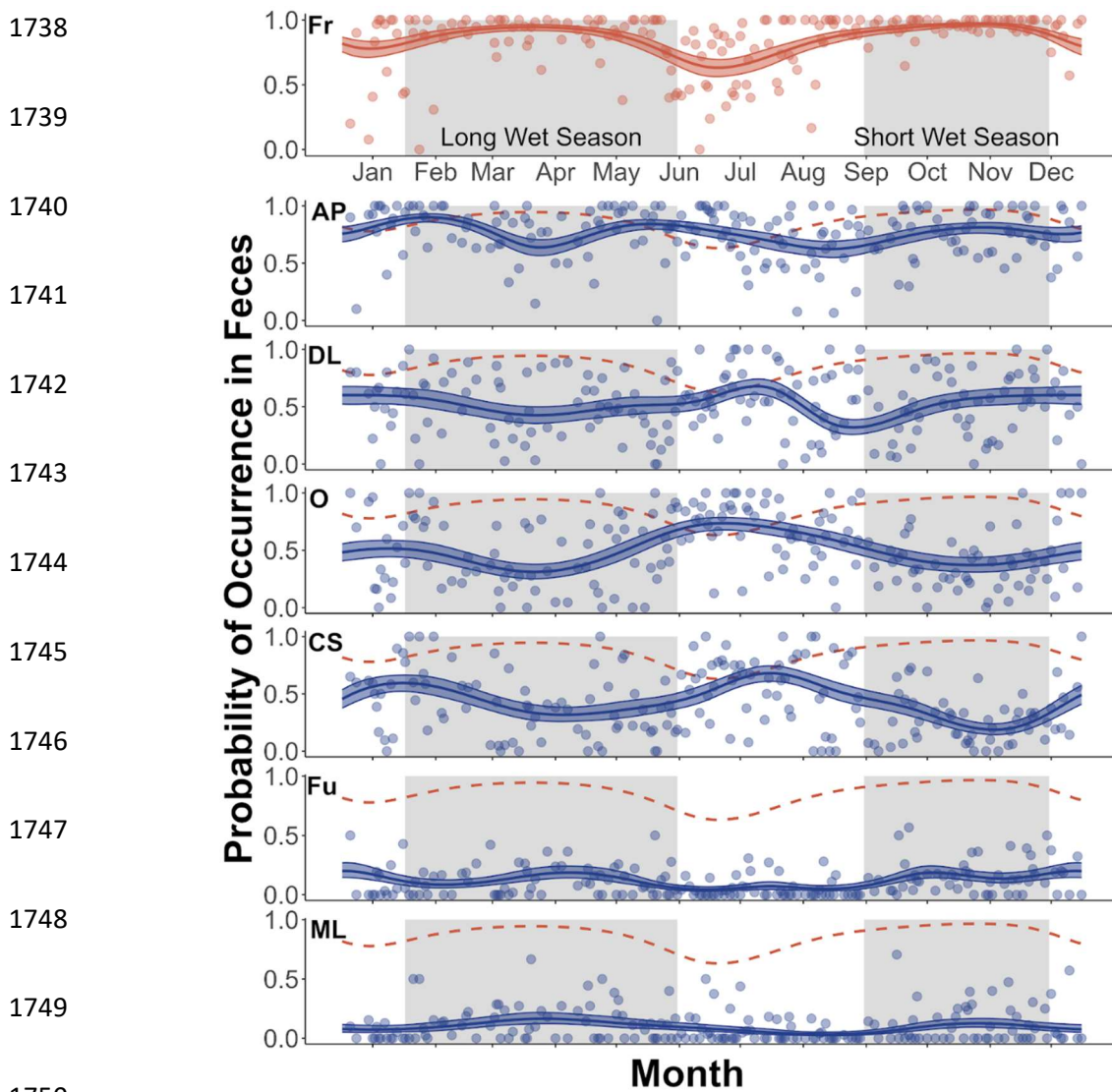
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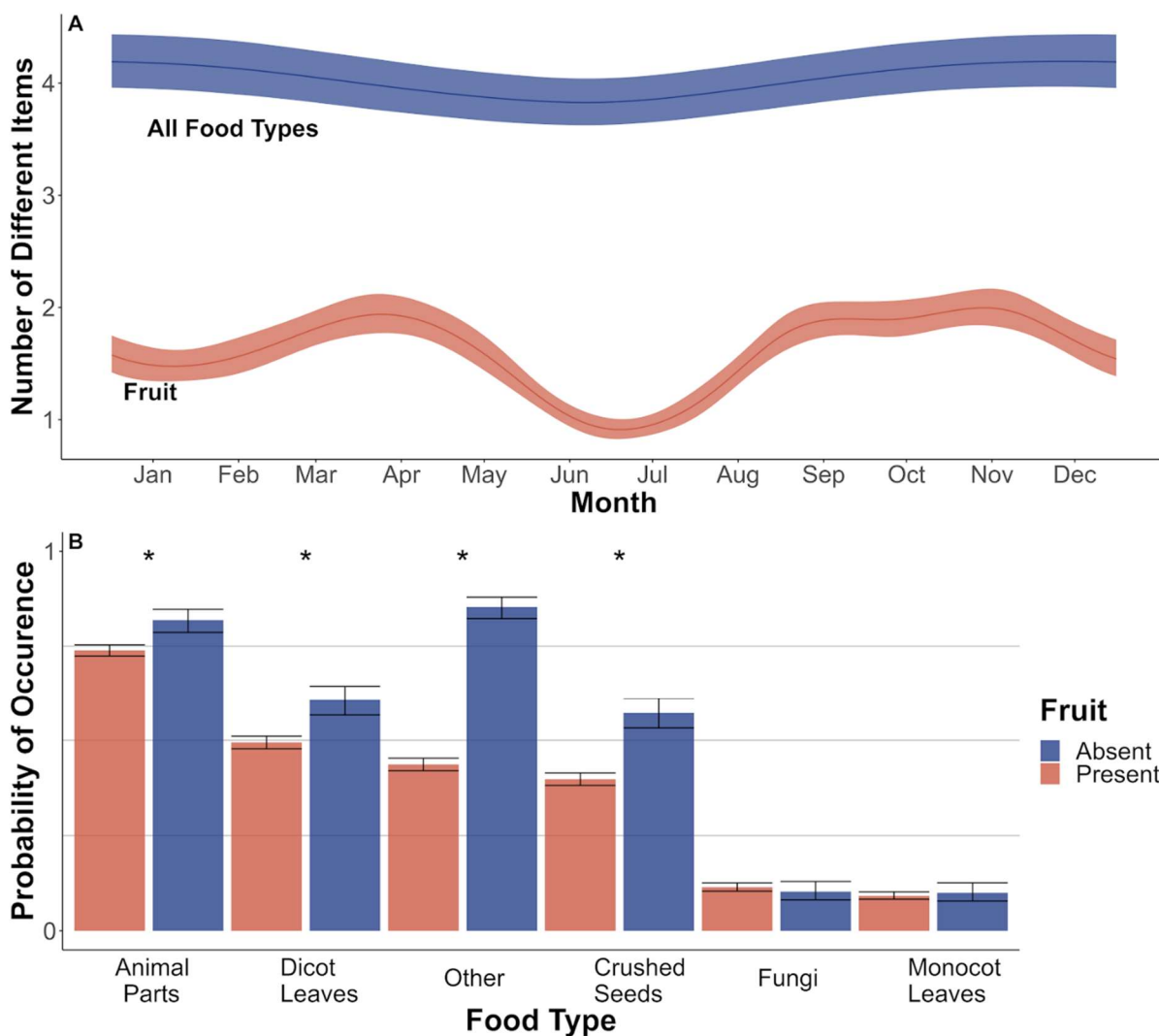


1750

1751 *Figure 3.3. Generalized additive model examining the effect of Day of Year on consumption of*
 1752 *major mandrill food types (Fr = Fruit, DL = Dicot Leaves, Fu = Fungi, O = Other, ML = Monocot*
 1753 *Leaves, CS = Crushed Seeds, AP = Animal Parts). Solid lines indicate the predicted binomial*
 1754 *probability of a given food type being found in a fecal sample, on a given day of the year.*
 1755 *Ribbons indicate the 95% confidence interval around the predicted binomial probability. The*
 1756 *model output is displayed over raw data to visualize the relationship between consumption*
 1757 *and day of year. Each data point represents one sampling day (n = 192), and their position on*
 1758 *the y axis indicates the proportion of fecal samples (total n = 4024) on that day containing a*
 1759 *given food type, averaged across all sampling years, and based on binomial presence versus*
 1760 *absence. Dark and light shaded areas are indicative of wet and dry seasons, respectively.*
 1761 *Dashed red lines illustrate the predicted binomial probability of a fecal sample containing fruit,*
 1762 *for comparison to other food types.*

1763 To measure diet breadth, we summed the number of items in a fecal sample belonging to
1764 different fruiting genera and to the alternative food type categories used above (e.g. 1x
1765 *Dialium sp.*, 1x *Uapaca sp.*, 1x dicot leaf, 1x animal part, equals a diet breadth of four).
1766 Generalized additive models revealed that the overall diet breadth of mandrills remains
1767 consistent throughout the year, with any one fecal sample containing four different items, on
1768 average (Figure 3.4A; Table A3.3). To measure fruit diet breadth, we summed the number of
1769 different fruiting genera present in a fecal sample (e.g. 1x *Dialium sp.*, 1x *Uapaca sp.* equals a
1770 fruit diet breadth of two). In contrast to diet breadth, fruit diet breadth was observed to
1771 fluctuate, peaking at roughly two genera of fruit in an average fecal sample in the two wet
1772 seasons, and falling in the dry seasons; especially the long dry season during which an average
1773 fecal sample was expected to contain only about one distinct fruit genus (Figure 3.4A; Table
1774 A3.4). Consumption of animal parts, crushed seeds, dicot leaves and other fibers was
1775 negatively associated with the consumption of fruit. In contrast, there was no evidence that
1776 consumption of monocot leaves and fungi was associated with fruit consumption (Table 3.1,
1777 Figure 3.4B).

1778



1779

1780 *Figure 3.4A. Generalized additive models illustrating the relationship between Day of Year and*
1781 *overall dietary breadth (number of different fruit genera plus different items from other food*
1782 *types) and dietary fruit breadth (number of different fruit genera) across 4024 fecal samples.*
1783 *Lines indicate the predicted number of different food or fruit types being found in a fecal*
1784 *sample, on a given day of the year (Day 1 = January 1st). Ribbons indicate the 95% confidence*
1785 *interval around the predicted numbers of different items. B. The binomial presence of major*
1786 *food groups in fecal samples also containing fruit and those in which fruit was absent. Bar*
1787 *height indicates the proportion of samples containing a food type, and error bars show the*
1788 *95% confidence intervals around those proportions (fruit present n = 3412, fruit absent n =*
1789 *612). Asterisks indicate statistically significant differences in the probability of food types*
1790 *occurring in the diet when fruit is present or absent, as indicated by chi-squared tests.*

1791

1792 *Table 3.1. Results of chi-squared tests examining differences in binomial presence of mandrill*
 1793 *food types in fecal samples, when fruit is present or absent in the sample.*

Type	Fruit	Proportion	95% CI Lower Bound	95% CI Upper Bound	<i>p</i>
Animal Parts	Present	0.74	0.72	0.75	<0.001
	Absent	0.82	0.79	0.85	
Dicot Leaves	Present	0.49	0.48	0.51	<0.001
	Absent	0.61	0.57	0.64	
Fungi	Present	0.12	0.1	0.13	0.41
	Absent	0.10	0.08	0.13	
Monocot Leaves	Present	0.09	0.08	0.1	0.6
	Absent	0.10	0.08	0.13	
Other	Present	0.44	0.42	0.45	<0.001
	Absent	0.85	0.82	0.88	
Crushed Seeds	Present	0.40	0.38	0.41	<0.001
	Absent	0.57	0.53	0.61	

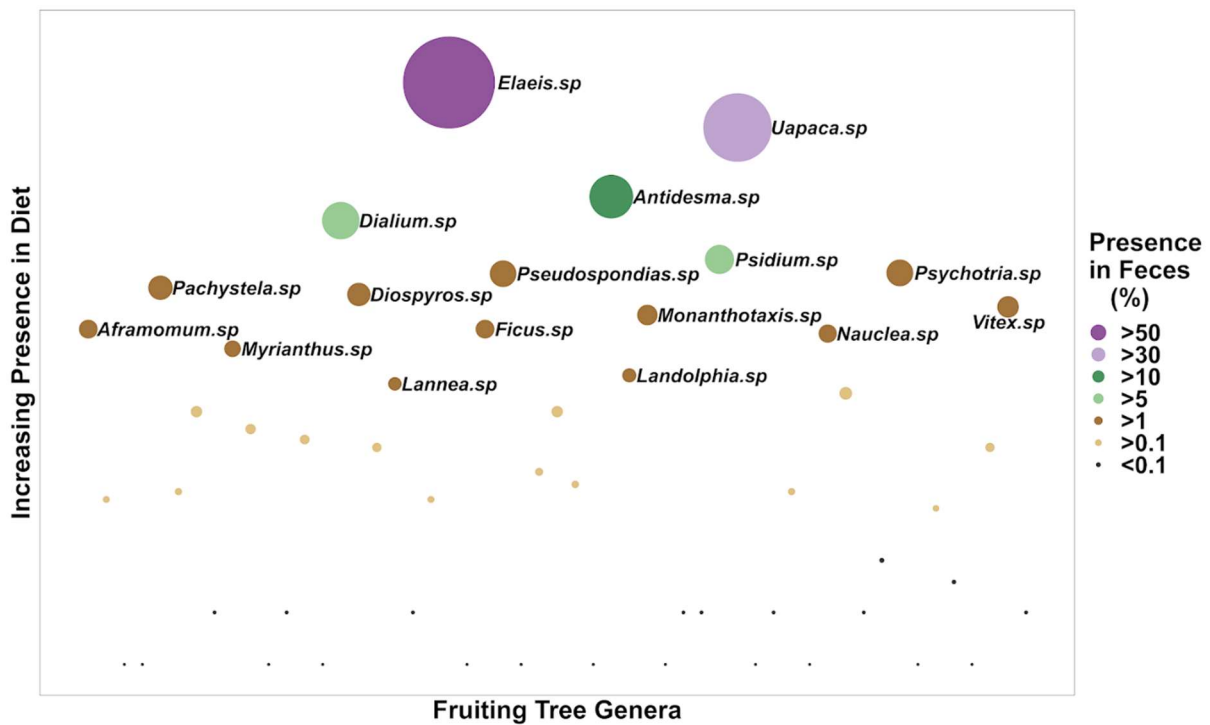
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1796 3.4.3 Fruit preference

1797 We found fruit genera to appear in the diet at various frequencies (Figure 3.5). A generalized
 1798 linear mixed model revealed a positive association between fruit availability and fruit
 1799 consumption by mandrills, across all genera tested (Table 3.2). Based on AIC scores, a model
 1800 containing random slopes, that allowed the relationship between availability and
 1801 consumption to vary between fruiting genera, was preferred (AIC = 16375) over models
 1802 containing only random intercepts (AIC = 17360) or only fixed effects (AIC = 25127). The
 1803 random slopes for each genus are displayed in Figure 3.6, and indicate that despite the overall
 1804 positive association across all genera, there was substantial variation among taxa in the extent
 1805 to which mandrill consumption covaries with availability. Furthermore, the most favored
 1806 genera were not always the most frequently consumed. For example, *Vitex sp.* were the most
 1807 preferred genera, indicated by the steepest positive slope between availability and
 1808 consumption, but were present in less than 5% of fecal samples. On the other hand, *Uapaca*
 1809 *sp.* were present in over 30% of fecal samples, but only the sixth most preferred of the genera
 1810 analyzed. In another comparison, *Detarium sp.* exhibited the weakest relationship between

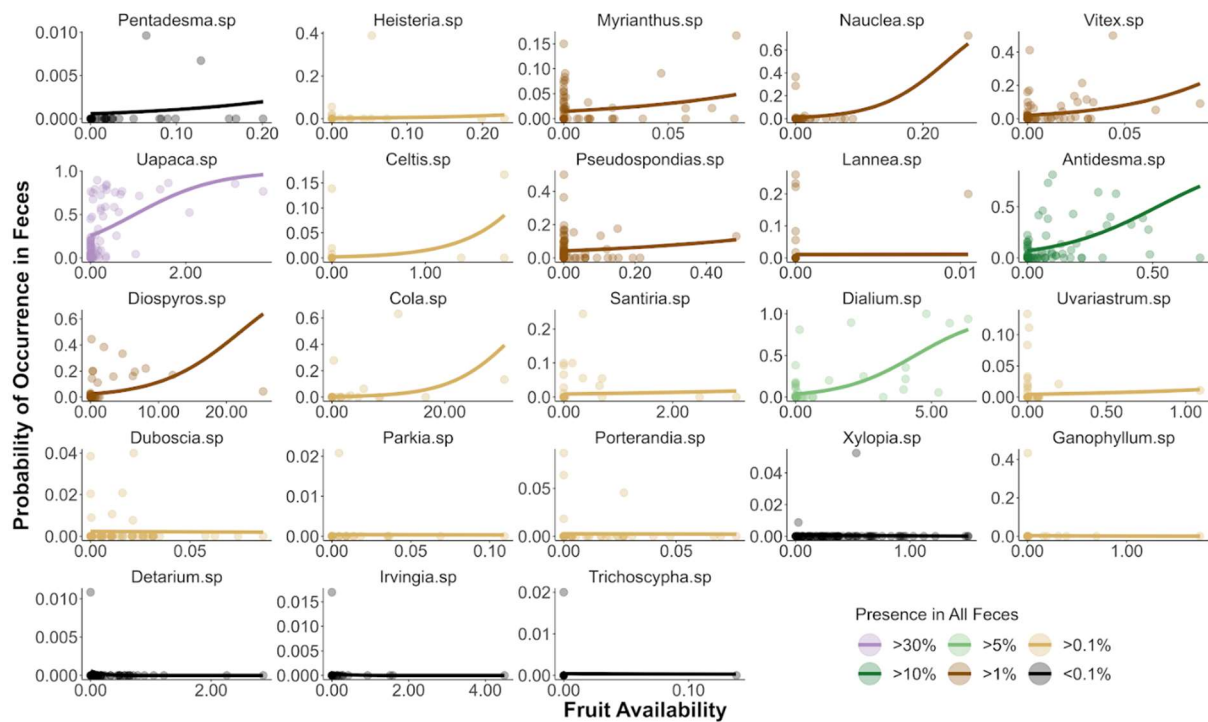
1811 availability and consumption frequency and *Pentadesma sp.* the fifth strongest relationship,
 1812 though both genera were present in less than 0.1% of fecal samples.
 1813



1814
 1815 *Figure 3.5. Fruit genera consumed by mandrills, ranked by proportional presence in all 4024*
 1816 *fecal samples. Each point represents a single genus, sized according the proportion of fecal*
 1817 *samples in which a genus was present, dispersed arbitrarily on the x axis to avoid overlaps and*
 1818 *ordered on the y axis by proportional occurrence in feces (log transformed).*

1819
 1820 *Table 3.2. Summary of generalized linear mixed model examining the relationship between*
 1821 *fruit availability and consumption by mandrills, with random slopes for each tree genus.*

	Estimate	Standard Error	z	p
Fixed Effects				
Intercept	-5.38	0.37	-14.53	<0.001
Scaled Fruit Availability Score	2.36	1.17	2.018	0.0436
Random Effects				
Genus	Random Intercept Variance	3.69		
Genus	Scaled Fruit Availability Score Variance	93.12		



1824 *Figure 3.6. The relationship between fruit availability and consumption of fruiting genera by*
 1825 *mandrills. Random slope predictions from a generalized linear mixed model are shown over*
 1826 *the observed fruit availability range for each genus. Points represent distinct sampling months*
 1827 *(n = 83), and their position on the x and y axes indicate the scaled fruit availability and*
 1828 *proportion of fecal samples (total n = 4024) containing a fruit genus during that month,*
 1829 *respectively. Scales differ between facets to avoid compressing data for genera with lower*
 1830 *availability. Intercepts of some species are > 0 because these fruits were consumed on the*
 1831 *ground after the ripe fruit our phenology monitoring detects in canopies were no longer*
 1832 *available. Effect sizes increase row-wise from left to right, with Vitex.sp exhibiting the largest*
 1833 *and Detarium.sp the smallest.*

1834

1835 We used generalized linear mixed models to examine whether maximum tree height, fruit
 1836 nutritional contents or seed size influenced mandrill fruit consumption. The association
 1837 between tree height and fruit consumption was statistically significant and negative [-0.048,
 1838 95%CI: -0.086 - -0.011, p = 0.011] (Figure 3.7), while the association between fruit lipid
 1839 contents and consumption was statistically significant and positive [0.072, 95%CI: 0.015 -

1840 0.13, $p = 0.014$] (Figure 3.8). However, the association between lipid content and
 1841 consumption appears to rely on a single high influence species of fruit: removing *Elaeis sp.*
 1842 (oil palm) from the data set resulted in no statistically significant relationship between lipid
 1843 content and fruit consumption. Associations between fruit consumption and contents of
 1844 protein, carbohydrates, fiber, water, phenols, and tannins were not statistically significant (all
 1845 $p > 0.09$). The model examining the relationship between seed size and the frequency at
 1846 which fruit genera were consumed suggested that seed size was not influential, as there were
 1847 no significant differences between the consumption frequencies of different seed size
 1848 categories (Table 3.3).

1849 *Table 3.3. Summary of generalized linear mixed model examining the relationship between*
 1850 *seed size and fruit consumption by mandrills.*

Estimate		Standard Error	<i>z</i>	<i>p</i>
Fixed Effects				
Intercept	-5.27	0.51	-10.44	<0.001
Seed Size: Medium	-0.26	0.51	-0.52	0.61
Seed Size: Small	0.31	1.03	0.30	0.77
Seed Size: Very Large	-3.44	2.23	-1.54	0.12
Seed Size: Very Small	-0.06	1.02	-0.06	0.96
Random Effects				
Genus	Random Intercept Variance	4.57		

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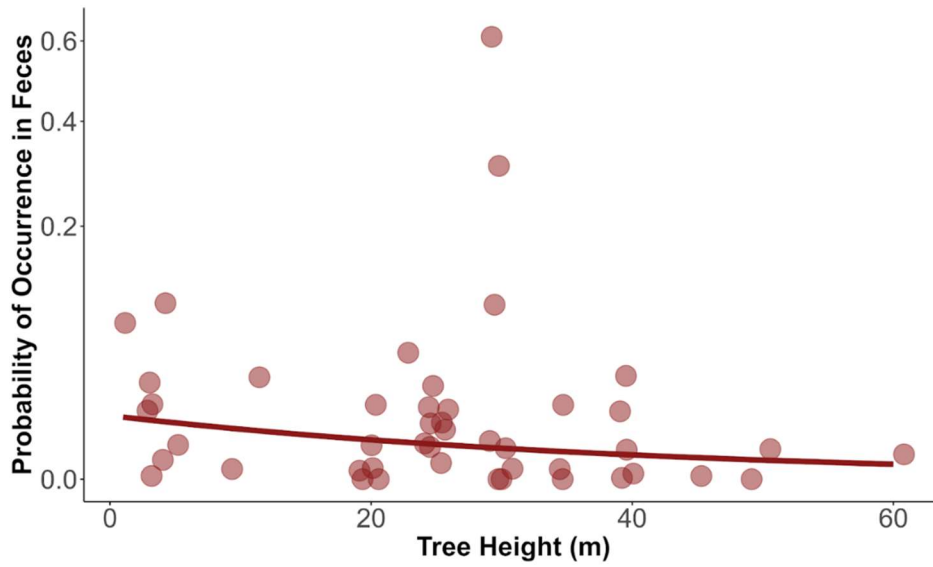
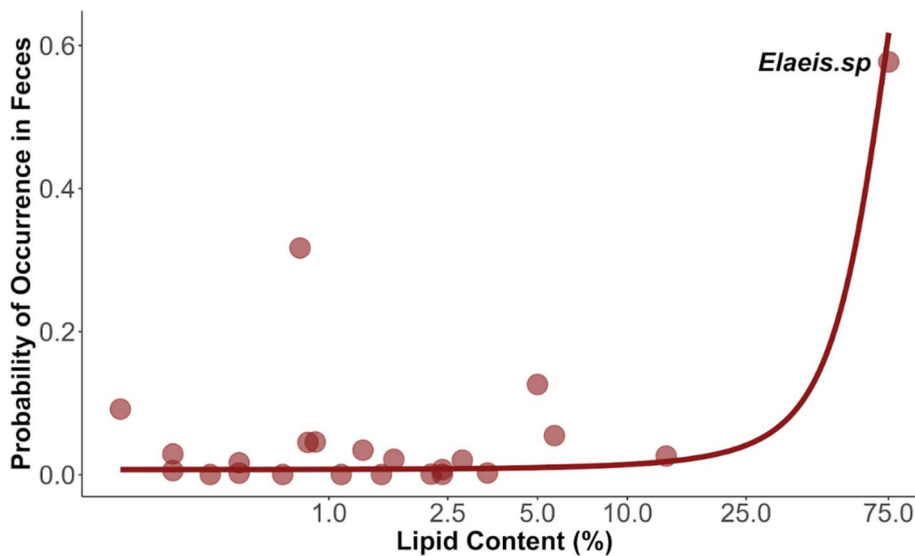


Figure 3.7. The relationship between maximum tree height and fruit consumption by mandrills. Points represent a single fruiting tree genus and their position on the x and y axes indicate the maximum height of each genus in the dataset and proportion of all 4024 fecal samples in which each genus was present, respectively. The line displays the predicted probability for trees of a given height being present in a mandrill fecal sample, based on a generalized linear mixed model. The y axis has been square root transformed and a jitter added to the data to better display overlapping points.

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1877 *Figure 3.8. The relationship between fruit lipid content and consumption by mandrills. The x*
1878 *axis has been log¹⁰ transformed to spread clustered data at lower values. Points represent a*
1879 *single fruiting tree genus and their position on the x and y axes indicate the mean lipid content*
1880 *of fruits of each genus in the dataset and proportion of all 4024 fecal samples in which each*
1881 *genus was present, respectively. The line displays the predicted probability for trees with a*
1882 *given lipid content being present in a mandrill fecal sample, based on a generalized linear*
1883 *mixed model. Removing *Elaeis sp.* (oil palm) from the data set resulted in no statistically*
1884 *significant relationship between lipid content and fruit consumption.*

1885

1886 **3.5 Discussion**

1887 Using rare, long-term data on mandrill feeding ecology and African tree phenology (Tutin *et*
1888 *al.*, 2017), we found that mandrills are highly frugivorous and that fruit consumption tracked
1889 availability. With generalized additive models, we also revealed that mandrills maintain a
1890 consistent diet breadth throughout the year, in spite of fluctuating fruit availability, via
1891 fallback food consumption of certain categories of dietary items. Finally, we found variation
1892 between fruiting genera in the extent to which they are preferred by mandrills, at least insofar
1893 as consumption covaries with availability.

1894 3.5.1 Mandrill Diet

1895 Fruit was the most frequently present food type in mandrill fecal samples (Figure 3.2) and it
1896 exhibited two annual consumption peaks, corresponding to the wet seasons, when fruit
1897 availability is highest (Figure 3.3). Similarly, we found strong positive covariance between fruit
1898 availability and fruit consumption (Table 3.2). Together these results indicate that fruit was
1899 the most preferred resource of our focal horde. In contrast, consumption of animal parts,
1900 dicot leaves, crushed seeds and other fibers tended to peak when fruit consumption was
1901 relatively low (Figure 3.3). The fact that mandrills displayed a consistent diet breadth
1902 throughout the year, despite fluctuating fruit consumption, confirms that increased dietary
1903 diversity occurred on a seasonal basis, when fruit consumption fell (Figure 3.4A).
1904 Furthermore, we found that consumption of animal parts, dicot leaves, crushed seeds, and
1905 other fibers covaried negatively with fruit consumption (Figure 3.4B). These four food types
1906 therefore appear to constitute important fallback foods for mandrills, on which they feed to
1907 make up for seasonal falls in fruit availability. Consumption of fungi and monocot leaves
1908 instead peaked during the two wet seasons (Figure 3.3), but the frequency of their
1909 consumption was lower than other fruit alternatives and not significantly related to fruit
1910 consumption (Figure 3.4B). These results are evidence that when alternate food types are
1911 available alongside fruit, they are not frequently consumed, further supporting that fallback
1912 foods are primarily consumed due to a lack of fruit. The only apparent exception to these
1913 trends is animal parts, consumption of which peaked during three periods: once at the end of
1914 the long wet season, when fruit consumption declined, but also at the beginning of the long
1915 wet season as fruit consumption increased, and during the short wet season, when fruit
1916 consumption was relatively high (Figure 3.3). Animal parts may thus also constitute an
1917 important and preferred resource of mandrills.

1918 Our results corroborate those of previous investigations indicating that mandrills are
1919 frugivorous omnivores, whose diets exhibit seasonal responses to fruiting tree phenology
1920 (Rogers *et al.*, 1996; Hongo *et al.*, 2018). Given the eight-year time period during which our
1921 data were collected, the large number of dietary samples, and a robust measure of fruit
1922 availability (Bush *et al.*, 2017, 2018, Cardoso *et al.*, 2020), our results provide an atypically
1923 robust reflection of the diet of the focal horde. In turn, the general between-study consistency
1924 in mandrill diets, across multiple locations (Jouventin, 1975; Lahm, 1986; White *et al.*, 2010;

1925 Nsi Akoue *et al.*, 2017; Hongo *et al.*, 2018), supports their classification as highly generalist
1926 frugivore-omnivores. Further study using advanced techniques, such as stable isotope
1927 analysis (Crowley, 2012) or metagenomics (Srivathsan *et al.*, 2016) may prove useful for
1928 identifying the full taxonomic diversity of mandrill diets, or for describing the impacts of
1929 extreme dimorphism and group size on mandrill feeding ecology. For example, species such
1930 as *Pentaclethra macrophyla* are known to be eaten by the focal horde, but the large seeds
1931 could not be identified in fecal samples, because they are crushed during mastication.

1932

1933 3.5.2 Mandrill Nutritional Strategy

1934 The feeding strategies employed by primates are thought to have evolved to allow sufficient
1935 micro- and macro-nutrients to be obtained from their habitat (Felton *et al.* 2009a). For
1936 example, spider monkeys (*Ateles chamek*) and chimpanzees (*Pan troglodytes*) appear to
1937 prioritize protein intake, balancing their consumption of carbohydrates and lipids accordingly
1938 (Felton *et al.*, 2009b; Uwimbabazi *et al.*, 2021). Conversely, mountain gorillas (*Gorilla beringei*)
1939 living in a protein rich habitat, target foods allowing energy intake to be maximized through
1940 sugars (Ganas *et al.*, 2008; Rothman *et al.*, 2011). To our knowledge no formal analyses of
1941 mandrill nutritional ecology are available, and so we also explored whether the preferences
1942 of our focal horde were related to fruit nutritional contents.

1943 Our data indicated that some fruit were consumed more than others as a function of
1944 availability (Figure 3.6), suggesting preference by the focal horde for certain fruit genera
1945 (Hongo *et al.*, 2018). These preferences are likely to be driven by the nutritional contents of
1946 particular fruits (Felton *et al.*, 2009a; 2009b). However, in our analyses of the relationship
1947 between fruit nutritional contents and consumption, we found a statistically significant
1948 association only between lipid content and consumption frequency. An association between
1949 fruit lipid content and consumption suggests that energy intake plays an important role in
1950 determining mandrill food selection. The significant association we found, however, was
1951 driven largely by consumption of oil palm fruits, which were by far the most frequently
1952 consumed resource (Figure 3.5) and contain 75% lipids. Such high consumption frequencies
1953 of palm fruits suggest that these are a keystone resource to mandrills in our study site, as for
1954 primates throughout Africa (Ornstein *et al.*, 2020).

1955 Unfortunately, we did not have abundance data for palm fruits, but palms are known to fruit
1956 year-round in the study site (White, 2007). Oil palms may therefore comprise an energy
1957 source to our focal horde that is consistently available throughout the year. If energy from
1958 palm fruits is targeted to a greater degree, year round availability of lipid rich fruits may
1959 explain why we found no meaningful association between carbohydrate content (an alternate
1960 source of energy) and fruit consumption. More specifically, an abundance of available lipid
1961 rich palm fruits may have reduced the need for mandrills to pursue caloric intake in the form
1962 of carbohydrates.

1963 We also did not find an association between fruit protein content and consumption
1964 frequency. A low influence of protein content on mandrill fruit selection could arise because
1965 fruits are typically low in protein, compared to other food items that primates feed on
1966 (Rothman *et al.*, 2014). Thus, individuals in the focal horde may not have selected fruits due
1967 to protein content, because they gain more substantial amounts of protein from other food
1968 sources. For example, invertebrates are often a key component of primate diets and can
1969 contain higher amounts of protein than other food types, such as fruit (Rothman *et al.* 2014).
1970 High nutritional value of invertebrates, in terms of protein content, may explain our finding
1971 that mandrills frequently consume animal parts when fruit availability is high. Indeed, animal
1972 prey often contributes a substantial amount of protein to primate macronutrient intake
1973 (O'Malley and Power, 2014; Bryer *et al.*, 2015). The quantitative contribution of animal prey
1974 to the diet of mandrills, both in terms of consumption volume and nutritional intake, may
1975 therefore be more substantial than our presence versus absence approach has revealed here.
1976 Future work to quantify the amount of animal prey consumed by mandrills would be useful
1977 and may demonstrate that animal parts are not simply a fallback food, but a preferred
1978 resource, potentially due to the need for high protein intake.

1979 We did not find statistically significant relationships between fruit consumption and contents
1980 of fiber, water, tannins, or phenols. We therefore did not find evidence of factors outside of
1981 macronutrients, such as plant secondary compounds, affecting fruit selection by mandrills.
1982 The consumption of fruit containing defensive compounds may form a part of mandrills'
1983 extremely generalist feeding strategy if they are less selective than other primates in terms
1984 of fruit quality. Alternatively, nutritional data on a greater range of fruit species may be
1985 needed to further examine the effect of secondary compounds. We only had nutritional

1986 contents data for a subset of the fruit genera consumed by our focal horde, and it would
1987 therefore be useful to increase the taxonomic coverage of our dataset, to carry out a more
1988 complete analysis of the nutritional ecology of the focal horde. It would also be interesting to
1989 compare the nutritional ecology of mandrills in different locations or habitat types, to
1990 examine the effects of the presence or absence of particular resources, such as oil palms
1991 (Hongo *et al.*, 2018). Nutritional analyses may also reveal the potential fitness consequences
1992 of obtaining preferred versus fallback foods and give insights into morphological trait
1993 evolution and the resilience of mandrills to environmental change.

1994

1995 *3.5.3 Feeding Competition*

1996 The numbers of mandrills present in a horde, the highest of any non-human primate
1997 (Abernethy *et al.*, 2002) mean that the food demands of a single group are extraordinarily
1998 high. The biomass of the mandrill horde is around 4852 kg (White *et al.*, 2010), roughly
1999 equivalent to the average group biomass for forest elephants at the site (4876 kg; White,
2000 1994) and close to twelve times the group biomass for the largest frugivorous primate, the
2001 lowland gorilla (414 kg; White, 1994). These high food demands, as well as rapid rates of patch
2002 depletion are the most likely explanation for mandrills also exploiting the largest home-ranges
2003 observed in wild primates (White *et al.*, 2010). It is also probably the case that the dietary
2004 generalism documented here and elsewhere (Hongo *et al.*, 2018) is, in part, an adaptation to
2005 extreme group sizes. We have documented that the diversity of food types eaten by mandrills
2006 increases when fruit availability is low. This ability to switch resources in response to a lack of
2007 fruit likely carries over to when individual animals are unable to access fruit in feeding
2008 patches. Thus, dietary generalism may allow individuals to continue to intake resources even
2009 when they lose out during scramble competition within a mandrill horde. Mandrills are also
2010 the most sexually dimorphic primate (Setchell, 2016) and sex differences in diet have been
2011 observed in other study sites, such as males consuming more hard foods than females (Nsi
2012 Akoue *et al.*, 2017; Percher *et al.*, 2017). Therefore, it could also be the case that mandrill
2013 dimorphism facilitates some degree of niche differentiation between sexes, that helps to
2014 alleviate feeding competition in large social groups.

2015 Mandrills must also cope with interspecific feeding competition. In Lopé, mandrills coexist
2016 with several other frugivores (11 other monkey species, chimpanzees, gorillas, red river hogs,
2017 and forest elephants) and our focal horde's dietary niche may be influenced by feeding
2018 competition and resource partitioning among species. Segregating habitats by height is one
2019 way in which coexisting frugivores may attempt to alleviate feeding competition (Sushma and
2020 Singh, 2006). Mandrills are often considered semi-terrestrial primates, as opposed to purely
2021 arboreal, because they forage mostly, but not exclusively, at ground level (Hoshino, 1985).
2022 However, the negative association we found between maximum tree height and fruit
2023 consumption was relatively weak (Figure 3.7) and so does not suggest that tree height greatly
2024 limits mandrill access to certain fruits or that arboreal habitats are avoided. Notably, oil palm
2025 and *Uapaca.sp* are medium sized trees (within our sample) and their fruits were more
2026 frequently consumed than any others. Furthermore, because we found only a modest effect
2027 of maximum height, and most trees are necessarily shorter than the species maximum height,
2028 it is very unlikely that mandrills are limited in terms of foraging height throughout most of
2029 their habitat.

2030 Rogers *et al.* (1996) found high levels of leaf litter disturbance when following a mandrill
2031 horde at Lopé, suggesting substantial amounts of ground level food was eaten. Consumption
2032 of fallen fruit and seeds at ground level most likely explains why we often observed
2033 consumption of fruit species when our phenology data indicated zero availability in the
2034 canopy (Figure 3.6). Our focal horde also exploited a 182 km² home-range during the same
2035 time period as feces were collected (White *et al.*, 2010). This estimate is much larger than
2036 those for sympatric frugivores at other sites in Central Africa, such as forest elephants at 75
2037 km² (Blake *et al.*, 2008), gorillas at 38 km² (Sanz, 2004) and chimpanzees at 18km² (Cipolletta,
2038 2004). Extreme fruit scarcity in 2004, the final year of fecal sample collection, also resulted in
2039 group-fissioning by the focal horde, with sub-groups breaking away from the horde to forage
2040 in other areas of the home-range (White, 2007). High mobility and social plasticity, alongside
2041 the dietary generalism documented here and by others (Rogers *et al.*, 1996, Nsi Akoué *et al.*,
2042 2017; Hongo *et al.*, 2018) may all play a role in allowing mandrills to coexist with many
2043 competing species. Examining the exact nature of feeding competition (and facilitation)
2044 between sympatric frugivores in Gabon would be another interesting avenue for future study.

2045

2046 3.5.4 Mandrill Conservation

2047 Evaluating the resilience of mandrills to environmental change is necessitated by their
2048 classification as Vulnerable on the IUCN Red List, with habitat degradation from climate
2049 change one of the threats identified (Abernethy and Maisels, 2019). The high social and
2050 dietary plasticity of mandrills may be a response to the extreme variation in resource
2051 availability in West Central Africa over the past millennia (Maley, 1996; White, 2001). Dietary
2052 flexibility, in the form of resource switching, could therefore confer some level of resilience
2053 to climate change-induced falls in fruit production (Korstjens and Hillyer, 2016). However,
2054 mandrill hordes are so large, with an approximate biomass of 4.1 tons (derived from
2055 Abernethy *et al.*, 2002) that availability of alternative foods may be limited. At our study site,
2056 climate change over the past three decades has resulted in a 1°C rise in temperature,
2057 alongside a 300mm fall in annual rainfall and a longer dry season (Bush *et al.*, 2020b).
2058 Consequently, fruit availability has fallen dramatically at Lopé in the last 30 years (Bush *et al.*,
2059 2020a).

2060 For comparison, forest elephants, the largest mammal found at Lopé and also a frugivore,
2061 have a comparable mean group biomass of around 4.8 tons (White, 1994). A comparison of
2062 recent and historical elephant fecal samples from Lopé indicated that elephants are
2063 consuming less fruit and much greater quantities of low-quality, fibrous plant tissues in 2022
2064 than they did in 1990 (Tejler *et al.*, unpublished data). This shift in diet, likely a response to
2065 lower fruit availability, has been linked to increased seasonal emaciation of elephants
2066 documented at Lopé (Bush *et al.*, 2020), suggesting that elephants have been unable to
2067 maintain the nutritional quality of their diet faced with current levels of tree productivity. As
2068 our fecal data were gathered between 1996-2004, they could provide a useful baseline for
2069 comparison to newer data, to examine whether the Lopé mandrills' diet has changed in the
2070 intervening 20 years. Contemporary data could highlight whether mandrills are consuming
2071 the same fruits as before, or whether they are having to consume fruit which we found to be
2072 less preferred. Additionally, repeating our analysis could indicate whether temporal
2073 consumption of fruit has remained consistent, or if fallback foods are forming a greater part
2074 of present day mandrill diets.

2075 An understanding of mandrill resilience to environmental change is crucial not only to protect
 2076 mandrills, but also to conserve forests throughout their range. Our study has highlighted that
 2077 mandrills primarily consume fruit, which likely means they play crucial functional roles,
 2078 including as seed dispersers and predators. We found that mandrills consumed fruit with a
 2079 range of seeds sizes at similar frequencies. However, forest elephants in Gabon have been
 2080 found to consume whole seeds up to 21 cm in diameter (Feer, 1995), which is a great deal
 2081 larger than the seeds of any fruit genus that appeared in mandrill fecal samples. Different
 2082 frugivores may therefore act as seed dispersers for distinct assemblages of fruiting trees and,
 2083 given that mandrill home-ranges are so large compared to many sympatric frugivores, they
 2084 may play a unique functional role as seed dispersers within Central African forests. In addition,
 2085 the importance of invertebrate foods we have documented means that mandrills also
 2086 influence the wider food web, by impacting invertebrate communities and the trophic
 2087 interactions those species are involved in.

2088 Our detailed analysis of how environmental parameters are influencing mandrill dietary
 2089 strategies reflect the importance of this species within forest phenology and trophic
 2090 structure, as well as its vulnerability to oncoming climate-driven environmental changes. For
 2091 a species with such a restricted global distribution, this is of importance to international
 2092 assessments (IUCN Red List) of the species' conservation priority.

2093 **3.6 Appendix**

2094 *Table A3.1. All plant taxa identified in mandrill feces. The lowest taxonomic level identified in*
 2095 *each family is highlighted in bold. Where a genus was only represented by a single species in*
 2096 *Lope National Park, this species was taken as present in the diet. Where no members of a*
 2097 *genus or family were identified to species level, the genus or family was taken as present in*
 2098 *the diet. This resulted in a minimum dietary diversity of 67 identified plant taxa and a*
 2099 *maximum dietary diversity of 119 plant taxa, if all species of each family found in mandrill*
 2100 *feces were assumed to be eaten. In addition to these 119 taxa, species known from*
 2101 *observational data to be consumed by the horde are highlighted in red, meaning the total*
 2102 *diversity of taxa consumed by mandrills is higher than we found through fecal analysis.*

Family	Genus	Species
Zingiberaceae	<i>Aframomum</i>	<i>sericeum</i> , <i>longipetiolatum</i> , <i>leptolepsis</i>
Annonaceae	<i>Annonidium</i>	<i>Floribundum</i>

Euphorbiaceae	<i>Antidesma</i>	venosum, vogelianum
Lauraceae	<i>Belschmeidia</i>	Fulva
Ochnaceae	<i>Campylospermum</i>	Elongatum
Zingiberaceae	<i>Costas</i>	afra
Ulmaceae	<i>Celtis</i>	tessmannii
Vitaceae	<i>Cissus</i>	dinklagei
Rutaceae	<i>Citrus</i>	limonum
Sterculiaceae	<i>Cola</i>	lizeae
Olacaceae	<i>Coula</i>	edulis
Euphorbiaceae	<i>Croton</i>	mubango
Leguminosae	<i>Cryptocephalum</i>	staudtii
Leguminosae	<i>Detarium</i>	macrocarpum
Leguminosae	<i>Dialium</i>	lopense, pachyphyllum, guineense, dinklagei, soyauxii
Ebenaceae	<i>Diospyros</i>	dendo, manni, polystemon, zenkeri
Euphorbiaceae	<i>Discoglyprena</i>	caloneura
Sapotaceae	<i>Donella</i>	ogowensis
Tiliaceae	<i>Duboscia</i>	macrocarpa
Areceaceae	<i>Elaeis</i>	guineensis
Annonaceae	<i>Enantia</i>	chlorantha
Moraceae	Ficus	<i>barteria, carica, cyathistipuloides, elasticoides, kimuenzensis, mucoso, ovata, polita, recurvata, subsagittifolia, sur, thonningii, variifolia</i>
Sapotaceae	<i>Gambeya</i>	subnuda
Sapindaceae	<i>Ganophyllum</i>	giganteum
Clusiaceae	<i>Garcinia</i>	gnetoides, ovalifolia
Olacaceae	<i>Heisteria</i>	parvifolia
Acanthaceae	<i>Hypoestes</i>	verticillaris
Irvingiaceae	<i>Irvingia</i>	gabonensis
Irvingiaceae	<i>Klainedoxa</i>	gabonensis
Apocynaceae	<i>Landolphia</i>	incerta, jumellei, manni
Anacardiaceae	<i>Lannea</i>	welwitschii
Rubiaceae	<i>Massularia</i>	acuminata
Leguminosae	Millettia	<i>barteri, griffoniana, laurenti, manni, sanagana, versicolor</i>
Annonaceae	Monanthataxis	congolensis, diclina, klainii
Annonaceae	<i>Monodora</i>	angolensis
Moraceae	<i>Musanga</i>	cecropioides
Moraceae	<i>Myrianthus</i>	arboreus
Lecythidaceae	<i>Napoleonaea</i>	vogelii
Rubiaceae	Nauclea	didderichii, latifolia, pobeguunii, vanderguchtii
Olacaceae	Ongokea	gore
Sapotaceae	<i>Pachystela</i>	brevipes
Leguminosae	Parkia	bicolor, filicoidea
Passifloraceae	<i>Paropsia</i>	grewiodes
Mimosaceae	<i>Pentaclethra</i>	macrophylla

Clusiaceae	<i>Pentadesma</i>	butyracea
Rubiaceae	<i>Polycoryne</i>	fernandensis
Rubiaceae	<i>Porterandia</i>	cladantha
Anacardiaceae	<i>Pseudospondias</i>	longifolia, microcarpa
Myrtaceae	<i>Psidium</i>	guineensis
Hypericaceae	<i>Psorospermum</i>	febrifugum
Rubiaceae	<i>Psychotria</i>	<i>penduncularis, venosa, voegliana</i>
Myristicaceae	<i>Pycnathus</i>	angolensis
Zingiberaceae	<i>Renealmia</i>	<i>cincinnata, macrocolea</i>
Humiriaceae	<i>Sacoglottis</i>	gabonensis
Burseraceae	<i>Santiria</i>	trimera
Flacourtiaceae	<i>Scottellia</i>	coriacea
Olacaceae	<i>Strombosiopsis</i>	tetrandra
Apocynaceae	<i>Strychnos</i>	<i>congolana, malacoclados</i>
Dilleniaceae	<i>Tetracera</i>	podotricha
Mimosaceae	<i>Tetrapleura</i>	tetraptera
Moraceae	<i>Treculia</i>	africana
Meliaceae	<i>Trichillia</i>	prieureana
Anacardiaceae	<i>Trichoscypha</i>	<i>abut, acuminata, anomala</i>
Euphorbiaceae	<i>Uapaca</i>	guinensis, heudelotii, paludosa, togoensis, vanhouttei
Annonaceae	<i>Uvaria</i>	versicolor, scabrida, klaineana, psorosperma
Annonaceae	<i>Uvariastrum</i>	pierreanum
Verbenaceae	<i>Vitex</i>	doniana
Annonaceae	<i>Xylopia</i>	aethiopica, hypolampra, parviflora, quintasii, staudtii
Cyperaceae	<i>Bulbostylis</i>	densa, laniceps
Cyperaceae	<i>Cyperus</i>	rotundus, sphacelatus
Cyperaceae	<i>Eleocharis</i>	variegata
Cyperaceae	<i>Fimbristylis</i>	pilosa
Cyperaceae	<i>Kyllinga</i>	echinata
Cyperaceae	<i>Rhynchospora</i>	corymbosa
Cyperaceae	<i>Scleria</i>	boivinii
Marantaceae	<i>Marantochloa</i>	cordifolia, filipes, purpurea
Marantaceae	<i>Megaphrynium</i>	macrostachyum, velutinum
Marantaceae	<i>Haumania</i>	liebrechtsiana
Marantaceae	<i>Hypselodelphys</i>	violacea
Marantaceae	<i>Sarcophrynium</i>	spp.
Marantaceae	<i>Trachyphrynium</i>	braunianum
Graminae		spp.

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2106 *Table A3.2. Summary of generalized additive model examining the effect of Day of Year on*
 2107 *consumption of major mandrill food types.*

Component	Term	Estimate	Standard Error	t	p
Parametric Coefficients	Intercept	1.19	0.12	9.93	<0.001
	Crushed Seeds	-1.49	0.05	-29.41	<0.001
	Dicot Leaves	-1.11	0.05	-22.32	<0.001
	Fruit	0.82	0.07	12.35	<0.001
	Fungus	-3.33	0.07	-49.97	<0.001
	Monocot Leaves	-3.53	0.07	-50.48	<0.001
	Other	-1.15	0.05	-22.85	<0.001
Component	Term	Effective Degrees of Freedom	Reference Degrees of Freedom	F	p
Smooth Terms	s(DayofYear): Animal Parts	7.73	8	139.45	<0.001
	s(DayofYear): Crushed Seeds	7.68	8	320.47	<0.001
	s(DayofYear): Dicot Leaves	7.82	8	141.43	<0.001
	s(DayofYear): Fruit	7.35	8	331.32	<0.001
	s(DayofYear): Fungus	7.74	8	100.63	<0.001
	s(DayofYear): Monocot Leaves	6.33	8	66.50	<0.001
	s(DayofYear): Other	6.32	8	323.30	<0.001
	s(Year)	7.81	8	140.09	<0.001
	Adjusted R²: 0.33, Deviance Explained: 0.27				

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2114 *Table A3.3. Generalized additive model examining the relationship between day of year and*
 2115 *overall diet breadth of mandrills.*

Component	Term	Estimate	Standard Error	<i>t</i>	<i>p</i>
Parametric Coefficients	Intercept	1.39	0.025	55.86	<0.001
Component	Term	Effective Degrees of Freedom	Reference Degrees of Freedom	<i>F</i>	<i>p</i>
Smooth Terms	s(DayofYear)	2.58	8	41.88	<0.001
	s(Year)	6.94	8	48.91	<0.001
Adjusted R²: 0.051, Deviance explained: 0.0526					

2116

2117 *Table A3.4. Generalized additive model examining the relationship between day of year and*
 2118 *breadth of fruit species consumption by mandrills.*

Component	Term	Estimate	Standard Error	<i>t</i>	<i>p</i>
Parametric Coefficients	Intercept	0.433	0.030	14.41	<0.001
Component	Term	Effective Degrees of Freedom	Reference Degrees of Freedom	<i>F</i>	<i>p</i>
Smooth Terms	s(DayofYear)	7.19	8	403.49	<0.001
	s(Year)	6.14	8	26.78	<0.001
Adjusted R²: 0.124, Deviance explained: 0.115					

2119

2120 **Chapter 4: Investigating stable isotope turnover and discrimination**
2121 **in mandrill (*Mandrillus sphinx*) hairs with a diet-switch experiment.**

2122 Joshua Bauld, David Lehmann, Jason Newton, Luc Bussi re and Katharine Abernethy
2123 conceived the research. Mandrill tissue samples were obtained by Joshua Bauld, Barth lemy
2124 Ngoubangoye, Thierry A. Tsoumbou and Cyr I. Moussadji-Kinga. Laboratory analysis was
2125 conducted by Joshua Bauld and Jason Newton. Data analysis was carried out by Joshua
2126 Bauld, with guidance from Luc Bussi re. Joshua Bauld wrote the chapter, with guidance
2127 from Luc Bussi re, Jason Newton, David Lehmann, Katharine Abernethy, and Isabel Jones.

2128 **4.1 Abstract**

2129 Stable isotope analysis of animal tissues can provide ecological insights, such as levels of niche
2130 overlap and the timing of nutritional stress. However, stable isotope analysis is optimized by
2131 species-specific knowledge of the relationship between diet and tissue stable isotope ratios
2132 and these data are unavailable for many species of primate. We aimed to address this by
2133 studying the incorporation of stable isotopes into the hairs of mandrills (*Mandrillus sphinx*),
2134 during a diet-switch that altered the isotopic signal of the diet fed to captive mandrills. We
2135 quantified the time taken for the stable isotope ratios of mandrill mustache hairs to express
2136 the diet-switch, finding the lag time of the hair nitrogen isotope signal to be 23 days and that
2137 of hair carbon to be 38 days. We also measured the growth rates of five mandrill hair types,
2138 which revealed variation between hair types, with growth rates ranging from 10.7 mm month⁻¹
2139 to 16.4 mm month⁻¹. Finally, after mustache hairs had reached or approached equilibrium
2140 with dietary nitrogen and carbon, we calculated tissue-diet discrimination factors for each
2141 isotope ratio. Hair-diet discrimination was 3.1 ‰ for $\delta^{15}\text{N}$ and 2.8 ‰ for $\delta^{13}\text{C}$. Our results
2142 increase the analytical precision of future investigations of primate stable isotope ecology.

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2147 **4.2 Introduction**

2148 The ratios of stable isotopes in animal tissues are often analyzed to investigate aspects of
2149 animal ecology (Ben-David and Flaherty, 2012). Isotopes are atoms of the same element with
2150 different numbers of neutrons. The ratios of stable isotopes in animal tissues reflect those of
2151 the diet an animal has consumed, as the atoms from food form the building blocks of animal
2152 tissues (Sulzman, 2007). This biochemical relationship between animal diets and tissues
2153 means inferences can be made about animal feeding ecology from tissue stable isotope ratios
2154 (Newton, 2016).

2155 In animals, the conversion of dietary protein to amino acids and carbohydrates tends to favor
2156 excretion of ^{14}N and uptake of ^{15}N (Kelly, 2000). Consequently, animal tissues are usually ^{15}N -
2157 enriched compared to the foods an animal has eaten, meaning the ratio of $^{15}\text{N}:^{14}\text{N}$ rises (Post,
2158 2002; Rioux *et al.*, 2020). Nitrogen stable isotope ratios have therefore proven useful in
2159 delineating trophic relationships among individuals and species in the same food web (DeNiro
2160 and Epstein, 1981; Hobson and Welch, 1992; Forero *et al.*, 2005; Hoondert *et al.*, 2021).
2161 Carbon stable isotope ratios in terrestrial ecosystems are largely influenced by photosynthetic
2162 mechanisms, with C4 plants relatively ^{13}C -enriched compared to C3 plants (Farquhar *et al.*,
2163 1989; Zhou *et al.*, 2022). These isotopic differences between plants in turn influence the
2164 carbon isotope ratios of their consumers, meaning carbon stable isotope ratios can
2165 distinguish, for example, browsing versus grazing herbivores and which of these prey types
2166 are predated by carnivores (Voigt *et al.*, 2013; Malindie *et al.*, 2020). Ratios of sulfur stable
2167 isotopes in plants largely reflect local geology meaning that, when spatial variation in
2168 geological sulfur ratios is present, sulfur isotope ratios in animal tissues may be used to track
2169 movement (Richards *et al.*, 2003; Kabalika *et al.*, 2020).

2170 This range of potential ecological information has made stable isotope analysis a useful tool
2171 in primatology (Sandberg *et al.*, 2012; Tsutaya *et al.*, 2022). For example, past studies have
2172 used the stable isotope ratios of primate tissues to draw inferences on subjects ranging from
2173 social dynamics (Oi *et al.*, 2021) and meat consumption (Oelze *et al.*, 2011), to species
2174 coexistence (Flores-Escobar *et al.*, 2020) and nutritional stress (Oelze *et al.*, 2020). The wide
2175 range of questions that may be addressed using stable isotope analysis means that application

2176 of the technique to primatology will likely continue to expand, especially for investigating
2177 species which are challenging to study through direct observation (Crowley *et al.*, 2016).

2178 Primate hair is a particularly useful tissue for several reasons (Oelze, 2016). Firstly, the durable
2179 structure of keratin, the main protein component of hair, means deterioration and changes
2180 in isotope ratios post-sampling are minimal, simplifying sample preservation (Lubec *et al.*,
2181 1987; Saul *et al.*, 2021). Secondly, under certain circumstances, such as when great apes leave
2182 night nests, hair may be sampled non-invasively (Wessling *et al.*, 2019; Oelze *et al.*, 2020).
2183 Finally, because hair growth occurs over many weeks, sequential isotopic variation between
2184 segments of hairs can reveal temporal ecological changes, such as weaning dynamics, within
2185 a single animal (Carboni *et al.*, 2022).

2186 To optimize the application of stable isotope analysis to primate hairs, especially when using
2187 sequential hair segments, knowledge of species specific hair growth rates, isotopic lag times
2188 and tissue-diet discrimination factors are useful. Hair growth rates describe the amount of
2189 time taken for an animal to produce hair of a certain length, allowing a date to be ascribed to
2190 a given hair segment (Guilfoyle *et al.*, 2022). Isotopic lag times indicate the period required
2191 for the isotope ratios in animal hair to reflect a change in diet (Oelze, 2016). Previous studies
2192 have used the term tissue turnover time to describe the time taken for hair to reflect a change
2193 in diet, which is the term used for metabolically active tissues, such as blood (Tieszen *et al.*,
2194 1983). However, hair is a metabolically inert tissue and the time taken for hair to reflect
2195 dietary changes represents a delay in hair expressing isotopic turnover in the metabolically
2196 active tissues that feed hair follicles, as opposed to true turnover within hair keratin (Ayliffe
2197 *et al.*, 2004; Rode *et al.*, 2016). We are therefore following the terminology of Oelze (2016)
2198 and using isotopic lag time to describe the time taken for hair stable isotope ratios to express
2199 a change in diet. Discrimination factors describe the typical magnitude of isotopic enrichment
2200 between animal food and tissue, which must be accounted for to accurately reconstruct diet
2201 composition from stable isotope data (Post, 2002; Pearson *et al.*, 2003; Cherel *et al.*, 2005).

2202 These species-specific measures are generally investigated using experiments with captive
2203 animals (Martinez del Rio, 2009). Regarding non-human primates, past studies have
2204 investigated stable isotope dynamics across several species and tissues (Reitsema, 2021;
2205 Nakashita *et al.*, 2013; Macharia *et al.*, 2014; Tsutaya *et al.*, 2017; Reitsema *et al.*, 2020).

2206 However, a previous review has highlighted the need to expand the range of species covered
2207 by captive experiments (Crowley, 2016). To our knowledge, no studies of stable isotope
2208 dynamics exist for the African Papionins (Papionina), a sub-tribe containing baboons (*Papio*
2209 spp.), geladas (*Theropithecus gelada*), mangabeys (*Lophocebus* spp. and *Cercocebus* spp.),
2210 drills and mandrills (*Mandrillus* spp.). Our study aimed to address this by investigating hair
2211 growth rates, isotopic lag times and tissue-diet discrimination factors for hair in mandrills
2212 (*Mandrillus sphinx*), using a diet-switch experiment. Such experiments involve altering the
2213 diet of captive animals, by adding foods of different isotopic composition, to measure the
2214 time taken for tissues to reflect the isotopic signal of the new diet.

2215

2216 **4.3 Materials and Methods**

2217 *4.3.1 Ethical Approval*

2218 The tissue sampling of CIRMF mandrills for measuring hair growth rates and stable isotope
2219 analysis received ethical approval from the University of Stirling, UK (AWERB (19/20) 034).

2220

2221 *4.3.2 Study Subjects, Housing and Diet*

2222 We conducted our experiment at the International Center for Medical Research in Franceville,
2223 Gabon. The center holds a large colony of semi-free-ranging mandrills, established in 1983.
2224 The original population consisted of 15 unrelated individuals, with any further changes in
2225 population size resulting from natural births and deaths over time. Our experimental group
2226 consisted of six male and seven female adult mandrills. One female, however, developed a
2227 respiratory infection and did not complete the experiment. Prior to the experiment the
2228 mandrills were housed in a 3.5 ha, naturally-forested enclosure where they had access to
2229 invertebrates and native plants, most of which use C3 photosynthesis. In addition to these
2230 natural foods, the group received bananas and a soy-based cake to provide carbohydrate and
2231 protein supplementation.

2232 To complete the diet-switch experiment, our study animals were transferred to a separate
2233 enclosure, containing rest sites and enrichment items, but without access to vegetation. In

2234 this enclosure the mandrills received the same banana and soy-based cake diet as prior to the
2235 experiment. However, the cake was modified for the duration of the experiment by adding
2236 blended tuna and maize during production. The soy-based cake was expected to be ¹⁵N-
2237 depleted compared to the native plants in the mandrills' previous enclosure, because soy is a
2238 nitrogen fixing plant and should therefore exhibit a lower ¹⁵N:¹⁴N ratio than C3 plants (Okito
2239 *et al.*, 2004). Maize is a C4 plant and is therefore predicted to be ¹³C-enriched compared to
2240 C3 plants and tuna, as a marine animal, is expected to be ³⁴S-enriched, relative to terrestrial
2241 foods (Finucane *et al.*, 2006; Nehlich *et al.*, 2015). The modifications made to the
2242 experimental food items were therefore anticipated to produce a depletion in mandrill tissue
2243 ¹⁵N and enrichment in ¹³C and ³⁴S.

2244

2245 *4.3.3 Tissue Sampling and Hair Growth Measurement*

2246 To facilitate transfer between enclosures and tissue sampling, each animal was immobilized
2247 using intramuscular injections of ketamine with a Telinject blowpipe, containing the
2248 anesthetic Imalgene1000 (10 mg/kg of body weight), under the supervision of vets BN and T-
2249 A T. Mustache hairs were sampled by plucking several hairs during transfer to the
2250 experimental enclosure, prior to the diet-switch, and on three subsequent occasions for each
2251 animal during the experiment. Three to four animals were re-sampled each week during the
2252 experiment, to capture the continuous response of hair isotope values to the diet-switch as
2253 opposed to sampling all animals on 3 discrete occasions.

2254 We preserved the plucked hair samples in polythene ziplock bags. On the second sampling
2255 date of each animal, we shaved a small patch of beard, eyebrow, mustache, occipital and
2256 thigh hair and recorded the location of the patch. Then on the third sampling date, we
2257 recorded the amount to regrowth of each hair type, to the nearest 0.5 mm. After sampling
2258 was complete on each animal's fourth sampling date, they were returned to their original 3.5
2259 ha enclosure.

2260

2261 *4.3.4 Stable Isotope Analysis*

2262 To analyze the isotopic lag time of mandrill mustache hairs, i.e. the time taken for hairs to
2263 isotopically express the diet-switch, we used the hairs plucked on each individuals' final
2264 sampling session. Analyzed hairs were therefore sampled during weeks 8-12 of the
2265 experiment and, because hair isotope ratios express the diet consumed over the weeks prior
2266 to sampling, the entire 12 weeks of the experiment was sampled by our approach. Analyzing
2267 multiple hair samples from each animal may have increased the chance of isotopically
2268 capturing the diet-switch in each individual, but for economic reasons this was not possible.

2269 Prior to analysis, we cleaned surficial contaminants and lipids from all hair samples using a
2270 2:1 chloroform:methanol solution. All hairs from an individual were allowed to soak for one
2271 hour in the solution, before being shaken vigorously and the solution poured away. We then
2272 left each sample to dry in an extractor unit. Once dried, we then cut hair into segments for
2273 sequential stable isotope analysis. Closest to the follicle, where hairs are thickest, we cut the
2274 hairs into 2 mm segments, approximating 5.6 days of hair growth (Results). Matching hair
2275 segments, for example, all segments 2-4 mm along each hair, were combined in a single tin
2276 cup for weighing and encapsulation. We used between 3-6 hairs per animal to ensure a 2 mm
2277 segment met the minimum mass requirement of 0.1 mg (scale readability: ± 0.001 mg),
2278 combining hairs as needed to analyze several segments per animal. As hairs thinned towards
2279 the distal end, the length of segments was increased as required for samples to weigh >0.1
2280 mg. To quantify the isotopic signal of the diet, we freeze dried samples of banana and soy-
2281 based cake and prepared samples >0.1 mg and >0.25 mg into tin cups. We used two weights
2282 for each sample, as 0.1 mg samples did not contain sufficient sulfur, and 0.25 mg contained
2283 too much carbon, for accurate analyses.

2284 We analyzed the stable isotope ratios of nitrogen, carbon and sulfur using a Thermo
2285 Scientific™ EA IsoLink™ IRMS System. Stable isotope ratios are expressed as ratios relative to
2286 international standards, which are: atmospheric N₂ for nitrogen, PeeDee Belemnite for
2287 carbon and Vienna-Canyon Diablo Triolite for sulfur stable isotope ratios. Relative ratios of
2288 heavy to light isotopes (e.g. ¹⁵N:¹⁴N) are expressed using delta notation (e.g. $\delta^{15}\text{N}$) and ‰
2289 (permil) units.

2290 During each run of the mass spectrometer, we also ran an internal laboratory standard of hair
2291 from a wild mandrill. We found the minimum reproducibility between samples of wild
2292 mandrill hair to be 0.10 ‰ for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ and 0.21 ‰ for $\delta^{34}\text{S}$.

2293

2294 *4.3.5 Statistical Analysis*

2295 We modeled hair growth using a linear mixed model with 'hair type' as a fixed factor, random
2296 intercepts for 'animal ID' and 'weekly hair growth' as the response variable. We used weekly
2297 hair growth in the model because this was the timescale on which growth was measured in
2298 the experiment. To generate 95% confidence intervals we bootstrapped the fixed effects for
2299 the mixed model, using 1000 replicates. We present the estimate of hair growth on a per
2300 month timescale for comparison to other studies, which we produced by dividing the weekly
2301 estimates by seven, to calculate daily growth rate, and multiplying this number by 30. We also
2302 present the daily growth rates in the results because we used daily growth rates to quantify
2303 the growth time of hair segments used during stable isotope analysis.

2304 To analyze isotopic lag time, we first quantified the time taken for a given segment of hair to
2305 be grown. We did this by dividing the central measure of a hair segment by the estimated
2306 daily growth rate for mustache hairs. For example, the section from 2-4mm is centered on
2307 3mm, therefore the growth time ascribed to that segment is: $3\text{mm}/0.357\text{mm day}^{-1} = 8.4$ days.
2308 The growth time for each segment was then subtracted from the sampling date for the
2309 relevant animal, to date each segment. We then calculated the time in days between the
2310 growth of each hair segment and the diet-switch (range: -79 - +70 days pre- and post-switch).
2311 Before conducting statistical analyses, we removed all data from segments of 0-2mm, as these
2312 segments include the hair follicle and follicles contain additional tissue types to hair, which
2313 may incorporate isotopes at different rates.

2314 To test whether the diet-switch had induced the expected isotope ratio changes in the hairs,
2315 we conducted linear models examining the difference in nitrogen, carbon, and sulfur stable
2316 isotope ratios between the earliest and most recently grown hair sections of each animal. We
2317 then compared the estimated magnitude of change between early and recent segments to

2318 the mass spectrometer's minimum reproducibility, to assess whether detectable changes had
2319 taken place in each stable isotope ratio, following the diet switch.

2320 Finding that detectable changes had occurred for nitrogen and carbon stable isotope ratios,
2321 we estimated the isotopic lag time of mustache hairs for each isotope, using non-linear mixed
2322 models. We included tissue stable isotope ratios of hair segments, either ' $\delta^{15}\text{N}$ ' or ' $\delta^{13}\text{C}$ ', as
2323 response variables, 'days' (before or after the diet-switch) as a fixed effect and random
2324 intercepts for 'animal ID'. We specified a self-starting four parameter logistic curve that
2325 required estimated starting values for the left-hand asymptote (mean hair isotope ratio
2326 before the diet-switch), the right hand asymptote (mean hair isotope ratio after isotopic
2327 equilibrium), a value for the inflection point of the curve (the x axis value for the midpoint
2328 between the two asymptotes) and a numeric scale parameter on the x axis. The model then
2329 estimated fitted values for each of these parameters via maximum likelihood ("ML").

2330 We took the fitted value for the inflection point as the estimated half-lives of $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$.
2331 We then calculated the time required for 95% of the total isotopic change within hairs to take
2332 place, based on the fitted values from each nonlinear mixed model and took these as the
2333 isotopic lag time for each isotope. Using the *getInitial* function from the 'stats' v4.1.3 (R Core
2334 Team, 2022) resulted in stable starting values for the $\delta^{15}\text{N}$ model, whereas the suggested
2335 starting values for the $\delta^{13}\text{C}$ model included a negative half-life estimate. Several iterations,
2336 with different starting values, were therefore required to model the isotopic lag time of $\delta^{13}\text{C}$
2337 and we selected the model with the lowest AIC (Burnham *et al.*, 2011). To generate
2338 confidence intervals for the half-life of each isotope, we conducted re-sampling with
2339 replacement, to generate 1000 simulated data sets from our original data. We then ran the
2340 same nonlinear mixed model on all 1000 data sets and took the 2.5% and 97.5% quantiles for
2341 the estimated inflection point, across all model runs, as the 95% confidence interval for $\delta^{15}\text{N}$
2342 and $\delta^{13}\text{C}$ half-lives.

2343 To calculate nitrogen and carbon discrimination factors, we used the following equation:

2344
$$\Delta X = \delta_{\text{hair}} - \delta_{\text{diet}}$$

2345 where ΔX is the discrimination factor for a given isotope, δ_{hair} is the stable isotope ratio of
2346 mandrill mustache hairs and δ_{diet} is the isotope ratio of the diet. To calculate δ_{diet} , we first

2347 used linear models with either ' $\delta^{15}\text{N}$ ' or ' $\delta^{13}\text{C}$ ' as response variables and 'food type' as a fixed
2348 effect, to estimate mean $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ of bananas and the soy-based cake. We then assumed
2349 a 50:50 intake ratio of each food type, as this was the ratio provided to the animals, to
2350 calculate δ_{diet} . As δ_{hair} , we took the predicted ' $\delta^{15}\text{N}$ ' or ' $\delta^{13}\text{C}$ ' of mandrill mustache hairs for
2351 Day 70 of the experiment, from the previous non-linear mixed models. To generate
2352 confidence intervals around each estimated discrimination factor, we took the predicted 2.5%
2353 and 97.5% quantiles for δ_{hair} on Day 70 of the experiment, from the previously described
2354 bootstrapping procedure. We then computed ΔX using the mean, lower bound and upper
2355 bound estimates of δ_{hair} , for each isotope.

2356

2357 4.3.6 Software

2358 We complete all analysis in R v4.1.3 (R Core Team, 2022). During analyses we used the 'lme4'
2359 v1.1-31 package (Bates *et al.*, 2015) for conducting linear mixed effects models, 'nlme' v3.1-
2360 161 (Pinheiro *et al.*, 2022) for non-linear mixed effects models, 'rsample' v1.1.1 (Frick *et al.*,
2361 2022), 'boot' v1.3-28.1 (Canty, 2022), 'broom.mixed' v0.2.9.4 (Bolker and Robinson, 2022),
2362 'purrr' v0.3.5 (Henry and Wickham, 2022) for bootstrapping and 'DHARMA' v0.4.6 (Hartig,
2363 2022) for checking model assumptions. We completed all data wrangling, plotting and
2364 produced all summary tables with the packages 'tidyverse' v1.3.2 (Wickham *et al.*, 2019),
2365 'lubridate' v1.9.0 (Grolemund and Wickham, 2011), 'ungevis' v0.1.0 (Wilke, 2022), 'ggplot2'
2366 v3.4.0 (Wickham, 2016), 'ggpubr' v 0.5.0 (Kassambara, 2022) and 'flextable' v0.8.3 (Gohel and
2367 Skintzos, 2022).

2368

2369

2370

2371 **4.4 Results**

2372 *4.4.1 Hair Growth*

2373 Using a linear mixed model, we found growth rates to be similar among beard, eyebrow and
2374 mustache hairs and to be slightly higher for occipital and thigh hairs (Table 4.1). Beard hairs
2375 grew at a rate of 10.8mm month⁻¹ (bootstrap 95% CI: 7.8 - 13.7mm month⁻¹), eyebrow hairs
2376 at 10.7mm month⁻¹ (4.74 - 16.6mm month⁻¹) and mustache hairs also at 10.7mm month⁻¹
2377 (4.77 - 16.6mm month⁻¹). Occipital hairs were the fastest growing hairs at 16.4mm month⁻¹
2378 (10.5 - 22.3mm month⁻¹) and thigh hairs grew at 13.4mm month⁻¹ (7.4 - 19.2mm month⁻¹)
2379 (Figure 4.1).

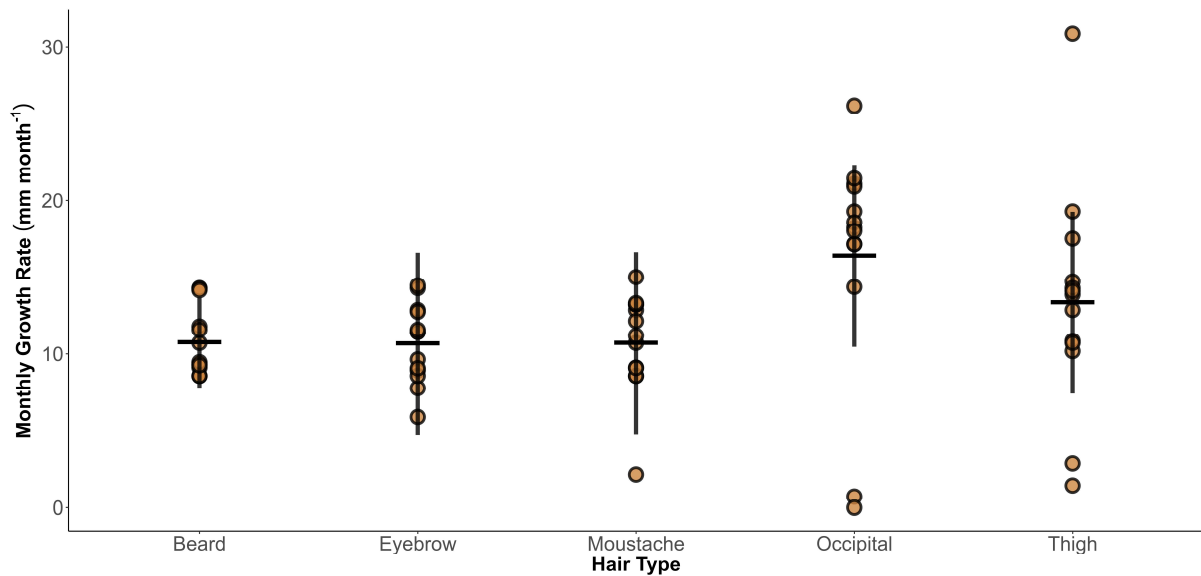
2380 *Table 4.1. Monthly and daily growth rates of each mandrill hair type.*

Hair Type	Daily Growth Rate (mm)	Lower Bound 95% CI	Upper Bound 95% CI	Monthly Growth Rate (mm)	Lower Bound 95% CI	Upper Bound 95% CI
Beard	0.36	0.26	0.46	10.78	7.8	13.67
Eyebrow	0.36	0.16	0.55	10.68	4.74	16.59
Mustache	0.36	0.16	0.55	10.72	4.78	16.62
Occipital	0.55	0.35	0.74	16.4	10.46	22.3
Thigh	0.45	0.25	0.64	13.36	7.41	19.26

2381

2382

2383

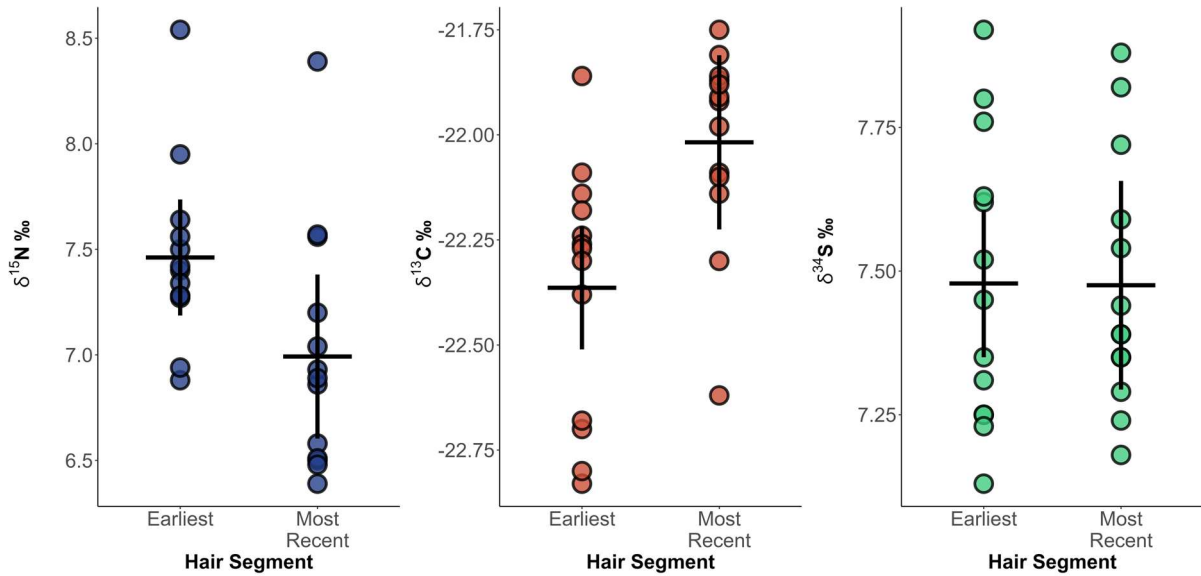


2384 *Figure 4.1. Growth rates of five mandrill hair types. Colored points show the observed growth*
 2385 *rates of each hair type, in each individual (n = 13). Black horizontal lines display the mean*
 2386 *growth rate for each hair type, vertical lines show the 95% confidence interval for the mean*
 2387 *estimate, based on 1000 bootstrap replicates.*

2388

2389 4.4.2 Isotopic Lag Times

2390 To examine whether the diet switch produced a change in the stable isotope ratios of mandrill
 2391 hairs, we conducted linear models examining the mean change between each animal's distal
 2392 hair segment (earliest grown) and the hair segments adjacent to the follicle (recently grown)
 2393 (Figure 4.2). $\delta^{15}\text{N}$ exhibited a statistically significant decrease between the earliest and most
 2394 recent segments (mean = -0.47‰ , 95% CI = ± 0.39 , $p = 0.026$). $\delta^{13}\text{C}$ showed a statistically
 2395 significant increase (0.35‰ , ± 0.21 , $p < 0.01$). The change in $\delta^{34}\text{S}$ between the earliest and
 2396 most recent hair segments was not statistically significant (-0.0031‰ , ± 0.09 , $p = 0.97$). The
 2397 minimum reproducibility of our mass spectrometer was 0.1‰ for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ and 0.21‰
 2398 for $\delta^{34}\text{S}$. Therefore the predicted changes between early and recent hair segments for $\delta^{15}\text{N}$
 2399 and $\delta^{13}\text{C}$ exceeded minimum reproducibility and were detectable, whereas the predicted
 2400 change in $\delta^{34}\text{S}$ was not.



2401 *Figure 4.2. Differences in $\delta^{15}\text{N}$, $\delta^{13}\text{C}$, $\delta^{34}\text{S}$ between the distal (earliest) ($n = 13$) and follicle*
 2402 *adjacent (most recent) ($n = 13$) hair segments of each individual mandrill. Black horizontal*
 2403 *lines display the mean growth rate for each hair type, vertical lines show the 95% confidence*
 2404 *interval for the mean estimate and data points represent the observed isotope ratio of each*
 2405 *element, in individual hair segments.*

2406

2407 To quantify the isotopic lag time of mandrill mustache hairs, we fitted non-linear mixed
 2408 models to the sequential $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ hair segment data. We did not attempt to quantify
 2409 the isotopic lag time of $\delta^{34}\text{S}$, as our previous analysis suggested that no detectable change in
 2410 $\delta^{34}\text{S}$ had taken place. We found the isotopic lag time for $\delta^{15}\text{N}$ to be shorter than $\delta^{13}\text{C}$, with
 2411 the model estimating a $\delta^{15}\text{N}$ half-life of 9.95 days (bootstrap 95% CI: 2.31 - 14.6 days) (Table
 2412 4.2; Figure 4.3). Isotopic equilibrium between mandrill mustache hairs and dietary nitrogen,
 2413 taken as the time required for 95% of the total isotopic change was 22.5 days. The isotopic
 2414 lag time for $\delta^{13}\text{C}$ was approximately double that of $\delta^{15}\text{N}$, with the $\delta^{13}\text{C}$ half-life estimated to
 2415 be 22.7 days (bootstrap 95% CI: -7.39 - 34.8 days) (Table 4.3; Figure 4.4). Isotopic equilibrium
 2416 between mandrill mustache hair and dietary carbon was predicted to be reached after 38.1
 2417 days.

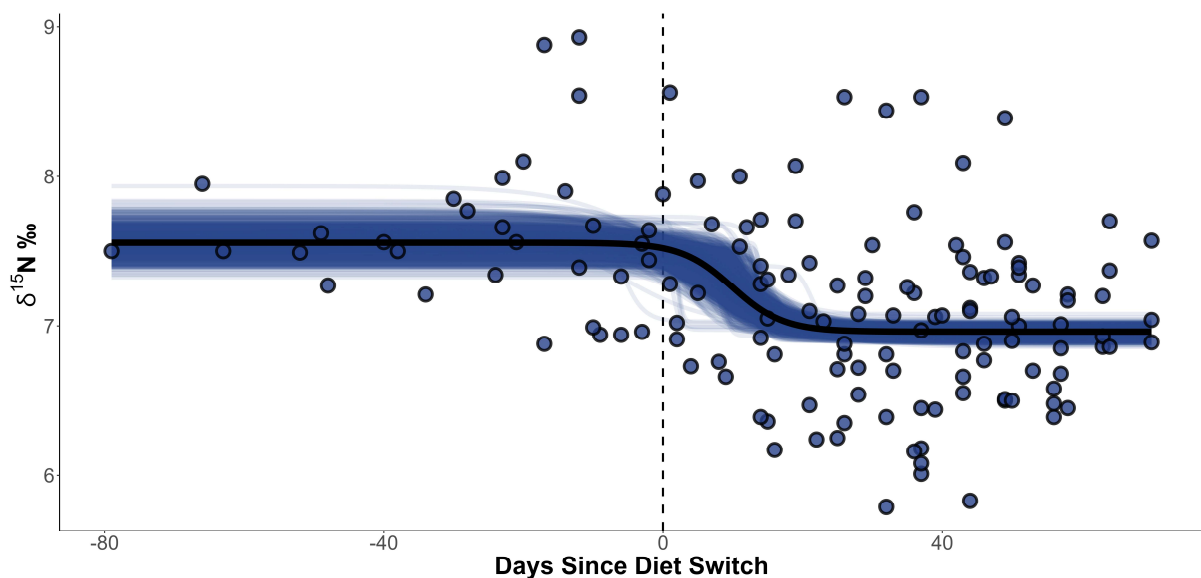
2418

2419 Table 4.2. Model summary for the non-linear mixed effects model analyzing the relationship
 2420 between days before or after the diet switch and mandrill mustache hair nitrogen stable
 2421 isotope ratio.

Estimate	Standard Error	df	t	p	
Fixed Effects					
Isotope Ratio Pre-Diet Switch	7.57	0.14	124	53.5	<0.001
Isotope Ratio Post-Diet Switch	6.96	0.15	124	45.8	<0.001
Half-life	9.95	1.91	124	5.2	<0.001
Scale Parameter	3.67	1.55	124	2.3	0.0197
Random Effects					
Isotope Ratio Pre-Diet Switch	Variance	0.18			
Isotope Ratio Post-Diet Switch	Variance	0.28			
Residual	Variance	0.084			

2422

2423



2424 Figure 4.3. The relationship between days before or after the diet-switch and mandrill
 2425 mustache hair $\delta^{15}\text{N}$. The solid black line displays the population level prediction of nitrogen
 2426 isotopic lag time and the blue lines illustrate uncertainty around that prediction, based on
 2427 1000 bootstrap replicates. Data points are raw stable isotope ratio data for each hair segment
 2428 used in the analysis (n segments = 140, n individual mandrills = 13).

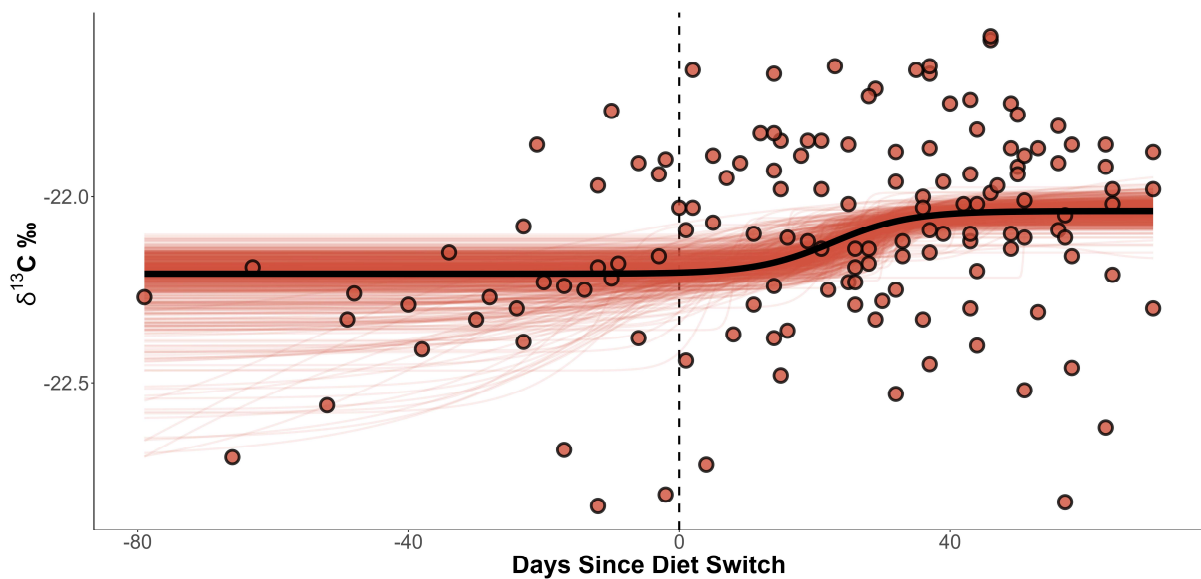
2429

2430 *Table 4.3. Model summary for the non-linear mixed effects model analyzing the relationship*
 2431 *between days before or after the diet switch and mandrill mustache hair carbon stable*
 2432 *isotope ratio.*

	Estimate	Standard Error	df	t	p
Fixed Effects					
Isotope Ratio Pre-Diet Switch	-22.2	0.068	124	-325.1	<0.001
Isotope Ratio Post-Diet Switch	-22.0	0.082	124	-269.8	<0.001
Half-life	22.72	4.23	124	5.4	<0.001
Scale Parameter	-5.7	3.5	124	-1.63	0.1
Random Effects					
Isotope Ratio Pre-Diet Switch	Variance		0.068		
Isotope Ratio Post-Diet Switch	Variance		0.048		
Residual	Variance		0.029		

2433

2434



2435 *Figure 4.4. The relationship between days before or after the diet-switch and mandrill*
 2436 *mustache hair $\delta^{13}\text{C}$. The solid black line displays the population level prediction of carbon*
 2437 *isotopic lag time and the orange lines illustrate uncertainty around that prediction, based on*
 2438 *1000 bootstrap replicates. Data points are raw stable isotope ratio data for each hair segment*
 2439 *used in the analysis (n segments = 140, n individual mandrills = 13).*

2440

2441 4.4.3 Hair-Diet Discrimination Factors

2442 To calculate discrimination factors we calculated the isotope ratio difference between δ_{hair} on
2443 Day 70 and δ_{diet} . Based on our estimates of isotopic lag times, mandrill mustache hairs should
2444 have been in equilibrium with dietary nitrogen before Day 70, but slightly short of equilibrium
2445 with dietary carbon. However, hair $\delta^{13}\text{C}$ at Day 70 was likely to be very close to the eventual
2446 asymptotic value at which equilibrium would have been reached (Figure 4.4). Calculating the
2447 difference between δ_{hair} and δ_{diet} resulted in discrimination factors of 3.1‰ (bootstrap 95%
2448 CI: 3.04 - 3.17‰) for $\delta^{15}\text{N}$ and 2.8‰ (bootstrap 95% CI: 2.76 - 2.86‰) for $\delta^{13}\text{C}$.

2449

2450 4.5 Discussion

2451 We set out to quantify growth rates, isotopic lag times and tissue-diet discrimination factors
2452 for mandrill hairs, because investigations of stable isotope dynamics within primates are
2453 sparse in the literature (Crowley *et al.*, 2016). Our experiment has revealed that mandrill hair
2454 growth rates varied between individuals and between body parts, suggesting that hairs
2455 should be sampled from the same location on different animals when sampling wild
2456 individuals. In our investigation of isotopic lag times for mandrill hairs, we found hair $\delta^{15}\text{N}$ to
2457 respond faster to the diet-switch, compared to hair $\delta^{13}\text{C}$, a result that mirrors other studies
2458 (Huelsemann *et al.*, 2009). Finally, we found isotopic discrimination factors between mandrill
2459 hairs and diet to be slightly higher for $\delta^{15}\text{N}$ than for $\delta^{13}\text{C}$, with both isotopes showing a similar
2460 magnitude of isotopic discrimination to other studies of captive primates. Our results can be
2461 applied to increase the analytical precision of future investigations of primate stable isotope
2462 ecology.

2463

2464 4.5.1 Hair Growth

2465 Knowledge of species-specific hair growth rates enhance stable isotope analysis, by allowing
2466 the time taken to grow a length of hair to be quantified and, therefore, for individual hair
2467 segments to be dated. We found substantial growth rate variation between individuals for all
2468 hair types, however, such inter-individual variation is impossible to account for when

2469 sampling wild individuals. Our measures of mean growth rate, that account for inter-
2470 individual variation, can therefore be applied during future studies of mandrill stable isotope
2471 ecology. For example, a 10.7cm length of mandrill mustache hair may be assumed to have
2472 taken one month to grow in wild individuals, based on our results. Our results also suggest
2473 that researchers should conduct hair sampling on the same body parts across individuals in a
2474 single study, to reduce error due to differing growth rates between hair types.

2475

2476 4.5.2 Isotopic Lag Times

2477 Before estimating the isotopic lag time of mandrill mustache hairs, we first examined the
2478 isotope ratios of the earliest and most recently grown hair segments, to test whether hair
2479 isotope ratios had responded to the diet switch. We found that hair $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ both
2480 responded to the diet switch, with $\delta^{15}\text{N}$ exhibiting a negative change and $\delta^{13}\text{C}$ showing a
2481 positive change. In contrast, we did not observe a change in $\delta^{34}\text{S}$ between the earliest and
2482 most recently grown hair segments. A lack of change in $\delta^{34}\text{S}$ may have resulted from low
2483 isotopic distinction between the experimental diet and the diet consumed by the study
2484 animals prior to the diet-switch. An additional possibility is that the isotopic lag time of hair
2485 $\delta^{34}\text{S}$ is too long to be captured by a 70 day experiment. A slower response of tissue $\delta^{34}\text{S}$ has
2486 previously been documented in mammals (Bahar *et al.*, 2009). In terms of mammal hairs, it
2487 has been hypothesized that $\delta^{34}\text{S}$ exhibit a longer lag time because sulfur is contained within
2488 the amino acids cysteine and methionine. Cysteine is a non-essential amino acid, meaning it
2489 may be synthesized from endogenous sulfur atoms and not only from dietary sulfur, leading
2490 to a delay in cysteine $\delta^{34}\text{S}$ reflecting a change in diet (Richards *et al.*, 2003). However, because
2491 our study animals were semi-free ranging, and their diet was not fully controlled prior to the
2492 experiment, distinguishing between these two possibilities was not possible.

2493 Having found hair $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ to have responded to the diet-switch, we then estimated the
2494 isotopic lag times of both elements. We found the half-life of $\delta^{15}\text{N}$ to be approximately 10
2495 days, and for hair nitrogen isotope ratios to be in equilibrium with the new diet after 22 days.
2496 Lag times for hair $\delta^{15}\text{N}$ are rare in the literature (Oelze, 2016). A previous investigation of a
2497 single cow found a half-life of 19 days for hair $\delta^{15}\text{N}$, approximately double our own estimate
2498 (Schwertl *et al.*, 2003). However, the same study also found hair to reach isotopic equilibrium

2499 after the equivalent of two half-lives, matching our results (Schwertl *et al.*, 2003). Similarly,
2500 an investigation of shaved rat hairs showed that hair $\delta^{15}\text{N}$ had reached equilibrium with a new
2501 diet after a maximum of 40 days, implying a half-life and lag time close to our own estimates
2502 (Caut *et al.*, 2008).

2503 Regarding the lag time of hair $\delta^{13}\text{C}$, our model estimated a half-life of 23 days, approximately
2504 double that of hair $\delta^{15}\text{N}$ and for isotopic equilibrium to be reached after 38 days. The isotopic
2505 lag time of mandrill mustache hair therefore appears similar to that of rats, with Caut *et al.*,
2506 (2008) finding that rat hair $\delta^{13}\text{C}$ also reached equilibrium after a maximum of 40 days. These
2507 results contrast somewhat with Tieszen *et al.*, (1983) who found a hair $\delta^{13}\text{C}$ half-life of 47.5
2508 days in gerbils, suggesting some interspecific differences in isotopic lag time for hair tissues.
2509 Our results are, however, consistent with studies of human hair, in which $\delta^{15}\text{N}$ has also been
2510 found to have a shorter lag time than $\delta^{13}\text{C}$ (Huelsemann *et al.*, 2009), which could suggest
2511 that this is a pattern to be expected in primate hairs.

2512

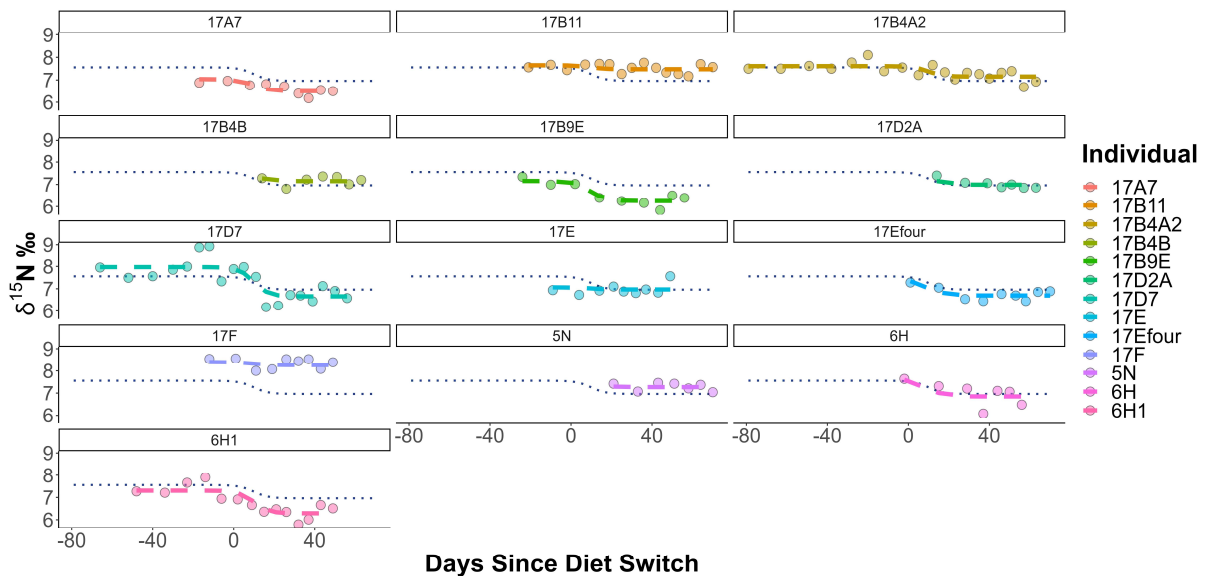
2513 4.5.3 Hair-Diet Discrimination Factors

2514 When comparing the predicted mean δ_{hair} on Day 70 of our experiment to δ_{diet} , we found
2515 discrimination factors for mandrill mustache hairs of 3.1‰ for $\delta^{15}\text{N}$ and 2.8‰ for $\delta^{13}\text{C}$. Prior
2516 investigations of tissue-hair discrimination factors for $\delta^{15}\text{N}$ have found values between 2.6‰
2517 in vervet monkeys (*Chlorocebus aethiops sabaesus*) (Macharia *et al.*, 2014) to 3.5‰ in
2518 Japanese macaques (*Macaca fuscata*) (Nakashita *et al.*, 2013), with chimpanzees (*Pan*
2519 *troglydytes*) and orang-utans (*Pongo pygmaeus*) in between (Tsutaya *et al.*, 2017; 2021). In
2520 terms of tissue-diet $\delta^{13}\text{C}$ discrimination factors, published values range between 2.6‰ in
2521 orang-utans (Tsutaya *et al.*, 2021) to 3.8‰ in chimpanzees (Tsutaya *et al.*, 2017). Our results
2522 are therefore within the ranges so far published for other non-human primate species and
2523 will aid future investigations aiming to accurately reconstruct mandrill diets from hair stable
2524 isotope ratios.

2525

2526 **4.6 Appendix**

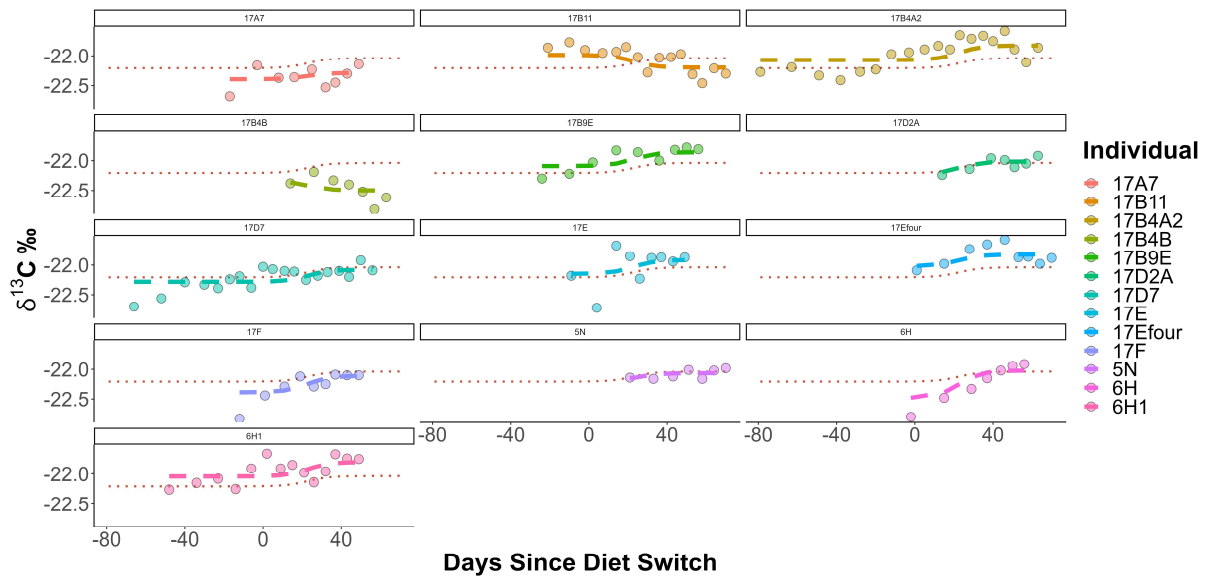
2527 Figures A4.1 and A4.2 illustrate the nonlinear relationships between days before or after the
 2528 diet switch and $\delta^{15}\text{N}$ or $\delta^{13}\text{C}$ of mandrill mustache hair, for each individual, given between
 2529 individual differences in random intercept. The predicted slope of each individual is displayed
 2530 alongside the raw stable isotope ratio of the mustache hair segments from each animal. The
 2531 plots show that there are differences between animals in the extent to which their data match
 2532 the population level predictions from the nonlinear mixed models presented in the main text.
 2533 As our nonlinear models contained random intercepts, these between animal differences are
 2534 accounted for during model fitting and we therefore believe our results are robust to inter-
 2535 individual variation in response to the diet-switch. The plots also, however, reveal the main
 2536 limitation of our study, which is that female mustache hairs are shorter than male hairs,
 2537 leading to less temporal coverage for each female. An interesting future analysis would be to
 2538 analyze additional hairs from the females, sampled earlier during the diet-switch experiment,
 2539 to increase the robustness of our estimates of hair isotopic lag times in mandrills.



2540 *Figure A4.1. The relationship between days before or after the diet-switch and mandrill*
 2541 *mustache hair $\delta^{15}\text{N}$ in each animal, given between individual differences in random intercept.*
 2542 *The dotted blue line indicates the population level prediction for $\delta^{15}\text{N}$ isotopic lag time and the*
 2543 *colored dashed lines display the $\delta^{15}\text{N}$ response of individual mandrills ($n = 13$) to the diet*
 2544 *switch. Colored points display the $\delta^{15}\text{N}$ of individual hair segments (total $n = 140$), which differ*
 2545 *in number between individuals due to inter-individual differences in moustache hair length.*

2546

2547



2548 *Figure A4.2. The relationship between days before or after the diet-switch and mandrill*
2549 *mustache hair $\delta^{13}\text{C}$ in each animal, given between individual differences in random intercept.*
2550 *The dotted orange line indicates the population level prediction for $\delta^{13}\text{C}$ isotopic lag time and*
2551 *the colored dashed lines display the $\delta^{13}\text{C}$ response of individual mandrills ($n = 13$) to the diet*
2552 *switch. Colored points display the $\delta^{13}\text{C}$ of individual hair segments (total $n = 140$), which differ*
2553 *in number between individuals due to inter-individual differences in moustache hair length.*

2554

2555 **Chapter 5: Stable isotope data suggest an association between**
2556 **extreme sexual dimorphism and resource competition in mandrills**
2557 **(*Mandrillus sphinx*).**

2558 Joshua Bauld, Jason Newton, David Lehmann, Katharine Abernethy, Isabel Jones and Luc Bussière
2559 conceived the research. Wild mandrills were sampled by David Lehmann, Lisa-Laure Nindiwe Malata,
2560 Brigitte Moussavou, Michel Louis Halbwax, Edmond Dimoto, Arthur Dibambou, Loic Makaga and Josué
2561 Edzang Ndong. Laboratory analysis was conducted by Joshua Bauld and Jason Newton. Joshua Bauld
2562 completed data analysis and wrote the chapter with guidance from Jason Newton, David Lehmann,
2563 Katharine Abernethy, Luc Bussière and Isabel Jones.

2564 **5.1 Abstract**

2565 Males and females often display niche divergence that is accompanied by sexual dimorphism
2566 in traits that are relevant to ecology, such as body size. Associations between sexual
2567 dimorphism and niche divergence suggest that selection mediated by ecology, as opposed to
2568 sexual or fecundity selection mediated by reproduction, could be involved in the evolution of
2569 sexual dimorphism. Here, we investigate niche divergence between male and female
2570 mandrills, which exhibit sexual dimorphism in body size, dental morphology, and social
2571 behavior, all of which may influence the feeding niche of each sex. To do so, we apply stable
2572 isotope analysis, which quantifies various features of animal niches, such as trophic level or
2573 basal carbon resource. We sequentially analyze the stable isotope ratios of mandrill mustache
2574 hairs, which allows multiple measures of an individuals' feeding niche to be obtained from a
2575 single tissue sample. Pooling all hair samples, we find the isotopic niches of male and female
2576 mandrills to be similar in size and to overlap extensively, suggesting dietary similarity at the
2577 level of sex. However, when analyzing mandrill isotopic niches at the individual level, we find
2578 that males are much less variable within individuals, compared to females. Lower male within
2579 individual variance suggests that males consume more consistent diets over time and that
2580 males experience different levels, or outcomes, of feeding competition to females. We end
2581 by discussing the implications of our results for the possibility that resource competition has
2582 shaped the evolution of extreme sexual dimorphism in mandrills.

2583

2584 **5.2 Introduction**

2585 *5.2.1 Ecological Sexual Dimorphism*

2586 Studies of animal niches have historically paid little attention to within-species differences
2587 among individuals (Bolnick *et al.*, 2003). However, it is increasingly clear that intraspecific
2588 niche differences are more common than previously thought (Bolnick *et al.*, 2007; Araujo *et*
2589 *al.*, 2011; Dall *et al.*, 2012). One frequently observed form of intraspecific difference is niche
2590 divergence between males and females (Summers *et al.*, 1990; Jones *et al.*, 2020), which may
2591 often be accompanied by sexual dimorphism in traits relevant to ecology, such as body size
2592 or feeding morphology, known as ‘ecological sexual dimorphisms’ (Temeles *et al.*, 2000; De
2593 Lisle and Rowe, 2015). Ecological sexual dimorphisms may arise because of sex differences in
2594 nutritional requirements, sexual selection, resource competition between males and females,
2595 or a combination of these factors (Shine, 1989; Kerneleguen *et al.*, 2015; De Lisle, 2023). At
2596 present, however, the relative importance of different evolutionary mechanisms, and
2597 particularly intersexual resource competition, to the evolution of ecological sexual
2598 dimorphisms is unknown (De Lisle, 2019; Janicke and Fromonteil, 2021).

2599 Irrespective of how ecological sexual dimorphisms evolve, associations between sexual
2600 dimorphism and intersexual niche divergence suggest an underappreciated relationship
2601 between sexual dimorphism and ecology (Chapter 2, Bauld *et al.*, 2022). Intersexual niche
2602 divergence may be indicated by sex differences in ecological means: for example, by sex
2603 differences in average trophic level (Drago *et al.*, 2015), food quality (Shannon *et al.*, 2013) or
2604 space use (Clutton-Brock *et al.*, 1987; Wang *et al.*, 2021).

2605

2606 *5.2.2 Variance Component Analysis*

2607 Another promising approach to revealing niche differences between sexes is to analyze the
2608 variance components of male and female niches (Roughgarden, 1972; de Lima *et al.*, 2019).
2609 Higher variance between individuals of one sex would imply that individuals of that sex are
2610 more ecologically diverged, relative to individuals of the sex displaying lower between
2611 individual variance. Any residual niche variance, unexplained by variance between individuals,
2612 may then be assumed to result from niche variance within individuals. Higher residual niche

2613 variance in one sex would therefore imply wider, more generalist, individual level niches for
2614 that sex, relative to individuals of the sex exhibiting lower residual niche variance (Newsome
2615 *et al.*, 2009; Yurkowski *et al.*, 2016). Applying variance component analysis to leopards
2616 (*Panthera pardus*), Balme *et al.* (2020) found the niches of individual males to be narrower
2617 than individual females, possibly because sexual size dimorphism allows males to consistently
2618 access larger prey. Variance component analysis thus provides an opportunity to identify
2619 subtle niche differences between males and females and help clarify how ecological sexual
2620 dimorphism arises.

2621 Balme *et al.*'s (2020) results, however, required the documentation of over 5000 leopard kills,
2622 and observational data on this scale is often difficult to obtain. Stable isotope analysis of
2623 animal tissues provides an alternative approach to gathering observational data for
2624 measuring the niche width of individuals and sexes (Newsome *et al.*, 2007; Shipley and
2625 Matich, 2020). The nitrogen stable isotope ratios ($^{15}\text{N}/^{14}\text{N}$) of animal tissues tend to vary with
2626 the trophic level at which an animal feeds (DeNiro and Epstein, 1981; Hobson and Welch,
2627 1992). In contrast, the carbon stable isotope ratios ($^{13}\text{C}/^{12}\text{C}$) of terrestrial animal tissues are
2628 largely determined by the plants on which they, or their prey, feed on (Codron *et al.*, 2007).
2629 This association between animal diet and tissue stable isotope ratios mean that inferences
2630 can be drawn about the feeding ecology of animals from the stable isotope ratios of their
2631 tissues (e.g. Codron *et al.*, 2006; Lewis *et al.*, 2018). For example, nitrogen and carbon isotope
2632 ratios may be considered as axes within Hutchinson's (1957) concept of ecological niches as
2633 an n -dimensional hypervolume (Bearhop *et al.*, 2004). The stable isotope ratios of animal
2634 tissues can then be analyzed in two dimensions, to construct ellipses that represent
2635 approximate characterizations of animal feeding niches (Jackson *et al.*, 2011). The ellipses of
2636 males and females may then be compared to consider niche overlap or sex differences in
2637 niche size (Thompson *et al.*, 2012; Lehmann *et al.*, 2015; French *et al.*, 2018).

2638 In the case of keratinous tissues, like hair or whiskers, dietary isotope ratios are incorporated
2639 throughout growth (McHuron *et al.*, 2016). Isotopic variation between sequential segments
2640 of linearly growing tissue thus provides a temporally structured record of dietary variation
2641 (Cherel *et al.*, 2009). This temporal record means long-term data on the feeding niche of
2642 individuals can be obtained from a single sampling event, negating the need for repeated
2643 observations of feeding behavior (Oelze, 2016). These data may then be used to conduct

2644 variance component analysis on the niches of males and females, such that: isotopic variance
2645 between individuals indicates the amount of ecological divergence between members of a
2646 given sex and residual isotopic variance, unexplained by variance between individuals, is
2647 assumed to result from isotopic variance along the hairs of individuals and is therefore a
2648 measure of the niche width of individuals of a given sex (Newsome *et al.*, 2009; Kerneleguen
2649 *et al.*, 2012; Yurkowski *et al.*, 2016).

2650 Variance component analysis of stable isotope data has previously been used to investigate
2651 ecological dimorphisms in a number of species. For example, similarity in the niche variance
2652 components of male and female seals (*Arctocephalus pusillus doriferus*) suggests that
2653 resource competition may not be more important than sexual selection to the evolution of
2654 size dimorphism in this species (Kerneleguen *et al.*, 2015). In other studies, sex differences in
2655 niche variance components have been attributed to restrictions in foraging close to juvenile
2656 rearing sites for females, but not males (Elliot Smith *et al.*, 2015; de Lima *et al.*, 2019; de Lima
2657 *et al.*, 2022; Franco-Trecu *et al.*, 2022). To our knowledge, no studies examining sex
2658 differences in isotopic niche variance components have yet suggested a role for resource
2659 competition in driving intersexual niche divergence or accompanying sexual dimorphism. In
2660 addition, no prior studies have used stable isotopes to examine sex differences in niche
2661 variance components in primates, despite evidence of intersexual niche divergence related
2662 to ecological sexual dimorphisms (Pokempner and Kamilar, 2008).

2663

2664 5.2.3 Mandrill Sexual Dimorphism

2665 Among primates, mandrills present a compelling case for a potential relationship between
2666 sexual dimorphism and niche divergence. Mandrills are the among the most size dimorphic
2667 primates (male mean = 28.9 Kg, female mean = 9.8 Kg, in our focal group, David Lehmann
2668 *unpublished data*), with males also possessing extravagantly colored ornamentation and
2669 4.5cm canines (Leigh *et al.*, 2008; Setchell, 2016). In semi-free ranging groups, females
2670 preferentially mate with the most brightly ornamented males (Setchell, 2005) and males
2671 exhibit a longer maturation period (Setchell *et al.*, 2001), alongside high reproductive skew
2672 (Setchell *et al.*, 2005); all of which suggest a role for sexual selection in the evolution of
2673 mandrill dimorphism. However, the diets of males, which possess larger jaws and teeth, have

2674 been found to include a greater proportion of hard foods than those of females, whereas
2675 females consume more invertebrates (Nsi Akoue *et al.*, 2017; Percher *et al.*, 2017). Most
2676 males also leave social groups outside of the breeding season, and are solitary for much of
2677 the non-breeding season (Abernethy *et al.*, 2002). These studies suggest an ecological
2678 component to mandrill sexual dimorphism and the potential for niche divergence between
2679 sexes. Niche divergence between males and females could in turn suggest a potential role for
2680 sex differences in nutritional requirements or resource competition in the evolution of
2681 mandrill dimorphism, alongside sexual selection.

2682 In this study, we aimed to quantify niche differences between male and female mandrills
2683 using stable isotopes. We sequentially analyzed the stable isotope ratios of nitrogen and
2684 carbon in segments of mandrill mustache hairs, to test the possibility of sex differences in
2685 isotopic niche. To facilitate our investigation, we first measured the range of isotopic variation
2686 to which mandrills in our study site may be exposed through their diets. To do so, we
2687 examined the isotope ratios of mandrill food items, testing whether these differed between
2688 food types, seasonally or spatially. These analyses allowed us to contextualize the mandrill
2689 stable isotope data, by revealing the factors that may drive individual isotopic variation.

2690 After establishing the possible sources of mandrill isotopic niche variation, we analyzed
2691 differences between the isotopic niches of males and females, in terms of size, shape and
2692 overlap of their two-dimensional isotopic niche ellipses. A previous analysis of mandrill fecal
2693 contents at our study site indicated that females are generalist feeders (Chapter 3).
2694 Comparing the isotopic niches of males and females thus allowed us to determine the amount
2695 of ecological divergence between males and females and whether males are also generalist
2696 feeders, at the level of sex.

2697 Next, we conducted variance component analysis on the isotopic niches of males and females,
2698 to describe sex differences in niche variation at the individual level. Using this approach, we
2699 able to quantify the amounts of isotopic variation between individuals of each sex, and thus
2700 measure the extent of feeding niche divergence between individual males and females. We
2701 then took the residual isotopic variance of each sex, unexplained by between individual
2702 variance, as a measure of the niche widths, and thus dietary generalism, of individual males
2703 and females. Given that males are solitary for much of the year and females are exclusively

2704 social (Abernethy *et al.*, 2002), females should experience higher levels of feeding
2705 competition and move between feeding patches more regularly (Sterck *et al.*, 1997). We
2706 therefore expected males, if solitary and feeding in distinct locations, to exhibit higher
2707 between individual isotopic variation and lower residual variation, which would indicate niche
2708 divergence between males and narrow individual feeding niches, respectively. This outcome
2709 would be consistent with males experiencing lower resource competition and so feeding
2710 more consistently than females, due to their solitary foraging strategy. Conversely, we
2711 expected females to display lower between individual isotopic variance and higher residual
2712 variance, indicating niche similarity between females and wide individual level niches,
2713 respectively. This outcome would be consistent with social females moving regularly between
2714 patches, because of high rates of patch depletion, and feeding opportunistically, due to high
2715 resource competition.

2716

2717 **5.3 Methods**

2718 *5.3.1 Authorization and Ethical Approval*

2719 The tissue sampling of mandrills, for stable isotope analysis, was approved by the Centre de
2720 la Recherche Scientifique et Technologique (CENAREST), Gabon (AR0036/16) and received
2721 ethical approval from the University of Stirling, UK (AWERB (16/17) 39).

2722

2723 *5.3.2 Study Site and Population*

2724 Our study was carried out in Lopé National Park, located in Gabon (-0.2N, 11.6E), Central
2725 Africa. Lopé National Park (4964 Km²) consists mostly of continuous, old growth forest, with
2726 savanna and gallery forests covering 10% of its northeast area. In this savanna-forest mosaic,
2727 gallery forests, with ground vegetation consisting of shrubs and lianas, spread along
2728 waterways throughout the savanna. Rainfall arrives seasonally, within a long (February to
2729 May) and a short (September to November) wet season. A long (June to August) and short
2730 (December to January) dry season take place between the two wet seasons (White, 1994,

2731 Bush *et al.*, 2020). For a more detailed description of the plant diversity in Lopé National Park
2732 see White and Abernethy (1997).

2733 Our study focused on a horde of mandrills occupying the forest-savanna mosaic and the
2734 northernmost continuous forest (White *et al.*, 2010). Observational counts suggest the horde
2735 usually consists of between 600-800 individuals, though counts as high as 1350 individuals
2736 have been recorded, when two hordes met (Abernethy *et al.*, 1997). More recent genetic
2737 work on the horde suggested an effective population size of approximately 1000 individuals
2738 (GuibingaMickala *et al.*, 2022). Most adult males leave the horde during the non-breeding
2739 season and during the breeding season adult and sub-adult males have never been observed
2740 to make up more than 12% of individuals (Abernethy *et al.*, 2002). As a result, a large number
2741 of our mandrill tissue samples were gathered during the breeding season, as conducting
2742 sampling during this period increased the chance of targeting adult males. These adult males
2743 were unknown as individuals prior to sampling because the horde is not fully habituated.
2744 Whether males were social or solitary during the feeding window represented by the isotope
2745 data in their mustache hairs was therefore unknown.

2746

2747 *5.3.3 Mandrill Food Item and Hair Sampling*

2748 We collected samples of mandrill food between February and May 2019, December 2019,
2749 and January 2020 and during May 2020. Our samples were therefore sourced during the long
2750 wet seasons of 2019 and 2020 and the 2019/20 short dry season, allowing us to test for the
2751 presence of seasonal and inter-annual shifts in the Lope National Park isotopic baseline.

2752 The focal horde is known to be frugivorous and so we prioritized collecting samples of fruit
2753 from the 15 species most frequently consumed by the horde (Chapter 2). Fruit available below
2754 2 m height was sampled directly from plants, otherwise relatively fresh fallen fruit, free of
2755 mold or bite marks, was collected. To ensure that we sampled the entire range of isotopic
2756 variation to which mandrills may be exposed through their diet, we also opportunistically
2757 sampled other fruit species, young leaves, flowers, seeds, and invertebrates. Sample
2758 collection was carried out throughout the horde's home-range, to ensure any spatial changes
2759 in the isotope ratios of mandrill food items were accounted for in our analysis. Each mandrill

2760 food sample was placed into a polypropylene centrifuge tube, alongside a small amount of
2761 indicating silica gel beads, to facilitate drying. All samples were identified to species level
2762 where possible and the corresponding tube labeled, alongside the sampling date. The location
2763 (longitude/ latitude) and a description of the microhabitat from which the sample was taken
2764 were also recorded. Upon return to the field station, each centrifuge tube was opened and
2765 placed into a large, sealed container, alongside a substantial amount of silica gel, to allow
2766 each sample to fully dehydrate.

2767 Mandrill mustache hairs were gathered between 11/05/2015 and 14/05/2019. Individual
2768 animals were anesthetized using Zoletil (Tiletamine and Zolazepam), administered via a
2769 teledart applied with a CO2 Injection Rifle (Daninject JM SP 25). Males received 80 mg
2770 Tiletamine and 80 mg of Zolazepam. Females received 25mg of Tiletamine and 25mg of
2771 Zolazepam. The age-sex class, weight, and tooth wear (for approximate aging) were recorded
2772 for each animal. Mustache hairs were sampled by plucking hair, including the follicle, and
2773 stored in polythene zip lock bags. Upon completion of tissue samples (and GPS collar fitting,
2774 see Chapter 6), each animal was allowed to recover within a soft bag, then released and
2775 visually monitored.

2776

2777 *5.3.4 Stable Isotope Analysis*

2778 All sample processing and analysis was conducted under laboratory conditions at the Scottish
2779 Universities Environmental Research Center, Glasgow, UK. Prior to analyzing the stable
2780 isotope ratios of mandrill food items, samples were freeze-dried to remove any remaining
2781 water. Samples were then placed into tin cups for weighing (scale readability: ± 0.001 mg) and
2782 encapsulation. Between freeze-drying, weighing and analysis, all plant tissues were stored in
2783 a desiccator, to prevent rehydration and spoiling. We initially carried out work to examine
2784 spatial variation in plant sulfur isotope ratios, but because of uncertainty over the period of
2785 time hair sulfur isotope ratios take to express a change in diet (Chapter 4), these results are
2786 not shown. Plant tissues often contained very low amounts of sulfur and high amounts of
2787 carbon, meaning two samples of differing mass were often needed to separately analyze
2788 sulfur and carbon isotope ratios. Nitrogen was present in intermediate amounts in plant
2789 tissues and so valid results were obtained from analysis of samples weighed for sulfur, of

2790 which we analyzed a greater number than for carbon. The sample sizes of plant tissue carbon
2791 ($n = 264$) and nitrogen isotope ratios ($n = 496$) therefore differ in our results.

2792 To prepare mustache hairs for analysis, hairs from each animal were placed in borosilicate
2793 vials containing a 2:1 chloroform:methanol solution and allowed to soak for one hour. After
2794 this period, vials were shaken vigorously to remove sediment and lipid contaminants from the
2795 hairs. We then left the tube to settle, allowing any lipid to rise to the top of the solution,
2796 before disposing of the waste liquid. The vials were then left open in an extractor unit, to
2797 allow the hairs to dry.

2798 After drying, we cut hairs into segments, to sequentially analyze variation in stable isotope
2799 ratios along the length of each hair. Nearest to the follicle, where hairs are thickest, we cut
2800 the hairs into 2mm segments, placing matching hair segments (i.e. all segments from 2-4mm)
2801 into a single tin cup for weighing and encapsulation. We used between 3-6 hairs per animal,
2802 to ensure that combined samples for each 2mm segment weighed over 100 μ g, which was the
2803 minimum mass required for analytical reproducibility (Chapter 4). As hairs became thinner
2804 towards the distal end, it was necessary to increase the length of segments used in each
2805 sample to meet the minimum mass requirement.

2806 We carried out stable isotope analysis using a Thermo Scientific™ EA IsoLink™ IRMS System,
2807 optimized for sulfur sensitivity. The stable isotope ratios of all samples are expressed using
2808 delta (δ) notation (McKinney *et al.*, 1950) and permil (‰) units, relative to an international
2809 standard, such that: $^{15}\text{N}/^{14}\text{N} = \delta^{15}\text{N}\text{‰}$ and $^{13}\text{C}/^{12}\text{C} = \delta^{13}\text{C}\text{‰}$. The international standard for
2810 nitrogen and carbon stable isotope ratios are atmospheric ^{2}N and PeeDee Belemnite,
2811 respectively.

2812

2813 *5.3.5 Statistical Analysis*

2814 To understand the factors that could influence isotopic variation in mandrill tissues, we first
2815 investigated isotopic variation between mandrill food types and whether the stable isotope
2816 ratios of forest plant tissues varied spatially or temporally. We initially used a biplot, with $\delta^{13}\text{C}$
2817 on the x axis and $\delta^{15}\text{N}$ on the y axis, to graphically explore the clustering of mandrill food items
2818 when grouped into 11 categories: C3 Fruit, C3 Leaf, C3 Seed, C3 Bulb, C3 Bark, C3 Sap, C3

2819 Flower, Forest Animal, Savanna Animal, Savannah Plant, Fungi. We then compared the 11
2820 category plot to one with five categories: Forest Animal, Forest Plant, Fungi, Savanna Animal,
2821 Savannah Plant, to decide which number of food type categories best described the isotopic
2822 clustering of mandrill food items. Using AIC scores (Burnham *et al.*, 2011) we then compared
2823 a linear mixed model containing food type categorical fixed factor and random intercepts for
2824 food species, to an intercept only model, containing only random intercepts for species, to
2825 determine which model best predicted food item $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$.

2826 To examine whether temporal or spatial variation of the isotopic baseline in mandrill habitats
2827 existed, we tested the effect of season and sampling location on the isotope ratios of forest
2828 plants. Forests are the principal habitat of mandrills and plants are the primary producers in
2829 forest habitats, as well as the main food source of mandrills. Therefore, consistent temporal
2830 or spatial variation in the isotope ratios of plants would suggest that isotopic baseline shifts
2831 occur within mandrill habitats, that would need accounting for when analyzing the isotope
2832 ratios of mandrill tissues. To explore the possibility of temporal shifts in isotopic baselines, we
2833 ran linear mixed models using season as a categorical predictor of forest plant $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$,
2834 with random intercepts for species. We then compared this model to an intercept only model,
2835 also containing random intercepts for species, using AIC. To examine the effect of sampling
2836 location on forest plant $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, we used linear mixed models containing either latitude
2837 or longitude as continuous fixed effects and random intercepts for species. We then
2838 graphically compared the predictions of these models to the raw data on which the models
2839 were based, to assess the effect size of latitude and longitude relative to the isotopic variation
2840 among forest plants.

2841 Before analyzing the isotopic niches of male and female mandrills, we first estimated the
2842 positions of each sex within isotopic space, allowing us to compare the isotope ratios of the
2843 mandrills we sampled to their food items and other studies of mandrill stable isotope ecology.
2844 We calculated mean values for each sex, for both isotopes, and constructed confidence
2845 intervals around the means using the 2.5% and 97.5% quantiles of the data for all mandrill
2846 hair segments. After estimating means and confidence intervals for each sex, we then
2847 adjusted these values by the hair-diet discrimination factors developed in Chapter 4.
2848 Discrimination factors describe the typical magnitude of isotopic enrichment between animal
2849 tissues and their dietary items, which must be accounted for before making inferences about

2850 animal diets from tissue isotope ratios. We then compared these adjusted tissue values to the
2851 isotope ratios of mandrill food items to consider the major food types consumed by mandrills
2852 in Lopé.

2853 To compare the isotopic niches of males and females, constructed ellipses representing the
2854 niche of each sex in two dimensions, with $\delta^{13}\text{C}$ on the x axis and $\delta^{15}\text{N}$ on the y axis. We
2855 compared ellipses constructed with three methods: standard ellipse area (SEA), representing
2856 two dimensional standard deviation; standard ellipse area corrected for small sample size
2857 (SEA_c); and Bayesian Standard Ellipse Areas (SEA_B), estimated with 10 000 posterior draws
2858 (see Jackson *et al.*, 2011). We found all three approaches to be comparable in terms of the
2859 estimated niche size of each sex. We therefore chose to use the standard ellipse areas for
2860 plotting, as they are convenient to display, and to compare the size and overlap of male and
2861 female niches using the Bayesian method, as this approach provides robust estimates of error.

2862 To conduct variance component analysis, we used intercept-only linear mixed models. We
2863 ran four separate models analyzing male and female variance in $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$. The isotope
2864 ratios of individual hair segments were included as the dependent variable and random
2865 intercepts were included for individual identity. We took the random intercept variance in
2866 each model as an estimate of isotopic variance between individual males or females and the
2867 residual variance as a measure of isotopic variance within individual of each sex. The sum of
2868 both variance components was taken as a measure of the total niche width of each sex, for
2869 each isotope (Newsome *et al.*, 2009; Yurkowski *et al.*, 2016).

2870

2871 5.3.6 Software

2872 All statistical analyses were conducted in R v4.1.3 (R Core Team, 2022). All data wrangling,
2873 plotting and production of tables was completed using the packages “tidyverse” v1.3.2
2874 (Wickham *et al.*, 2019), “lubridate” v1.9.0 (Grolemund and Wickham, 2011), “ggplot2” v3.4.0
2875 (Wickham, 2016) and “flextable” v0.8.3 (Gohel and Skintzos, 2022). Creation and
2876 manipulation of shapefiles and rasters was carried out with the packages “sf” v1.0-9
2877 (Pebesma, 2018), “terra” v1.6-47 (Hijmans, 2022) and “raster” v3.6-11 (Hijmans, 2022b).
2878 Stable isotope ellipses were constructed with the package “SIBER” v2.1.6 (Jackson *et al.*,

2879 2011). Linear mixed models were run using the packages “lme4” v1.1-31 (Bates *et al.*, 2015)
2880 and “nlme” v3.1-161 (Pinheiro *et al.*, 2022).

2881

2882 **5.4 Results**

2883 *5.4.1 Mandrill Food Stable Isotope Ratios*

2884 To contextualize the stable isotope ratios of mandrill mustache hairs, we analyzed the stable
2885 isotope ratios of plant and invertebrate foods available in the home-range of our study group.
2886 Firstly, to understand how mandrills may be exposed to isotopic variation through the various
2887 food types they consumed, we produced biplots displaying the $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ ratios of all
2888 mandrill food items. Grouping the foods into 11 food types (C3 Fruit, C3 Leaf, C3 Seed, C3
2889 Bulb, C3 Bark, C3 Sap, C3 Flower, Forest Animal, Savanna Animal, Savannah Plant, Fungi)
2890 resulted in numerous overlapping pair-wise combinations (Figure A5.1). Instead, grouping
2891 food types into five categories (Forest Animal, Savanna Animal, Forest Plant, Savannah Plant,
2892 Fungi) produced categories that were ecological meaningful, for example mean forest plant
2893 $\delta^{15}\text{N}$ was lower than mean forest animal $\delta^{15}\text{N}$, though extensive overlap still existed between
2894 food types (Figure 5.1). To investigate whether the five food type categories explained
2895 isotopic variation between food types, we compared linear mixed effects models containing
2896 food type as an independent variable with intercept only models, using AIC. AIC was lower for
2897 the food type model both for $\delta^{15}\text{N}$ (AIC = 2354.5 for food type model vs. 2387.8 for intercept
2898 only model) and for $\delta^{13}\text{C}$ (food type model AIC = 1451.3, intercept only model AIC = 1545.4),
2899 and so the food type models were taken as the preferred models. Invertebrates from the
2900 forest or savanna were ^{15}N -enriched relative to plants and fungi and fungi were ^{15}N -enriched
2901 relative to plants (Table 5.1). In addition, invertebrates and plants from the savanna were ^{13}C -
2902 enriched relative to invertebrates, plants and fungi from the forest and fungi were ^{13}C -
2903 enriched relative to forest invertebrates, which were in turn ^{13}C -enriched relative to forest
2904 plants (Table 5.2).

2905

2906

2907 *Table 5.1. The effect of food type on mandrill food items $\delta^{15}N$.*

		Estimate	Standard Error	t
Fixed Effects				
Intercept		6.29	0.35	17.92
Savannah Animal		0.53	1.88	0.28
Fungi		-1.50	0.83	-1.79
Forest Plant		-2.65	0.44	-6.01
Savannah Plant		-2.13	1.37	-1.55
Random Effects				
Species	Intercept Variance	2.92		
Residual	Observation Variance	2.46		

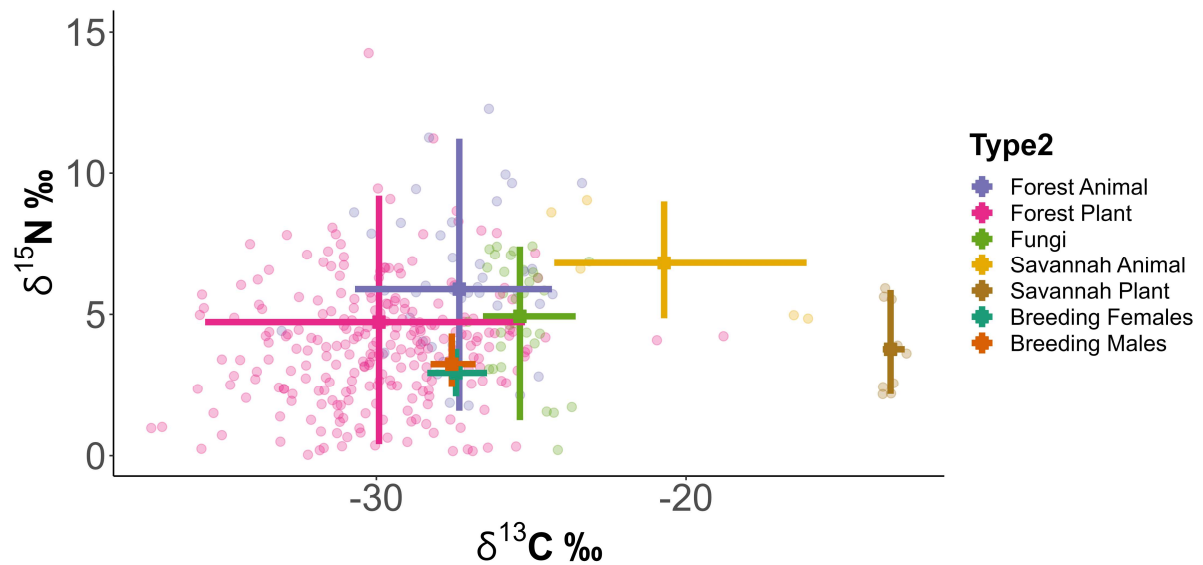
2908

2909 *Table 5.2. The effect of food type mandrill food items $\delta^{13}C$.*

		Estimate	Standard Error	t
Fixed Effects				
Intercept		-27.29	0.39	-69.53
Savannah Animal		6.59	2.03	3.24
Fungi		1.92	0.91	2.11
Forest Plant		-2.50	0.49	-5.07
Savannah Plant		13.87	1.49	9.34
Random Effects				
Species	Intercept Variance	3.37		
Residual	Observation Variance	3.02		

2910

2911



2912 *Figure 5.1. Isotopic variation between mandrill food items, when grouped into five categories.*
 2913 *Each circular point represents at individual food item sample (forest plant n = 246, forest*
 2914 *animal n = 43, fungi n = 32, savannah plant n = 9, savannah animal n = 5). Square points*
 2915 *indicate the mean of each food type and colored lines display the 95% confidence interval*
 2916 *around the mean, based on the 2.5% and 97.5% quantiles of the raw data in each category.*
 2917 *Means and 95% confidence intervals for stable isotope ratios of male (n individuals = 9, n hair*
 2918 *segments = 169) and female (n individuals = 12, n hair segments = 108) mandrill mustache*
 2919 *hairs are shown for comparison, after adjustment for tissue-diet discrimination factors*
 2920 *(Chapter 4).*

2921 To quantify the isotopic baseline values for the primary habitat of the focal horde, we fitted
 2922 intercept only models estimating the mean $\delta^{13}\text{C}$ or $\delta^{15}\text{N}$ values for forest plant (primary
 2923 producer) tissues in our study site (Table 5.3; Table 5.4). The mean $\delta^{13}\text{C}$ of forest plants was -
 2924 28.8‰ (95% Confidence Interval: -29.4 to -28.1‰) and the mean $\delta^{15}\text{N}$ of forest plants was
 2925 3.6‰ (3.1 to 4.1‰). To assess the presence of temporal isotopic baselines shifts in our study
 2926 area, we assessed whether forest plant isotope ratios varied between the long rainy seasons
 2927 in 2019 and 2020 and the 2019/20 short dry season. We found the AIC for intercept only
 2928 models for both $\delta^{15}\text{N}$ (AIC = 1,953.7) and $\delta^{13}\text{C}$ (AIC = 1,089.0) to be slightly lower than that for
 2929 models including season as predictors of forest plant $\delta^{15}\text{N}$ (AIC = 1,957.4) or $\delta^{13}\text{C}$ (AIC =
 2930 1,091.2), thus the intercept only models were taken as the preferred models.

2931 To analyze whether the isotope ratios of mandrill food items varied spatially, we conducted
 2932 linear mixed effects models examining the impact of latitude and longitude on the isotope
 2933 ratios of forest plants (Table A5.1, A5.2, A5.3, A5.4). Comparing predictions from each of these
 2934 models against the raw isotope ratios of forest plants indicated that any linear relationships
 2935 between latitude or longitude and $\delta^{13}\text{C}$ or $\delta^{15}\text{N}$ were modest, relative to the total isotopic
 2936 variation between plants (Figure A5.2, A5.3, A5.4, A5.5). Weak linear relationships imply a
 2937 lack of isotopic gradients and that isotopic variation is instead spatially heterogeneous (Figure
 2938 5.2; Figure 5.3).

2939 *Table 5.3. Model estimating mean $\delta^{13}\text{C}$ of forest plants in Lopé National Park.*

		Estimate	Standard Error	t
Fixed Effects				
Intercept		-29.78	0.35	-85.38
Random Effects				
Species	Intercept Variance	4.94		
Residual	Observation Variance	3.31		

2940

2941 *Table 5.4. Model estimating mean $\delta^{15}\text{N}$ of forest plants in Lopé National Park.*

		Estimate	Standard Error	t
Fixed Effects				
Intercept		3.65	0.24	14.98
Random Effects				
Species	Intercept Variance	2.26		
Residual	Observation Variance	2.54		

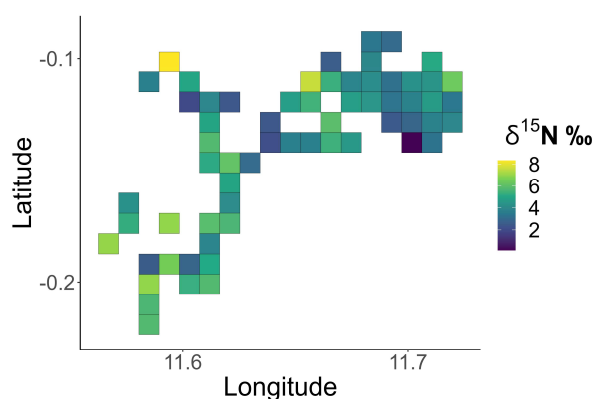
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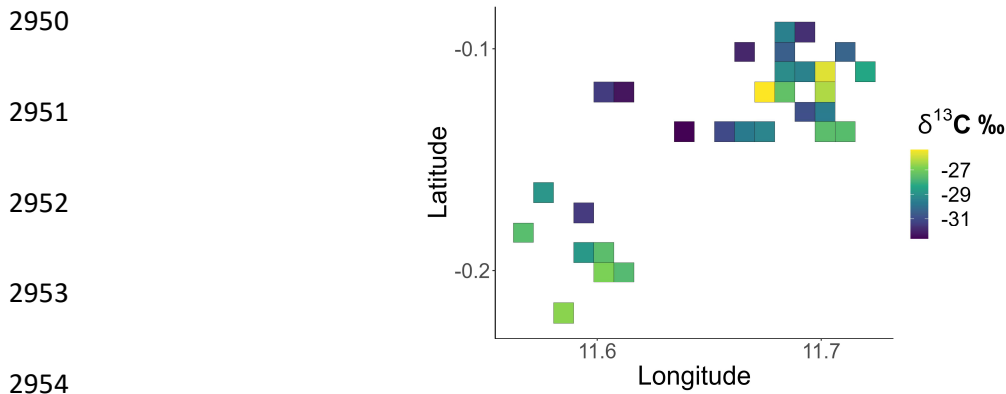
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2947 *Figure 5.2. Spatial variation in $\delta^{15}\text{N}$ of mandrill food items ($n = 471$). Pixels are equal to 1 km^2
 2948 and colored according to the mean stable isotope ratio of mandrill food items samples in that
 2949 pixel (sample sizes vary between pixels).*



2955 *Figure 5.3. Spatial variation in $\delta^{13}\text{C}$ of mandrill food items ($n = 251$). Pixels are equal to 1 km^2*
 2956 *and colored according to the mean stable isotope ratio of mandrill food items samples in that*
 2957 *pixel (sample sizes vary between pixels).*

2958

2959 5.4.2 Mandrill Isotopic Niches

2960 After segmenting mandrill moustache hairs, our final sample sizes included 160 segments
 2961 from nine individuals and 96 segments from 12 individuals for males and females,
 2962 respectively. Mean $\delta^{13}\text{C}$ of each sex was -24.60 ‰ (95% Confidence Interval: -24.90 to -24.30
 2963 ‰) and -24.70 ‰ (-25.0% to -24.50 ‰) for females and males, respectively. Mean $\delta^{15}\text{N}$ was
 2964 6.10 ‰ (5.93 to 6.27 ‰) and 6.33 ‰ (6.05 to 6.61 ‰) for females and males, respectively.
 2965 After adjusting for tissue-diet discrimination factors (Chapter 4), the mean values of male and
 2966 female mustache hair $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ were in similar locations, relative to mandrill food items
 2967 (Figure 5.1). Mean $\delta^{13}\text{C}_{\text{adjusted}}$ of each sex was -27.40 ‰ for females (95% Confidence Interval:
 2968 -27.70 to -27.10 ‰) and -27.50 ‰ (-27.80 to -27.30 ‰) for males. Mean $\delta^{15}\text{N}_{\text{adjusted}}$ was 3.00
 2969 ‰ (2.83 to 3.17 ‰) and 3.23 ‰ (2.95 to 3.51 ‰) for females and males, respectively.

2970 We compared the isotopic niche areas of both sexes using standard ellipses (SEA), standard
 2971 ellipses with a correction for small sample size (SEA_c) and Bayesian inference (SEA_B). We found
 2972 little difference between ellipse areas estimated by different approaches (Table 5.5). We
 2973 therefore chose to plot the standard ellipse areas of each sex (Figure 5.4) and to compare the
 2974 sizes and overlap between the niches of males and females using Bayesian inference. The
 2975 isotopic niches of female and male mandrills were similar in size and location within isotopic
 2976 space, but differed somewhat in shape, with the female niche obviously wider on the $\delta^{13}\text{C}$

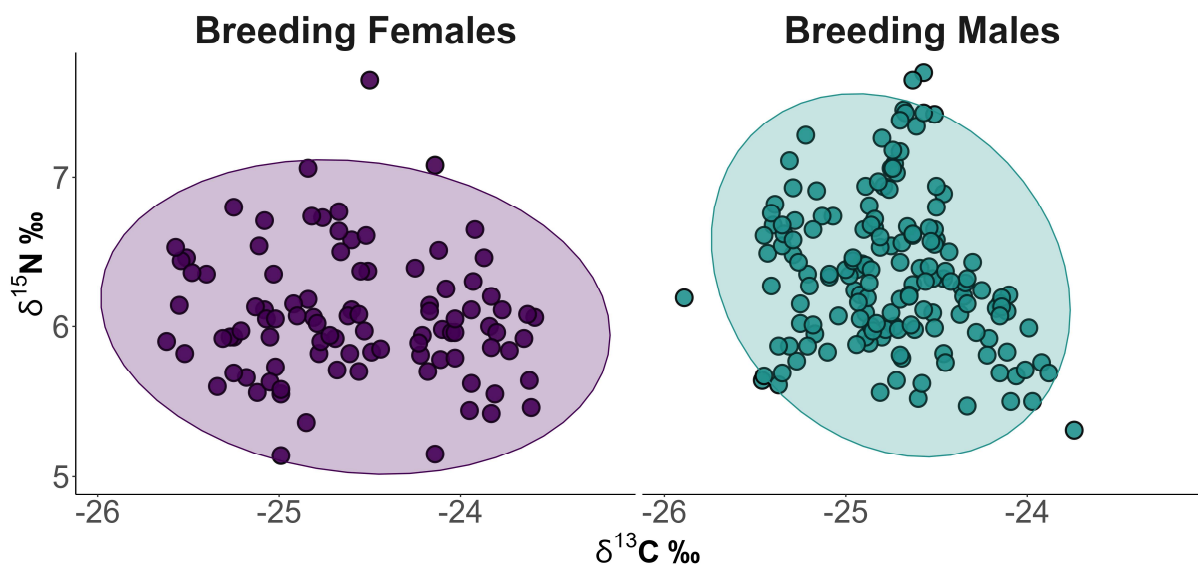
2977 axis and male niche slightly wider on the $\delta^{15}\text{N}$ axis (Figure 5.4). When comparing 10 000
 2978 posterior draws for each sexes niche size (SEA_B), the isotopic niche of females was larger than
 2979 that of males in 93.7% of cases (Figure 5.5). Regarding overlap between the ellipses of each
 2980 sex, the area of overlap covered 82.5% (95% Credible Interval: 68.7-95.9%) of the male niche
 2981 (Figure 5.6) and 67.2% (53.2-81.3%) of the female niche.

2982 *Table 5.5. Isotope niche sizes of male and female mandrills.*

Class	SEA	SEA _c	SEA _B	SEA _B 95% Credible Interval
Breeding Females	0.74	0.75	0.73	0.61-0.91
Breeding Males	0.61	0.61	0.60	0.52-0.70

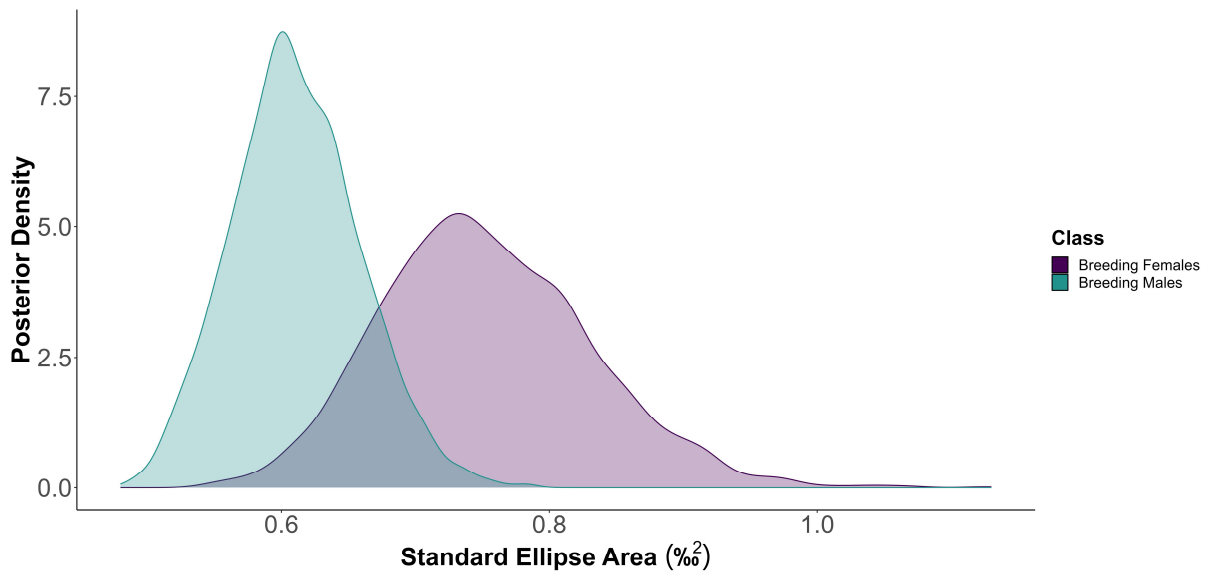
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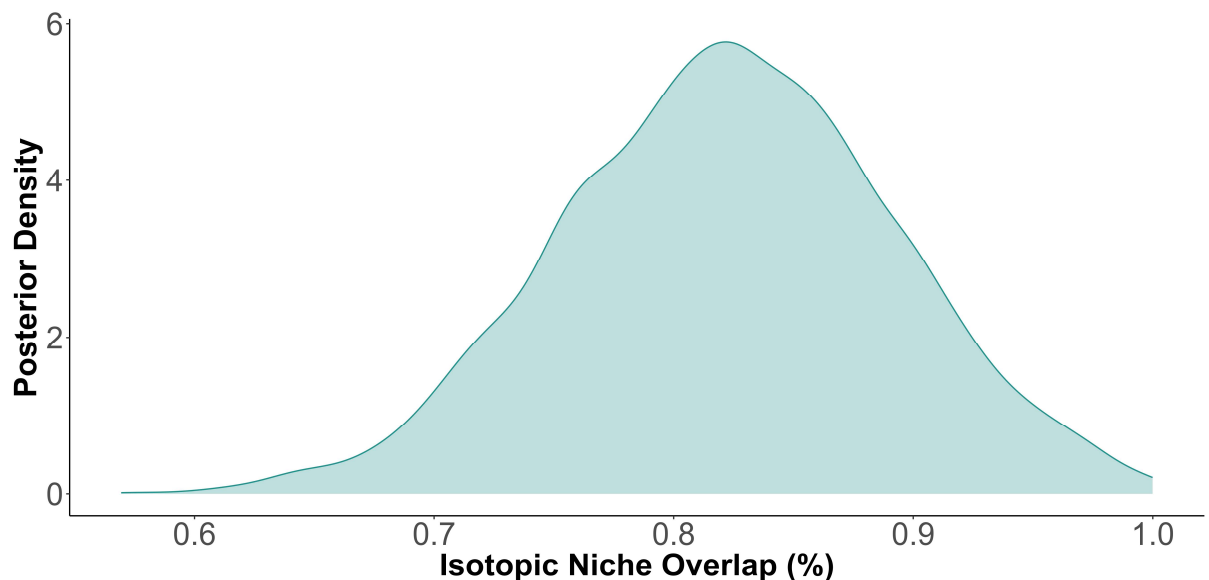
2985 *Figure 5.4. Isotopic niches of male (n individuals = 9, n hair segments = 169) and female*
 2986 *mandrills (n individuals = 12, n hair segments = 108), based on all hair segments from all*
 2987 *individuals of each sex. Each point represents the position in isotope space of each hair*
 2988 *segment and standard ellipse areas (SEA) represent the 95% isotopic niche area of each sex,*
 2989 *based on those positions.*

2990



2991 *Figure 5.5. Posterior distributions of 10 000 draws for female (n individuals = 12, n hair*
2992 *segments = 108) and male (n individuals = 9, n hair segments = 169) standard ellipse areas,*
2993 *estimated via Bayesian inference (SEA^B), using $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ stable isotope ratios of mustache*
2994 *hair segments.*

2995



2996 *Figure 5.6. Posterior distributions of 1000 random draws for the area of the male (n individuals*
2997 *= 9, n hair segments = 169) ellipse (SEA_B) overlapped by the female (n individuals = 12, n hair*
2998 *segments = 108) ellipse, using $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ stable isotope ratios of mustache hair segments.*

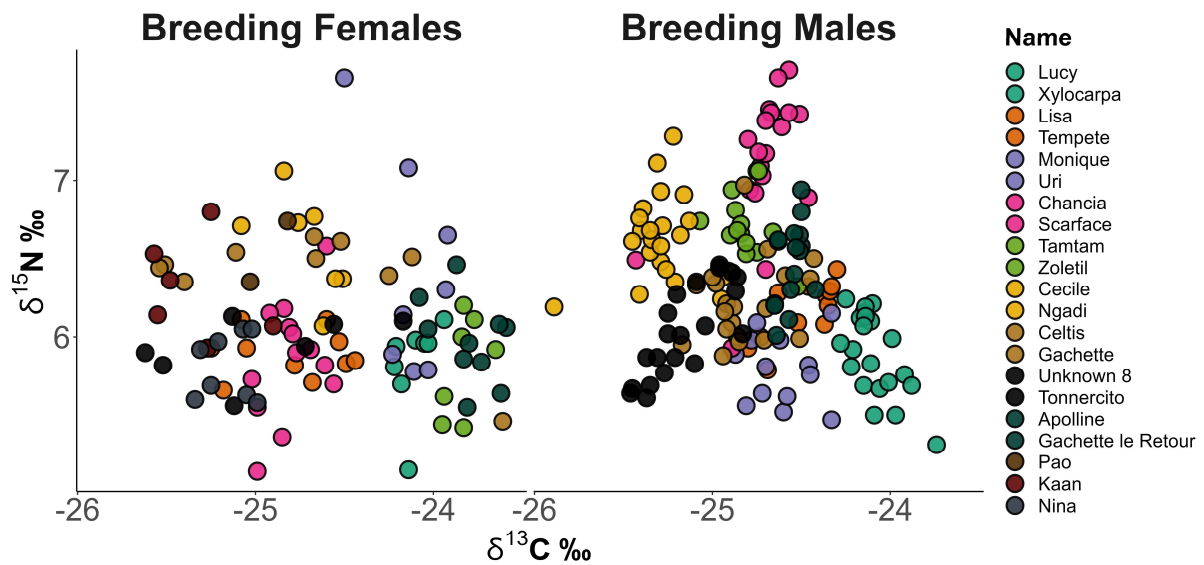
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3000 *5.4.3 Variance Component Analysis*

3001 After finding that female and male niches were similar in size and location, we next examined
3002 whether the niche variance components differed between the sexes. To explore these
3003 differences in the raw data, we first produced biplots displaying the $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ ratios of all
3004 mandrill hair segments. Individual males showed substantial clustering within individuals and
3005 variation between individuals. In contrast, data for individual females was much less clustered
3006 within individuals, suggesting more dietary variation within individuals and less between
3007 individuals (Figure 5.7). To confirm these observations, we examined the variance
3008 components of the $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ niche of each sex, using linear mixed models. Consistent
3009 with the ellipses in Figure 5.4, total $\delta^{13}\text{C}$ niche width (the sum of random intercept and
3010 residual variance) was higher in females than males, whereas total $\delta^{15}\text{N}$ niche width was
3011 higher in males than in females (Figure 5.8).

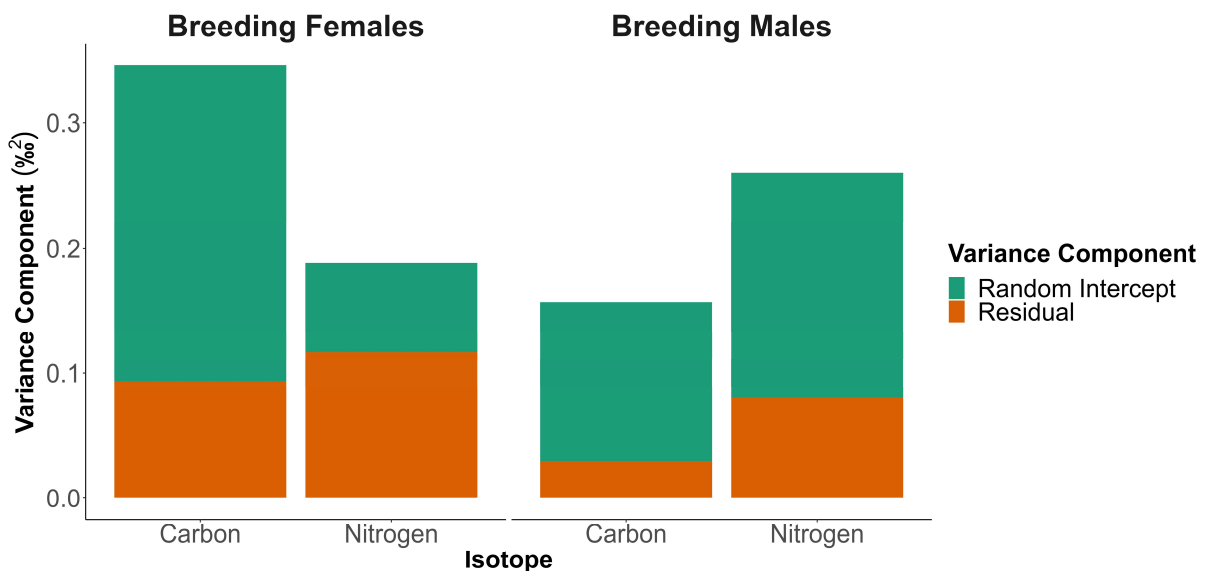
3012 The niche variance components, in terms of random intercept (between individual) and
3013 residual (within individual) variance, also differed between sexes (Figure 5.8). Regarding $\delta^{13}\text{C}$,
3014 female random intercept variance (0.25‰^2) was the largest of all variance components and
3015 approximately double that of male random intercept variance (0.13‰^2). Similarly, female
3016 $\delta^{13}\text{C}$ residual variance (0.096‰^2) was over three times greater than that of males (0.03‰^2).
3017 In contrast to all other niche components, female $\delta^{15}\text{N}$ random intercept variance (0.08‰^2)
3018 was smaller than male random intercept variance (0.18‰^2). In terms of residual $\delta^{15}\text{N}$
3019 variance, female variance (0.12‰^2) again exceeded male variance (0.08‰^2). For both $\delta^{13}\text{C}$
3020 and $\delta^{15}\text{N}$, random intercept variance exceeded residual variance in males. In females, random
3021 intercept variance exceeded residual variance for $\delta^{13}\text{C}$, but for $\delta^{15}\text{N}$ residual variance
3022 exceeded random intercept variance. Summary tables for each variance component model
3023 are available in the Appendix (Tables A5.5; A5.6; A5.7; A5.8).

3024



3025 *Figure 5.7. $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ ratios of all male (n individuals = 9, n hair segments = 169) and female*
 3026 *(n individuals = 12, n hair segments = 108) mandrill hair segments. Points represent the*
 3027 *position of each hair segment in isotopic space. Points colors correspond to individual*
 3028 *mandrills, to illustrate the area of isotopic niche space occupied by each animal.*

3029



3030 *Figure 5.8. Random intercept (between individual) and residual (within individual) $\delta^{13}\text{C}$ and*
 3031 *$\delta^{15}\text{N}$ ratios of male (n individuals = 9, n hair segments = 169) and female (n individuals = 12, n*
 3032 *hair segments = 108) mandrill hair segments, estimated via linear mixed models. Y axis units*
 3033 *are ‰^2 as the niche components constitute variance calculated by squaring the standard*
 3034 *deviation, which are in ‰ units.*

3035

3036 **5.5 Discussion**

3037 We aimed to investigate whether mandrills exhibit niche divergence between males and
3038 females, at the levels of sex and individual, using stable isotope analysis of mandrill mustache
3039 hair segments. Our analysis of mandrill food items suggested that, on average, invertebrates
3040 were ^{15}N -enriched relative to plants and food items in savannas were ^{13}C -enriched relative to
3041 those in forests, though there was substantial variation in each category. Higher isotopic
3042 variation in mandrill tissues thus likely indicates greater diversity in food consumption. When
3043 analyzing the isotopic niches of male and female mandrills, we found both sexes to be in
3044 similar locations within isotopic space, suggesting little dietary niche divergence overall
3045 between males and females. However, variance component analysis revealed sex differences
3046 in niche variation at the individual level, consistent with males and females experiencing
3047 differing levels, or outcomes, of resource competition.

3048

3049 *5.5.1 Isotopic Variation in Mandrill Food Items*

3050 Before interpreting the stable isotope ratios from mandrill hair samples, it was first important
3051 to describe isotopic variation within the study site, to understand the potential sources of
3052 isotopic niche variation for individual mandrills (Roberts *et al.*, 2017). Visualizing the isotope
3053 ratios of mandrill food items in two dimensions (i.e. $\delta^{13}\text{C}$ versus $\delta^{15}\text{N}$), it was clear that using
3054 11 categories did not lead to obvious clustering within plant parts (Figure A5.1). For example,
3055 fruit stable isotope ratios were not obviously distinct from those of leaves, similar to findings
3056 at another site in Gabon (Oelze *et al.*, 2014). Splitting the data into five categories (C3 Plants,
3057 Forest Animals, C4 plants, Savanna Animals and Fungi) appeared to be a more parsimonious
3058 and meaningful way to categorize the samples. Using five categories was preferred to an
3059 intercept only model for predicting $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ of mandrill food items, though substantial
3060 isotopic variation within categories and overlap between categories remained present in the
3061 data (Figure 5.1). These results suggest that inferences about mandrill diets from tissue stable
3062 isotope ratios must be limited to a fairly coarse resolution: greater isotopic $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$

3063 along the length of a hair may indicate greater variation in trophic level or plant species
3064 consumption, respectively.

3065 The mean $\delta^{13}\text{C}$ of forest plants in Lopé National Park was -29.8‰ (95% CI: -30.4 to -29.1‰),
3066 which is similar to Loango National Park, Gabon ($-30.5 \pm 3.5 \text{‰}$, Oelze *et al.*, 2014).
3067 Furthermore, the mean $\delta^{15}\text{N}$ of 3.7‰ (3.2 to 4.1‰) for forest plants in Lopé was close to the
3068 average for Loango plants ($3.3 \pm 2.5\text{‰}$). Thus the isotopic baseline within the home-range of
3069 our mandrill horde is similar to those of other primates studied within Gabon, which could
3070 facilitate inter-site comparisons.

3071 To test whether temporal isotopic baseline shifts occurred in Lopé, we analyzed differences
3072 in forest plant isotope ratios between the 2019 and 2020 long wet seasons and the 2019/20
3073 short dry season. We found that models containing season as an independent variable did
3074 not explain forest plant isotope ratios better than intercept only models, suggesting that no
3075 seasonal or inter-annual shifts in isotopic baseline took place. We also tested whether
3076 simple spatial gradients in $\delta^{15}\text{N}$ or $\delta^{13}\text{C}$ were present within the hordes home-range. Each
3077 model examining the effect of latitude and longitude on the isotope ratios of forest plant
3078 tissues suggested very modest effects, relative to the total isotopic variation we found
3079 between plants (Figure A5.2; A5.3; A5.4; A5.5). Isotopic variation in plants was instead
3080 spatially heterogeneous (Figure 5.2; 5.3). The lack of consistent spatial or temporal variation
3081 of the isotopic baseline in Lopé, compared to isotopic differences between plant species and
3082 food types, means we find no reason to systematically account for time and space in our
3083 analysis of mandrill stable isotope data. Instead, we can be assured that local variation in
3084 isotopic signal is a modest source of error, and may not interfere with the detection of strong
3085 patterns related to dietary variation among or within individual mandrills.

3086

3087 5.5.2 Mandrill Isotopic Niches

3088 The mean stable isotope ratios of mandrills in Lopé were remarkably similar to the results
3089 obtained by Oelze *et al.* (2020) for mandrill hairs in Lékédi National Park, also in Gabon. The
3090 $\delta^{13}\text{C}$ ratios of Lopé males (-24.73‰ , 95% CI: -24.96 to -24.49‰) and females (-24.61‰ , -24.90
3091 to -24.31‰) were ^{15}N -enriched by approximately 0.5‰ compared to Lékédi (mean both

3092 sexes: -25.1‰). The $\delta^{15}\text{N}$ values of Lopé males (6.33‰, 6.05 to 6.61‰) and females (6.10‰,
3093 5.93 to 6.27‰) were even closer to values from Lékédi (mean both sexes: 6.0‰) (Oelze *et*
3094 *al.* 2020). The similarity in stable isotope ratios of hairs between Lopé and Lékédi suggest
3095 broad consistency in the dietary niches of mandrills between these two sites, as has been
3096 found during inter-site comparisons of mandrill diets using observational and fecal contents
3097 data (Nsi Akoué *et al.*, 2017; Hongo *et al.*, 2018; Chapter 3). Adjusting the Lopé mandrill data
3098 by the appropriate hair-diet discrimination factors for mandrills (Chapter 4) placed both male
3099 and female mandrills within the isotopic ranges of Forest Plants, Forest Animals and Fungi in
3100 Lopé (Figure 5.1). The isotope ratios of mandrill mustache hairs in our study are thus
3101 consistent with previous feeding studies showing that mandrills are omnivorous and feed
3102 largely in C3 forest food webs (Nsi Akoué *et al.*, 2017; Hongo *et al.*, 2018; Chapter 3).

3103 The isotopic niches of male and female mandrills were similar in size and location within
3104 isotopic space (Table 5.5; Figure 5.4), as well as similar in size to the isotopic niches of other
3105 primates (Hermsmeyer, 2019; Flores-Escobar *et al.*, 2020). Previous work has shown that the
3106 focal horde are highly generalist omnivores, but with a diet dominated by fruit (Rogers *et*
3107 *al.*, 1996; White, 2007). However, because the number of adult males in the horde is generally
3108 less than 12% (Abernethy *et al.*, 2002), knowledge of the feeding niches of mandrills in Lopé
3109 is mostly limited to females and juveniles (Chapter 3). Our finding that the isotopic niche of
3110 males is 82.5% overlapped by females indicates that, at the level of sex, male and female diets
3111 are alike in our study site, as in other locations (Nsi Akoué *et al.*, 2017). The female isotopic
3112 niche, however, was larger than the male niche across 93.7% of the posterior draws for each
3113 sex's ellipse area, suggesting that females have a slightly wider diet breadth than males.
3114 Feeding competition among individuals is expected to increase with group size (Sterck *et al.*,
3115 1997), potentially leading to higher dietary diversity for groups containing more individuals,
3116 as individuals must forage more opportunistically. The larger isotopic niche of females, which
3117 are permanent horde members, may therefore be a first indication that they experience more
3118 feeding competition than males, which are temporary horde members, and often forage
3119 alone (Abernethy *et al.*, 2002).

3120

3121 5.5.3 Isotopic Niche Variance Components

3122 To further investigate the possibility that male and female mandrills experience differing
3123 levels, or outcomes, of feeding competition, we conducted variance component analysis on
3124 the isotopic niches of each sex. Summing the random intercept (between individual) and
3125 residual (within individual) variance components revealed that total $\delta^{13}\text{C}$ niche width of
3126 females was greater than that of males. Conversely, the total $\delta^{15}\text{N}$ niche width of males was
3127 larger than females. These results mirrored the differing isotopic niche shapes of each sex
3128 (Figure 5.4). However, somewhat conflicting with the high overlap between male and female
3129 isotopic niches, plotting the individual level raw data showed isotopically clustered males and
3130 dispersed females, suggesting niche differences between the sexes at the individual level
3131 (Figure 5.7).

3132 The magnitude of $\delta^{15}\text{N}$ between individual variance (random intercept variance) was the only
3133 variance component found to be larger in males than females. Importantly, the raw data
3134 ranges of $\delta^{15}\text{N}$ were very similar between sexes (Figure 5.7), suggesting that greater trophic
3135 level variance among males is not because large size provides access to higher trophic level
3136 prey, as suggested for other dimorphic species (Louis *et al.*, 2021). High $\delta^{15}\text{N}$ between
3137 individual variance in males may be because they are solitary (Abernethy *et al.*, 2002) and
3138 feed in distinct locations to other males and the horde. If the prey present in these locations
3139 differ in trophic level, individual males would be expected to feed at different trophic levels
3140 for extended periods, leading to larger differences between males in average trophic level,
3141 compared to females.

3142 An alternative explanation could be that males specialize on distinct resources, leading to
3143 consistent dietary differences between individuals. For example, an adult male has been
3144 observed carrying part of a duiker carcass, though whether the mandrill in question made the
3145 kill or scavenged the prey is unknown (Jouventin, 1975). However, this explanation seems less
3146 likely, as it would require all males to have developed their own foraging or hunting
3147 specializations, for the consistent inter-male trophic differences we have found here. The
3148 overlap between male and female niches we have found (Figure 5.5), alongside the broad
3149 similarity of male and female diets observed by others (Nsi Akoué *et al.*, 2017), indicate that
3150 males are generalist foragers, in a similar manner to females. It is thus more probable that

3151 the between individual trophic variance of males is due to opportunistic foraging in separate
3152 locations that contain distinct prey types.

3153 Somewhat unexpectedly, because all females were within the same horde and often feeding
3154 in similar locations, $\delta^{13}\text{C}$ between individual variance was much greater in females than in
3155 males and was the largest of all niche variance components. Thus, plant consumption differed
3156 more between individual females than individual males. The large absolute measure of
3157 between individual $\delta^{13}\text{C}$ variance in females suggests that differences between females in
3158 plant consumption are consistent over time. Mandrill hordes are formed of philopatric
3159 females in a matrilineal and highly nepotistic hierarchy (Setchell, 2000; Charpentier *et al.*,
3160 2022). It is possible that contest competition among matrilines leads to consistent differences
3161 between matrilines, and therefore individual females, in resource access over time. In
3162 another population, high ranking females have been found to spend less time foraging than
3163 low ranking females, implying that resource access is related to hierarchical dominance in
3164 mandrills (Dezeure *et al.*, 2022). Alternatively, the reproductive status of individual females
3165 may have differed, given that mandrill hordes contain females that are ovulating, pregnant,
3166 carrying infants (and therefore lactating), or relatively free of reproductive constraints, in
3167 varying proportions throughout the year (Hongo *et al.*, 2016). Lactation is associated with
3168 increased energy intake in female primates (Cui *et al.*, 2018), meaning $\delta^{13}\text{C}$ divergence
3169 between female mandrills could have been driven by variation in food selection between
3170 females in different reproductive states.

3171 After accounting for between individual random intercept variation, there remained residual
3172 variance in each model, which we took to represent within individual isotopic variation
3173 (Newsome *et al.*, 2009). Regarding $\delta^{13}\text{C}$, female residual variance was over three times larger
3174 than male residual variance, suggests that individual females consumed a wider diversity of
3175 plant foods than individual males. The same outcome followed for $\delta^{15}\text{N}$, for which residual
3176 variance was also greater for females than males. Individual females thus also exhibited more
3177 trophic level diversity than individual males. Considering residual variance in both isotopes
3178 together, these results imply that, over the period of time represented by hair isotope data,
3179 female diets were more diverse than male diets, at the individual level. Thus, females
3180 exhibited larger isotopic niches at the individual level, whereas male niches were relatively
3181 narrow.

3182 The niche width differences between individual males and females may be a consequence of
3183 differing levels of feeding competition experienced by each sex. Feeding competition and
3184 movement between feeding patches is expected to increase with group size (Sterck *et*
3185 *al.*, 1997, Gillespie and Chapman, 2001) and mandrills live in exceptionally large social groups
3186 (Guibinga Mickala *et al.*, 2022). It therefore follows that scramble competition should be high
3187 within mandrill hordes and that hordes must regularly move on to new feeding patches. If
3188 female mandrills are social and males often solitary (Abernethy *et al.*, 2002), females may
3189 experience higher levels of scramble competition and somewhat random resource access.
3190 Food selection by individual females may therefore be relatively opportunistic, which would
3191 explain high within individual variance in females. In contrast, solitary males should
3192 experience lower feeding competition and may be able to remain in the same feeding patches
3193 for extended periods, as a single individual should deplete resources at a much slower rate
3194 than a large group of animals. Lower feeding competition and longer patch residency times
3195 would allow male mandrills to feed more consistently on a narrower range of preferred
3196 resources, producing low within individual feeding variance.

3197

3198 *5.5.4 A possible role for resource competition in the evolution of mandrill sexual dimorphism?*

3199 The sex differences in within- and between-individual niche variation we have documented
3200 here are suggestive of an interplay between sexual dimorphism and feeding ecology in
3201 mandrills. Sex differences may exist in nutritional requirements (Maklakov *et al.*, 2008) and
3202 these have been proposed to explain dietary differences between male and female primates
3203 (Rothman *et al.*, 2008). Experimental evidence also suggests that sex differences in nutritional
3204 requirements may interact with resource competition to produce ecological sexual
3205 dimorphism (De Lisle, 2023). In mandrills, females have been observed to more frequently
3206 consume animal prey than males, perhaps because of higher protein requirements (Nsi Akoué
3207 *et al.*, 2017). However, the high amounts of overlap between male and female isotopic niches
3208 we have found here suggest that any differences in nutritional requirements do not lead to
3209 substantial niche divergence between the sexes in mandrills. It is possible that the large size
3210 of male mandrills counterbalances the energetic costs of pregnancy and lactation to females,
3211 reducing sex differences in nutritional requirements (Key and Ross, 1999). Divergence in male

3212 and female nutritional intake may therefore be too subtle to have played a substantial role in
3213 the evolution of mandrill sexual dimorphism.

3214 A lot of evidence points towards sexual selection as a driver of mandrill dimorphism (Setchell,
3215 2016). Males are 3.4 times larger than females and possess 45mm canines and colorful
3216 ornamentation, with each of these traits influencing male rank attainment or female mate
3217 choice (Setchell, 2005; Leigh *et al.*, 2008; Setchell *et al.*, 2008). Males also exhibit high
3218 reproductive skew, develop more slowly and first reproduce at an older age than females
3219 (Dixson *et al.*, 1993; Setchell *et al.*, 2005; Charpentier *et al.*, 2005). These results are strong
3220 evidence that males invest time and energy into the development of secondary sexual traits
3221 that maximize mate acquisition, as would be expected under sexual selection. However,
3222 sexual dimorphism driven by sexual selection is a common occurrence among primates
3223 (Plavcan, 2001) and the unanswered question is why mandrill dimorphism is so extreme. The
3224 sex differences in dietary generalism we have found here, more specifically, wider feeding
3225 niches in individual females and narrower feeding niches in individual males, may offer some
3226 explanation.

3227 Modeling suggests that resource competition between males and females could, in isolation,
3228 produce sexual dimorphism (Lande, 1980; Slatkin, 1984; Bolnick and Doebeli, 2003). But
3229 perhaps a more likely scenario is that resource competition may act to exaggerate existing
3230 sexual dimorphism produced by sexual selection (De Lisle, 2019). For example, the extent of
3231 carpet python (*Morelia spilota*) size dimorphism changes depending on the range of prey sizes
3232 available in a given location (Pearson *et al.*, 2002). In the case of mandrills, social differences
3233 between males and females may explain why individual males exhibited lower within
3234 individual isotopic variation than females over time. Seasonal influxes of males into mandrill
3235 hordes occur during the breeding season, when the number of tumescent females is highest
3236 (Hongo *et al.*, 2016), suggesting that many males are solitary outside of the breeding period.
3237 Female primates are proposed to live in groups to reduce individual predation risk (Clutton-
3238 Brock and Janson, 2012) and males must therefore forgo this protection when solitary. The
3239 relative dietary consistency exhibited by males suggests that lower feeding competition is one
3240 of the benefits males gain from leaving hordes, because lower individual level feeding
3241 variation implies more regular access to preferred resources. However, leaving hordes means
3242 that males are unable to form long-term social relationships with females. Consequently,

3243 upon joining the group breeding males must rapidly establish dominance over other males
3244 and signal their quality to females, in a dense forest environment (Abernethy *et al.*, 2002).
3245 The intensity of intra- and inter-sexual mating competition may therefore be relatively high
3246 in mandrills, as a result of males rejoining hordes on a seasonal basis. In this manner, the
3247 response of male mandrills to within-group resource competition may have increased sexual
3248 selection for large body size, long canines, and colorful ornamentation and explain why
3249 mandrill dimorphism is so extreme compared to other primate species.

3250 The sex differences in dietary generalism we have found thus give some indication that sexual
3251 selection and resource competition could have interacted to produce extreme sexual
3252 dimorphism in mandrills. However, using only stable isotope data, we cannot rule out that
3253 the dietary sex differences we have documented are simply ecological consequences of size
3254 dimorphism due to sexual selection. For example, sexual selection may have produced large
3255 male mandrills that are able to dominate feeding patches whilst in social groups, which would
3256 also lead to relatively high isotopic / dietary consistency in males. We cannot exclude this
3257 scenario, as we do not know the social status of males prior to capture and hair sampling. One
3258 way to address this is to analyze spatial data obtained from GPS collars fitted at the same
3259 tissues were sampled. Alongside high feeding competition, another ecological cost to the
3260 large size of mandrill hordes is that extensive daily travel is required for horde members to
3261 consume sufficient food (White *et al.*, 2010). If daily energetic expenditure on locomotion is
3262 greater for large males (Key and Ross, 1999), lower travel costs may be an additional
3263 advantage for males that leave mandrill hordes. Analyzing sex differences in home-range use
3264 will reveal when and why males leave mandrill hordes and may give more support to a role
3265 for ecologically-mediated selection in the evolution of mandrill dimorphism (Chapter 6).

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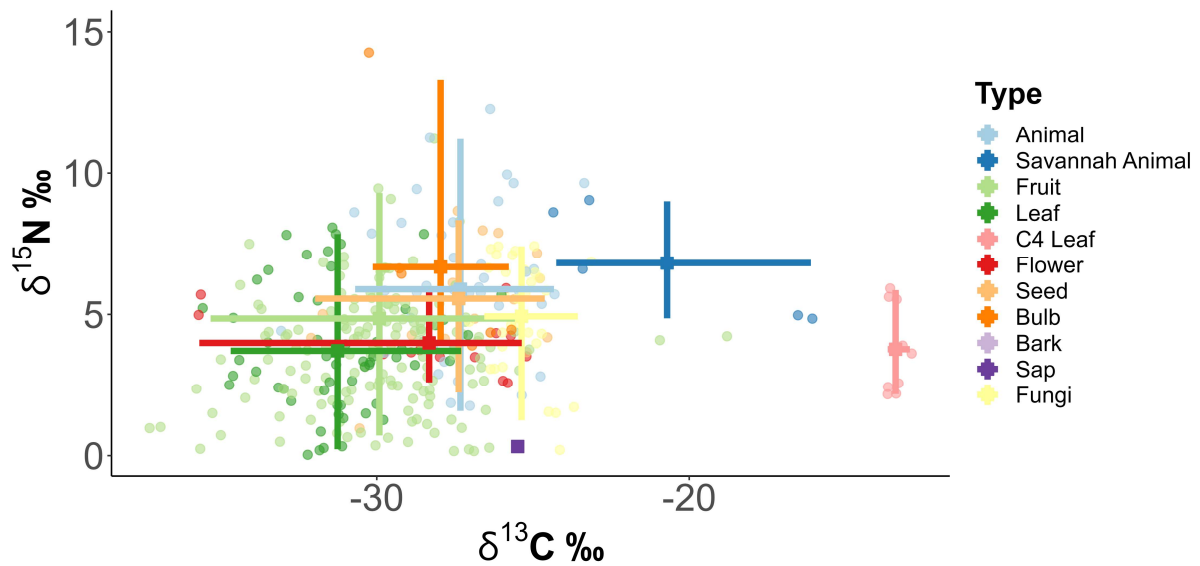
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3268

3269 **5.6 Appendix**

3270 Grouping mandrill food types into 11 categories produced multiple overlapping pair-wise
3271 combinations of categories, for example, fruits, leaves and flowers largely overlap each other
3272 (Figure A5.1).

3273



3274 *Figure A5.1. Isotopic variation between mandrill food items, when grouped into 11 categories*
3275 *(fruit n = 155, leaf n = 56, seed n = 15, flower n = 15, bulb n = 6, sap n = 1, animal n = 43, fungi*
3276 *n = 32, C4 leaf n = 9, savannah animal n = 5). Each data circular point represents at individual*
3277 *food item sample. Square points indicate the mean of each food type and colored lines display*
3278 *the 95% confidence interval around the mean, based on the 2.5% and 97.5% quantiles of the*
3279 *raw data in each category.*

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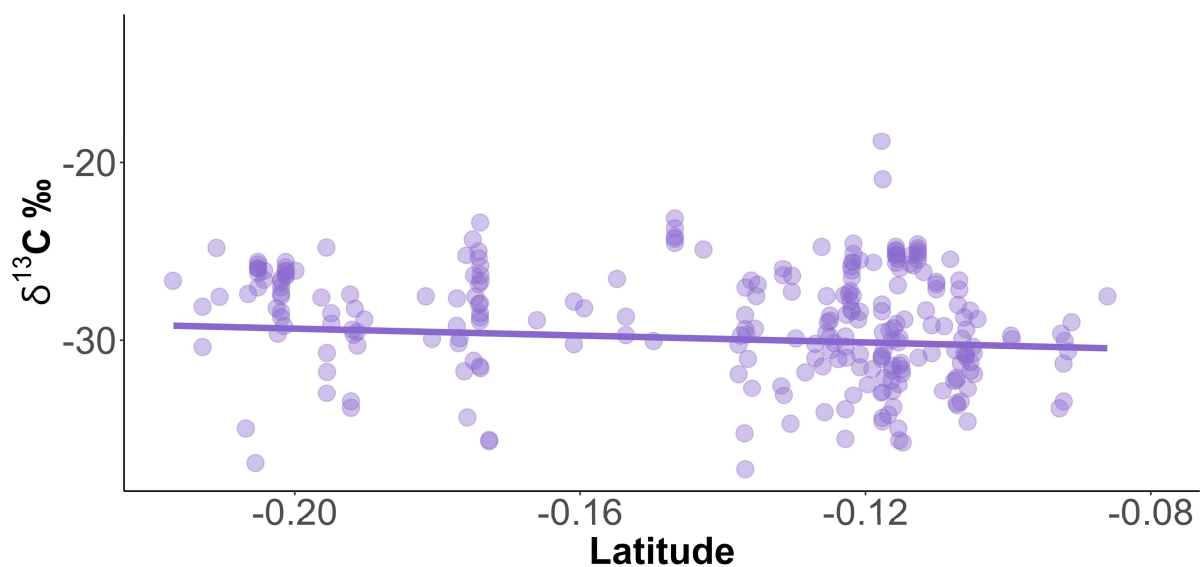
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3285 Analyzing linear relationships between longitude or latitude with $\delta^{13}\text{C}$ or $\delta^{15}\text{N}$, did not
 3286 produce strong linear relationships, relative to the total isotopic variation between samples
 3287 of mandrill food items. Thus, spatial gradients in the stable isotope ratios of mandrill food
 3288 items did not appear to exist in the study site (Figure A5.2; A5.3; A5.4; A5.5).

3289 *Table A5.1. The effect of latitude on mandrill food item $\delta^{13}\text{C}$.*

		Estimate	Standard Error	t
Fixed Effects				
	Intercept	-31.28	0.99	-31.66
	Latitude	-9.64	6.56	-1.47
Random Effects				
Species	Intercept Variance	5.53		
Residual	Observation Variance	3.76		

3290



3291 *Figure A5.2. The relationship between latitude and $\delta^{13}\text{C}$ of mandrill food items. Each point*
 3292 *represents displays the latitude and isotope ratio of an individual food sample ($n = 192$). The*
 3293 *purple line displays the predicted relationship between latitude and food item $\delta^{13}\text{C}$.*

3294

3295

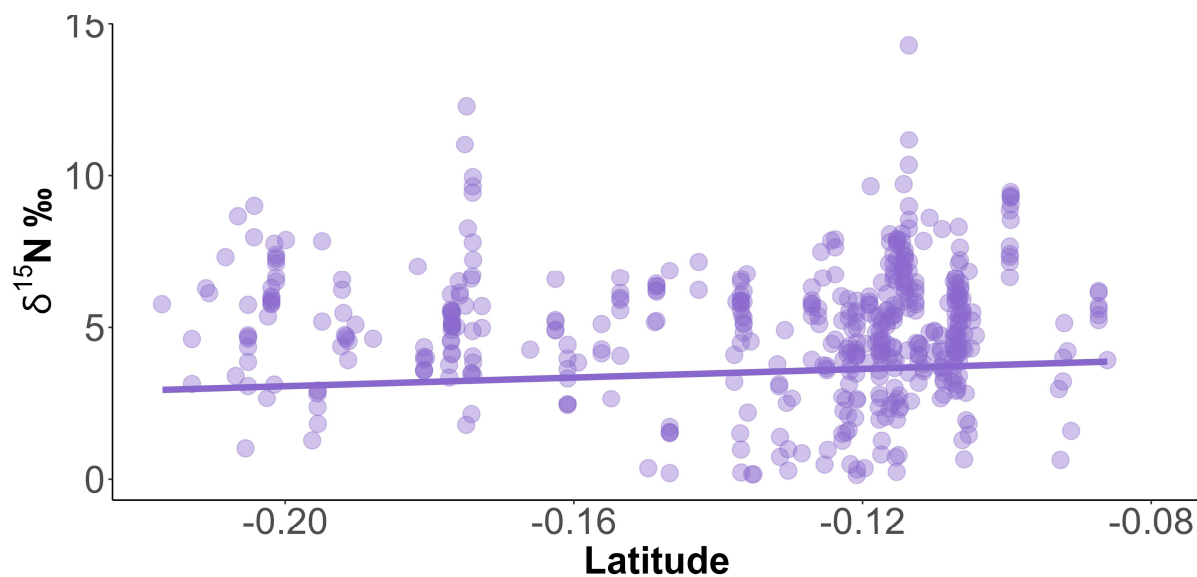
3296

3297

3298 *Table A5.2. The effect of latitude on mandrill food item $\delta^{15}N$.*

		Estimate	Standard Error	t
Fixed Effects				
Intercept		4.49	0.51	8.74
Latitude		7.05	3.15	2.24
Random Effects				
Species	Intercept Variance	2.59		
Residual	Observation Variance	2.19		

3299



3300 *Figure A5.3. The relationship between latitude and $\delta^{15}N$ of mandrill food items. Each point*
 3301 *represents displays the latitude and isotope ratio of an individual food sample ($n = 424$). The*
 3302 *purple line displays the predicted relationship between latitude and food item $\delta^{15}N$.*

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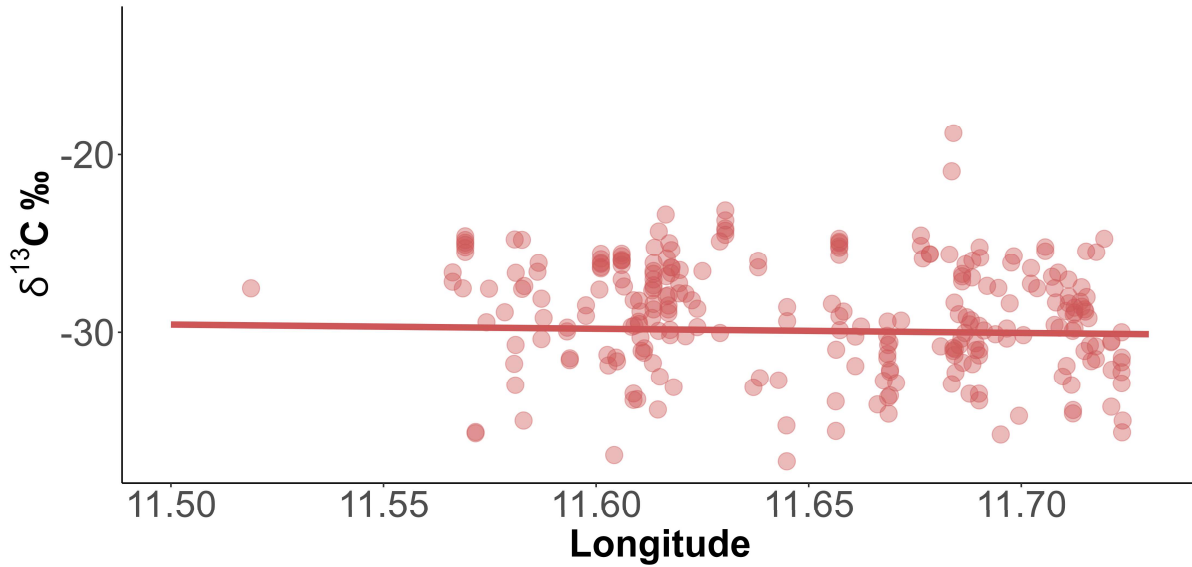
3308

3309

3310 *Table A5.3. The effect of longitude on mandrill food item $\delta^{13}\text{C}$.*

		Estimate	Standard Error	t
Fixed Effects				
	Intercept	-2.57	49.66	-0.05
	Longitude	-2.35	4.26	-0.55
Random Effects				
Species	Intercept Variance	5.52		
Residual	Observation Variance	3.80		

3311



3312 *Figure A5.4. The relationship between longitude and $\delta^{13}\text{C}$ of mandrill food items. Each point*
 3313 *represents displays the latitude and isotope ratio of an individual food sample ($n = 192$). The*
 3314 *red line displays the predicted relationship between longitude and food item $\delta^{13}\text{C}$.*

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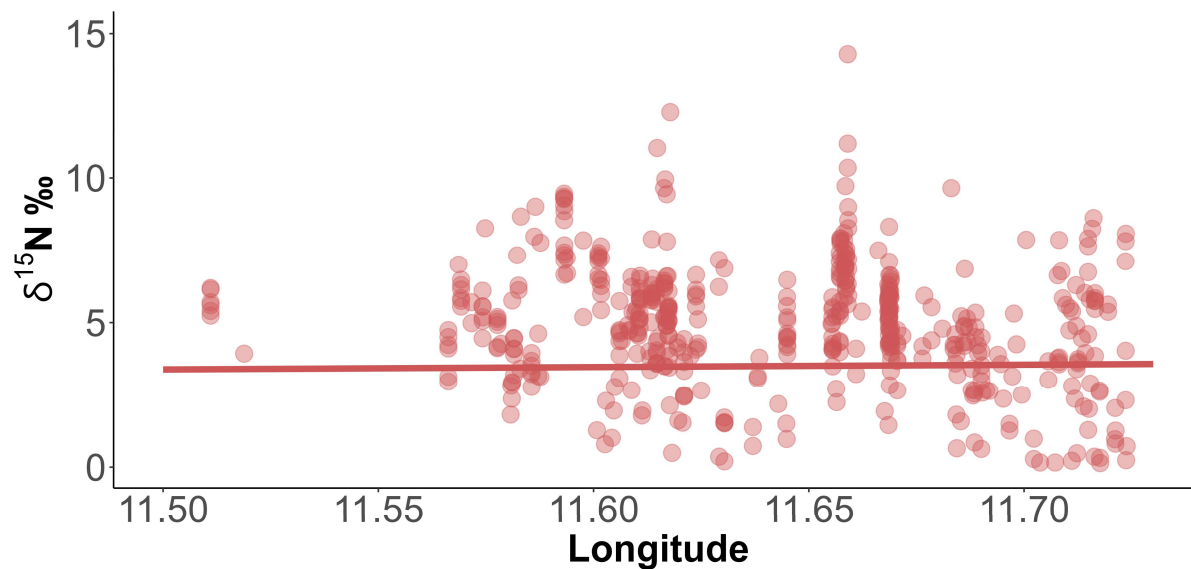
3319

3320

3321

3322 Table A5.4. The effect of longitude on mandrill food item $\delta^{15}\text{N}$.

		Estimate	Standard Error	t
Fixed effects				
Intercept		-5.98	22.21	-0.27
Longitude		0.82	1.90	0.43
Random effects				
Species	Intercept Variance	2.32		
Residual	Observation Variance	2.24		



3323 Figure A5.5. The relationship between longitude and $\delta^{15}\text{N}$ of mandrill food items. Each point
 3324 represents displays the latitude and isotope ratio of an individual food sample ($n = 424$). The
 3325 red line displays the predicted relationship between longitude and food item $\delta^{15}\text{N}$.

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3333 The following four tables display the summaries of linear mixed models used to conduct
 3334 variance component analysis on the isotopic niches of male and female mandrills.

3335 *Table A5.5. Female $\delta^{15}N$ variance components.*

		Estimate	Standard Error	df	z
Fixed Effects					
	Intercept	6.10	0.09	84	71.01
Random Effects					
Name	Intercept Variance	0.08			
Residual	Observation Variance	0.12			

3336 *Table A5.6. Male $\delta^{15}N$ variance components.*

		Estimate	Standard Error	df	z
Fixed Effects					
	Intercept	6.33	0.14	151	44.30
Random Effects					
Name	Intercept Variance	0.18			
Residual	Observation Variance	0.08			

3337 *Table A5.7. Female $\delta^{13}C$ variance components.*

		Estimate	Standard Error	df	z
Fixed Effects					
	Intercept	-24.61	0.15	84	-165.10
Random Effects					
Name	Intercept Variance	0.25			
Residual	Observation Variance	0.096			

3338 *Table A5.8. Male $\delta^{13}C$ variance components.*

		Estimate	Standard Error	df	z
Fixed Effects					
	Intercept	-24.73	0.12	151	-206.47
Random Effects					
Name	Intercept Variance	0.13			
Residual	Observation Variance	0.03			

3339

3340

3341 **Chapter 6: Sexual dimorphism in mandrill (*Mandrillus sphinx*)**
3342 **spatial behavior. A role for travel costs in the evolution of sexual**
3343 **dimorphism?**

3344 Joshua Bauld, Jason Newton, David Lehmann, Katharine Abernethy, Isabel Jones and Luc Bussière
3345 conceived the research. Wild mandrills were collared by David Lehmann, Lisa-Laure Nindiwe Malata,
3346 Brigitte Moussavou, Michel Louis Halbwax, Edmond Dimoto, Arthur Dibambou, Loic Makaga and Josué
3347 Edzang Ndong. Joshua Bauld completed data analysis and wrote the chapter with guidance from Jason
3348 Newton, David Lehmann, Katharine Abernethy, Luc Bussière and Isabel Jones.

3349 **6.1 Abstract**

3350 Larger bodied animals typically occupy wider home-ranges than smaller animals. However,
3351 larger animals incur greater energetic costs of body maintenance, which may trade-off against
3352 the energy available for locomotion and limit the maximum home-range that an animal can
3353 occupy. In species with extreme sexual size dimorphism, this trade-off could place divergent
3354 limitations on the maximum home-range size of each sex. Mandrills are the most sexually
3355 dimorphic primate, in terms of body size, and the larger males often leave social groups
3356 outside of the breeding season, suggesting sex differences in space use. We fitted GPS collars
3357 to 12 female and seven male mandrills, to explore possible sex differences in home-range
3358 size, habitat selection, Brownian motion variance (a proxy for behavioral state changes), daily
3359 travel distance and travel speed. We found home-range size and habitat selection to be
3360 similar between the sexes during the breeding season, but to differ during the non-breeding
3361 season. In addition, males exhibited lower Brownian motion variance (fewer switches
3362 between feeding, resting and travel) during the non-breeding season and a substantial
3363 increase during the breeding season. These results suggest that collared males were more
3364 often solitary during the non-breeding season and more often social during the breeding
3365 season. Males also, on average, traveled shorter distances, at slower speeds, during the non-
3366 breeding season, indicating that males expend less energy on travel when solitary. These
3367 results suggest that energetic trade-offs have heavily impacted the mandrill mating system
3368 and have potentially contributed to the exaggeration of mandrill sexual dimorphism.

3369

3370 6.2 Introduction

3371 Sexual dimorphism is most commonly attributed to sexual selection, but may also be driven
3372 by additional evolutionary mechanisms (Shine, 1989; Janicke and Fromonteil, 2021; Chapter
3373 2). For example, modeling and experimental work suggest that sexual dimorphism could also
3374 evolve via resource competition and ecological character displacement between sexes
3375 (Slatkin, 1984; Bolnick and Doebeli, 2003; De Lisle and Rowe, 2015; Li and Kokko, 2021).
3376 Selection mediated by reproduction and ecology need not be considered in isolation, and
3377 could even interact to exaggerate sexual dimorphism (Lande, 1980; De Lisle, 2019). Indeed,
3378 an interaction between sexual selection and resource competition, but also energetic trade-
3379 offs related to locomotion (flight), may best explain body size dimorphism in giant petrels
3380 (*Macronectes* spp.), the most sexually dimorphic seabirds (González-Solís *et al.*, 2000). Sex
3381 differences in energetic trade-offs may therefore be a useful and understudied factor for
3382 explaining the existence of sexual dimorphism, particularly in cases of extreme size
3383 dimorphism.

3384

3385 6.2.1 Sexual dimorphism and energetic trade-offs

3386 Animal home-range size tends to scale positively with body size (Lindstedt *et al.*, 1986; Tucker
3387 *et al.*, 2014; Tamburello *et al.*, 2015). Furthermore, in sexually dimorphic species, the larger
3388 sex has often been found to occupy more extensive home-ranges (Mysterud *et al.*, 2001;
3389 Ofstad *et al.*, 2016). However, individuals of larger species have higher basal metabolic rates
3390 and therefore greater energetic costs of maintenance, alongside higher energetic costs of
3391 travel (Taylor *et al.*, 1982; White *et al.*, 2019). Trade-offs may therefore exist between the
3392 energy an organism requires for maintenance against the energy available for locomotion,
3393 potentially reducing the home-range area that an animal can occupy (Boratyński, 2020). In
3394 species displaying sexual size dimorphism, it is possible that this trade-off could constrain the
3395 maximum home-range of the larger sex, which could suffer higher maintenance costs to
3396 trade-off against locomotion. Sex differences in energetic trade-offs could suggest an
3397 influence of travel costs on the evolution of sexual dimorphism.

3398 In addition to the energetic costs of movement, home-range use by animals may be
3399 influenced by other factors including food distribution (Di Bitetti, 2001), water availability
3400 (Gehrt and Fritzell, 1998), and the distribution of rest sites and predators (Coleman and Hill,
3401 2014). In group-living animals, such as primates, group size is also relevant, because groups
3402 containing more individuals should deplete feeding patches, such as fruiting tree canopies,
3403 faster than small groups (Chapman, 1988). Larger groups may therefore travel further each
3404 day in search of food (Stevenson and Castellanos, 2001). The result is a positive relationship
3405 between group biomass and home-range size in social animals (Clutton-Brock and Harvey,
3406 1977; Makarieva *et al.*, 2005). Thus, other ecological factors, such as food availability, may
3407 need to be accounted for when investigating a potential influence of energetic trade-offs and
3408 space use on the evolution of sexual dimorphism.

3409

3410 6.2.2 Mandrill Sexual Dimorphism

3411 Among primates, mandrills (*Mandrillus sphinx*) are the most sexually dimorphic in terms of
3412 body size, with males approximately three times heavier than females (male mean = 28.9 Kg,
3413 female mean = 9.8 Kg, in our focal group, David Lehmann *unpublished data*). In addition,
3414 males possess 4.5 cm canines and brightly colored ornamentation on their faces, rumps, and
3415 genitalia (Setchell, 2005; Leigh *et al.*, 2008; Setchell *et al.*, 2008). Though other members of
3416 the Papionine lineage exhibit sexual dimorphism in size and canine length, as well as male
3417 ornamentation (Bergman *et al.*, 2009; Galbany *et al.*, 2015), mandrill dimorphism, in
3418 particular the ornamentation of males, is uniquely extravagant (Darwin, 1871).

3419 Another uncommon aspect of mandrill ecology is the group structure and mating system.
3420 Many Papionines, such as macaques (*Macaca* spp.), live in multi-male, multi-female groups,
3421 with resident males (Higham *et al.*, 2012). Alternatively, many baboons (*Papio* spp.) live in
3422 hierarchical societies, consisting of one-male units, containing a single dominant male and
3423 several females (Swedel, 2002; Goffe *et al.*, 2016). Most male mandrills, however, only join
3424 social groups during the breeding season, when the number of fertile adult females is highest
3425 (Hongo *et al.*, 2016). This seasonal influx of males may intensify sexual selection, and perhaps
3426 explain why mandrills are exceptionally dimorphic, because males must rapidly establish
3427 dominance and demonstrate their quality to females (Abernethy *et al.*, 2002).

3428 It is currently unknown why male mandrills leave social groups outside of the breeding
3429 season, but travel costs relating to their large body size are one possibility (Hongo *et al.*,
3430 2016). Mandrills live in exceptionally large social groups, termed 'hordes', often numbering
3431 600-800 individuals (Abernethy *et al.*, 2002). The biomass of mandrill hordes is thus
3432 exceedingly high (approximately 4.9 tons for our focal horde), leading to the largest known
3433 home-ranges among non-human primates (>180km²; White *et al.*, 2010). The extensive travel
3434 required to sustain several hundred animals may be impossible for male mandrills to maintain
3435 year-round, due to energetic trade-offs between locomotion and maintenance, with a body
3436 mass much higher than that of females.

3437 In this study, we aimed to compare male and female spatial behavior during the breeding and
3438 non-breeding seasons, to address the question of why males seasonally leave social groups.
3439 To do so, we fitted 12 adult female and seven adult male mandrills with GPS collars, to track
3440 space use by individuals of each sex between 2017 and 2022. To investigate whether collared
3441 males were likely to have left the horde, we analyzed sex differences in home-range use,
3442 habitat selection and Brownian motion variance (Horne, *et al.* 2007), during the breeding and
3443 non-breeding seasons. If the home-range sizes and habitat choices of males are more similar
3444 to females during the breeding season and diverge during the non-breeding season, this
3445 would make it unlikely that males exploit the same areas as females outside of the breeding
3446 season and imply that they are not with the horde.

3447 Brownian motion variance (σ^2_m) is a measure of the irregularity of animal movements, based
3448 on the actual and expected locations of an animal (Horne *et al.*, 2007) and can be interpreted
3449 as a proxy measure for the behavioral state of an animal. For example, alternations between
3450 resting or feeding and travel between feeding patches would be expected to produce higher
3451 motion variance. Mandrills are frugivorous (Rogers *et al.*, 1996; Hongo *et al.*, 2018) and
3452 frugivores may exhibit high motion variance when regularly alternating between feeding and
3453 searching for new patches of clumped fruit resources. Larger groups of animals should move
3454 between patches at higher rates than smaller groups or solitary animals (Stevenson and
3455 Castellanos, 2001). Thus, hordes of mandrills would be expected to display higher motion
3456 variance than solitary males. If the motion variance of males is more similar to females during
3457 the breeding season, but lower during the non-breeding season, this would further indicate
3458 that males are not foraging with the horde outside of the breeding season.

3459 We then address the question of why male mandrills may opt to leave hordes outside of the
3460 breeding season, in terms of energetic trade-offs. First, using GPS collar data, we analyze
3461 seasonal changes in the daily travel distance and speed of both sexes. We expected to find
3462 that males travel similar distances to females, at similar speeds, during the breeding season.
3463 Conversely, during the non-breeding season, we expected males to travel shorter distances,
3464 at slower speeds, which would indicate that they expend less energy on travel when there
3465 are few fertile females available. Because GPS data is obtained remotely, we could not be
3466 certain that males were solitary or social (hereafter, “group status”) on days when they were
3467 not observed. We therefore corroborated the data from more recent GPS collars, by analyzing
3468 data collected via direct observations between 1998 and 2004 (White, 2007). These data
3469 constituted daily path lengths of known individuals, including females, social males, and
3470 solitary males. We therefore analyzed the effects of sex and season on the daily path lengths
3471 of individuals whose group status was known, to ascertain whether these data produced
3472 similar results to GPS data. Furthermore, fruit availability data was available for the time
3473 period that focal follows were conducted. We therefore also analyzed the impact of fruit
3474 availability on mandrill daily path lengths, to test whether fruit availability had a greater
3475 impact on mandrill travel than sex, breeding seasonality and group status.

3476

3477 **6.3 Methods**

3478 *6.3.1 Authorization and Ethical Approval*

3479 The collaring of mandrills was approved by the Centre de la Recherche Scientifique et
3480 Technologique (CENAREST), Gabon (AR0036/16) and received ethical approval from the
3481 University of Stirling, UK (AWERB (16/17) 39).

3482

3483 *6.3.2 Study Site and Population*

3484 Our study was conducted in Lopé National Park, Gabon (-0.2N, 11.6E). The park is 90%
3485 covered by continuous forest, with a forest-savanna mosaic covering 10% of the northern and
3486 eastern areas. In the mosaic, savanna constitutes the largest habitat type and is intersected

3487 by strips of gallery forest running along waterways. We considered three major habitat types
3488 relevant for mandrill habitat selection, given the approximate home-range of the focal horde:
3489 continuous forest, gallery forest and savanna. Approximately 1400 mm of rain falls across an
3490 annual cycle, which is split into four seasons: a long wet season (February to May), a long dry
3491 season (June to August), a short wet season (September to November), and a short dry season
3492 (December to January) (White, 1994, Bush *et al.*, 2020). The seasonality in rainfall leads to
3493 temporal variation in fruit production, which impacts the diets of primates within the park
3494 (Tutin *et al.*, 1997; Chapter 3).

3495 We fitted GPS collars to members of a horde that largely occupies the forest-savanna mosaic
3496 (White *et al.*, 2010). The number of individuals in the horde is usually between 600-800, but
3497 can rise to over 1000 (Abernethy *et al.*, 2002). Horde size is so variable because the number
3498 of older males fluctuates seasonally (Abernethy *et al.*, 2002) and because the group exhibits
3499 highly flexible, fission-fusion behavior in response to seasonal changes in fruit production
3500 (White *et al.*, 2010). Furthermore, some males may remain within the horde to pursue
3501 breeding opportunities that arrive outside of the main breeding season (Hongo *et al.*, 2016).
3502 These two factors mean that, when analyzing the spatial behavior of individuals using
3503 remotely sensed GPS collar data, the size of (sub-)groups in which females are present and
3504 male group status cannot be certain. However, given that male group status is related to
3505 seasonal availability of fertile females (Abernethy *et al.*, 2002; Hongo *et al.*, 2016), it is likely
3506 that males are more often solitary in the non-breeding season. Similarly, if horde fissioning is
3507 related to fruit availability, female (sub-)groups are probably often larger during periods of
3508 fruit abundance (White *et al.*, 2010).

3509

3510 *6.3.3 Animal Capture and Collar Fitting*

3511 We fitted collars to mandrills between 01/08/1998 and 15/11/2022. From 1998 to 2008 these
3512 were VHF collars, requiring close (<1km) contact with the animals to verify positions and
3513 group compositions. From 2017 onwards, we used GPS collars and remote tracking alongside
3514 VHF capabilities. Consequently our data covers multiple seasons and years but, crucially, GPS
3515 locations were returned for females and males across 346 and 365 days of the year,

3516 respectively, during the 2017-2022 period. Thus, we were able to analyze the relationship
3517 between day of year and the spatial behavior of both sexes.

3518 Individual animals were anesthetized using Zoletil (Tiletamine and Zolazepam), administered
3519 via a teledart applied with a CO2 Injection Rifle (Daninject JM SP 25). Males received 80 mg
3520 Tiletamine and 80 mg of Zolazepam. Females received 25mg of Tiletamine and 25mg of
3521 Zolazepam. The approximate age in years, age-sex class and weight were recorded for each
3522 animal. A Telonics GPS Iridium collar weighing <3% of animal body mass (a maximum of 5% is
3523 recommended for terrestrial primates) was then fitted to each animal. The collars also carried
3524 a VHF transmitter for tracking individuals and this was tested before fitting each collar. Once
3525 collars were fitted, each animal was allowed to recover within a soft bag, then released and
3526 visually monitored.

3527

3528 *6.3.5 Focal Follows*

3529 To record the daily path length of known individuals, focal follows were carried out between
3530 17/09/1998 and 06/05/2004. Focal follows were conducted by recording the sleeping site of
3531 an individual and re-locating the individual the following morning using VHF radiotelemetry.
3532 A given individual was followed for the entire day and bearings taken via radiotelemetry to
3533 record the location of the individual every 30 minutes. Daily path length was then calculated
3534 using the minimum distances through forest habitat, or using a savanna crossing of <180m
3535 between half-hourly locations and are therefore conservative estimates. For more details on
3536 the calculation of daily path lengths using focal follows see White (2007).

3537

3538 *6.3.6 Analyzing Home-Range Use*

3539 We analyzed GPS location data using the R Programming Language (R Core Team, 2022) and
3540 the 'move' v4.1.1 package (Kranstauber *et al.*, 2022). Because the frequency at which
3541 successful fixes were obtained varied between individuals, we filtered the data to include only
3542 the first fix each day and only fixes obtained before 08:01 am. Our data thus generally
3543 corresponds to the daily 'waking site' of each individual, or a nearby location in the case of

3544 fixes obtained closer to 08:01 am. Using this approach maximized the number of individuals
3545 we were able to include in our analysis, but has the disadvantage of quantifying daily distance
3546 and travel speed using euclidean distances (rather than travel through appropriate habitats)
3547 between waking sites. Mandrill movement may deviate substantially from straight lines
3548 between daily locations (White *et al.*, 2010), but we considered it likely that longer estimated
3549 euclidean distances between waking sites would generally correspond with greater actual
3550 travel distances and speeds. We calculated distances and speeds between daily waking sites
3551 using the inbuilt *distance* and *speed* functions from the 'move' package.

3552 To calculate home-range size and Brownian motion variance, we fitted dynamic Brownian
3553 bridge movement models (Kranstauber *et al.*, 2012), again using the 'move' package. These
3554 models estimate home-ranges based on conditional random walks between locations and are
3555 advantageous over some other home-range estimators because they consider the order and
3556 timing of locations, thus accounting for spatio-temporal autocorrelation (Kranstauber *et al.*,
3557 2012). For applying the dynamic Brownian bridge movement models to daily location data,
3558 we set the margin to 3, the window size to 7, meaning mandrill movement was modeled
3559 based on sequential one week chunks of data (Silva *et al.*, 2020). The location error was set
3560 to 5, thus our models assumed that GPS collar location fixes were accurate to within 5m. The
3561 extent was set to 0.85 and the dimSize to 1000 through an iterative process, to ensure the map
3562 area was large enough to contain each estimated home-range. We estimated home-ranges
3563 using subsets of 30 daily locations for each animal. Using 30 day subsets of data allowed us
3564 to produce monthly home-range estimates that were reproducible between animals for
3565 which we had differing amounts of data. We were also able to produce multiple home-range
3566 estimates for males and females in both the breeding and non-breeding seasons, to analyze
3567 the effects of sex and season on home-range size. For three mandrills (two females, one
3568 male), shorter periods of between 20-30 days were used, to include these individuals in the
3569 analysis. Whether or not these individuals were included in the analysis did not change our
3570 qualitative conclusions. We took the 95% isopleth of each estimated home-range as the
3571 monthly home-range for each animal. The mean day of year for each 30 day period (i.e. the
3572 middle day) was used to allocate each home-range estimate to the breeding or non-breeding
3573 season. In addition to the order and timing of location fixes, dynamic Brownian bridge
3574 movement models utilize an estimate of Brownian motion variance (σ^2m) for estimating

3575 home-ranges. We therefore extracted the motion variance estimates for each animal to
3576 analyze seasonal changes in the frequency of behavioral state changes for males and females.

3577

3578 *6.3.7 Statistical Analysis*

3579 To analyze seasonal changes in monthly home-range size we used a generalized linear mixed
3580 model, with monthly home-range size as the response variable and sex ('male', 'female'),
3581 season ('breeding', 'non-breeding') and their interaction as independent variables. We also
3582 included a random intercept for individual identity, to account for individual differences in
3583 home-range size.

3584 To analyze temporal variation in habitat selection, we first assigned a habitat type to each
3585 daily first GPS collar location, using the *st_within* function from the 'sf' package (Pebesma, *et*
3586 *al.* 2018). We then fitted a multinomial generalized additive mixed model, with the binomial
3587 presence of individuals in each habitat type as the response variable. As fixed factors, we
3588 included a smoother for day of year, habitat type ('gallery forest', 'continuous forest',
3589 'savanna'), sex and their interactions. Day of year was included as a cyclic cubic regression
3590 term to allow the intercepts for early January and late December to align. Random intercepts
3591 for individual identity were also included, to account for differences between individuals in
3592 the relationship between habitat selection and day of year. We did not fit a global effect for
3593 day of year, instead fitting individual curves for each habitat type, each with their own
3594 wiggleness (equivalent to Model I in Pederson *et al.*, 2019). Thus, individual curves were
3595 created describing the separate relationships between day of year and selection of each
3596 habitat type, by each sex.

3597 To analyze temporal changes in the motion variance, daily travel distance and travel speed of
3598 each sex, we also fitted generalized additive models, with each of these variables as a
3599 response. Excepting the lack of a fixed factor for habitat type, these generalized additive
3600 mixed models were fitted following the same procedure as the model examining habitat
3601 selection. For all generalized additive mixed models, restricted maximum likelihood was used
3602 to estimate smoothing parameters.

3603 Next, we used a linear mixed model to investigate the effect of sex and season on the daily
3604 path lengths of known individuals, recorded using focal follows. Daily path length was
3605 included as the response variable and a 'group type' categorical variable describing individual
3606 sex, group status and season ('horde member all seasons', 'social male all seasons', 'solitary
3607 male breeding season', solitary male non-breeding season') used as a fixed effect. 'Horde
3608 member' instead of 'female' was used as a category because one individual was a juvenile
3609 male. However, because juvenile males are also permanent horde members, tracking this
3610 individual was equivalent to tracking adult female horde members. We also included random
3611 intercepts for individual identity to account for individual differences in daily path length.

3612 To investigate whether changes in fruit availability, related to climatic seasonality, was as or
3613 more important than breeding seasonality, we also analyzed the impact of fruit availability on
3614 daily path lengths. For a detailed description of how fruit availability was calculated, refer to
3615 Chapter 3. To analyze the impact of fruit availability, we constructed linear mixed models with
3616 daily path length as a response variable, 'group type' as a categorical fixed factor, monthly
3617 fruit availability as a continuous fixed factor and random intercepts for individuals. We ran
3618 this model with and without an interaction between 'group type' and monthly fruit
3619 availability. We also constructed a linear mixed model with monthly fruit availability as the
3620 sole predictor of daily path length, with random intercepts for individuals, to examine if fruit
3621 availability had an effect of daily path length independent from group type. We compared the
3622 four candidate models for explaining variation in daily path length using AIC scores, taking the
3623 model with the lowest score as the preferred model (Burnham *et al.*, 2011).

3624 We graphically examined the residuals of all models, to ensure that model assumptions were
3625 met. In the analysis of daily path lengths based on focal follows, a gaussian error structure
3626 was suitable, with well distributed residuals. For the habitat selection model, a binomial error
3627 structure produced an appropriate residual distribution. For models examining temporal
3628 changes in monthly home-range size, daily travel distance, travel speed and motion variance,
3629 we found a gaussian error structure to be unsuitable. Each of these measures sometimes
3630 involved low means that approached the lower bound of displacement (0). Because a gamma
3631 distribution is constrained to be positive but can also include continuous (i.e. non-integer)
3632 numbers, it closely matches the properties of the data. We therefore specified a gamma error
3633 structure, which led to well distributed residuals.

3634

3635 6.3.8 Software

3636 All analyses were completed in R v4.1.3 (R Core Team, 2022). Data wrangling, plotting and
3637 production of tables were completed using the packages ‘tidyverse’ 1.3.2 (Wickham *et al.*,
3638 2019), ‘lubridate’ v1.9.0 (Grolemund and Wickham, 2011), ‘ggplot2’ v3.4.0 (Wickham, 2016)
3639 and ‘flextable’ v0.8.3 (Gohel and Skintzos, 2022). Creation and manipulation of shapefiles was
3640 conducted using the ‘sf’ v1.0.9 (Pebesma, 2018) and ‘concaveman’ v1.1.0 (Gombin *et al.*,
3641 2020) packages. Home-range calculation was carried out with the packages ‘move’ v4.1.1
3642 (Kranstauber *et al.*, 2022) and ‘moveViz’ v0.10.5 (Schwalb-Willmann *et al.*, 2020). All statistical
3643 models, checks of residuals and predictions were completed with ‘boot’ v1.3-28.1 (Canty and
3644 Ripley, 2022), ‘mgcv’ v1.8-41 (Wood, 2017), ‘glmmTMB’ v1.1.5 (Brooks *et al.*, 2017) and
3645 ‘DHARMA’ v0.4.6 (Hartig, 2022).

3646

3647 6.4 Results

3648 6.4.1 The effect of sex on space use

3649 To explore whether male and female spatial behavior varied between the breeding and non-
3650 breeding seasons, we examined seasonal changes in monthly home-range size, habitat use
3651 and Brownian motion variance. Males appeared to be more spatially similar to females during
3652 the breeding season and to use somewhat more distinct areas during the non-breeding
3653 season (Figure 6.1). Regarding seasonal changes in monthly home-range size, a generalized
3654 linear mixed model (Table 6.1) suggested that male and female home-ranges were similar in
3655 size during the breeding season. However, the sexes showed distinct changes in the non-
3656 breeding season, with females displaying larger home-ranges and male home-ranges
3657 shrinking (Figure 6.2)

3658

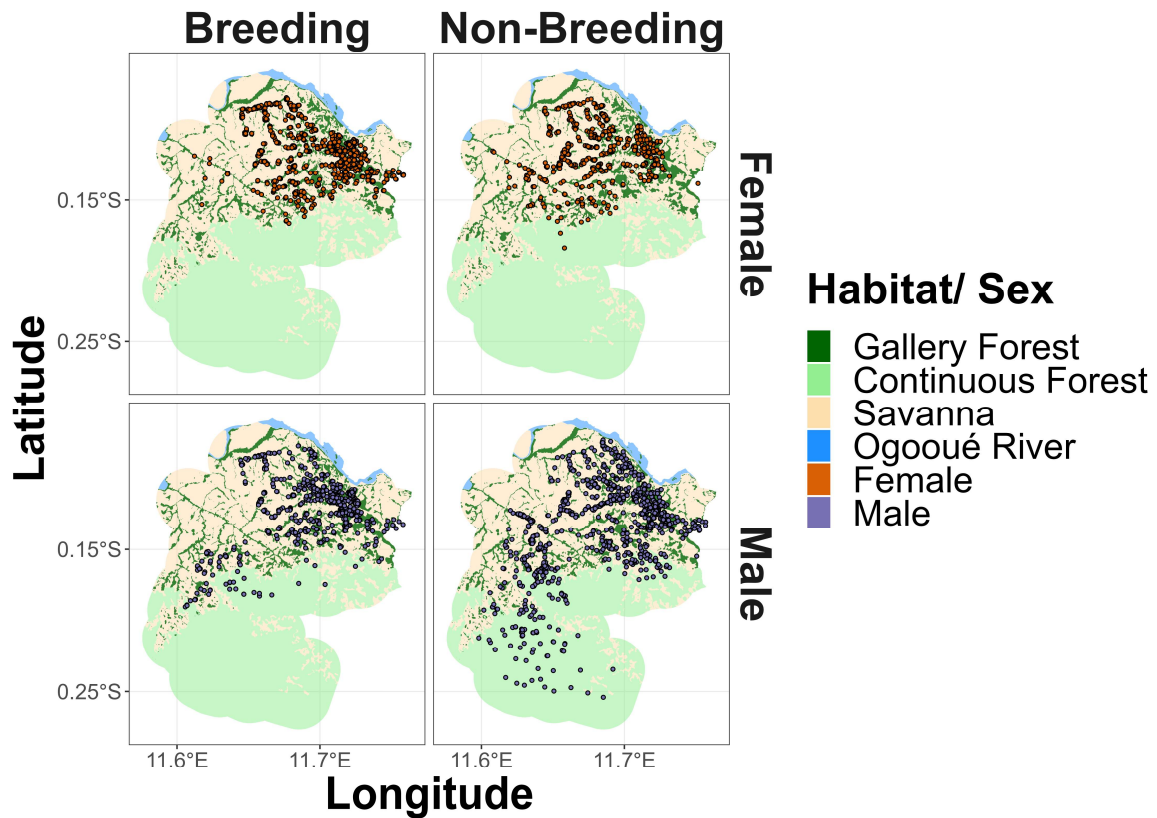
3659 *Table 6.1. Summary of a generalized linear mixed model examining male and female*
3660 *monthly home-range sizes during the breeding and non-breeding seasons.*

	Estimate	Standard Error	z	p
Fixed Effects				
Intercept	0.055	0.007	8.24	<0.0001
Male	0.004	0.010	0.36	0.72
Non-Breeding	-0.018	0.008	-2.26	0.024
Male: Non-Breeding	0.031	0.014	2.27	0.023
Random Effects				
Individual	Intercept Variance	0.000049		

3661

3662 A generalized additive model revealed that male habitat selection was also more similar to
3663 females during the breeding season and more distinct during the non-breeding season (Figure
3664 6.3). Females were present almost exclusively within gallery forests throughout the year, with
3665 some use of continuous forest during the non-breeding season. In contrast, males showed
3666 greater use of continuous forest during the non-breeding season, coupled with a reduction in
3667 gallery forest use. As for home-range size, males were much more similar to females during
3668 the breeding season and largely used gallery forests. The use of savanna habitats by both
3669 sexes was very low and occurred at similar frequencies between the breeding and non-
3670 breeding seasons.

3671 The sexes also differed in temporal patterns of Brownian motion variance. A generalized
3672 additive model indicated female motion variance fluctuates regularly throughout the year,
3673 with peaks in the breeding and non-breeding seasons. Conversely, male motion variance
3674 showed a clear peak only during the breeding season (Figure 6.4).



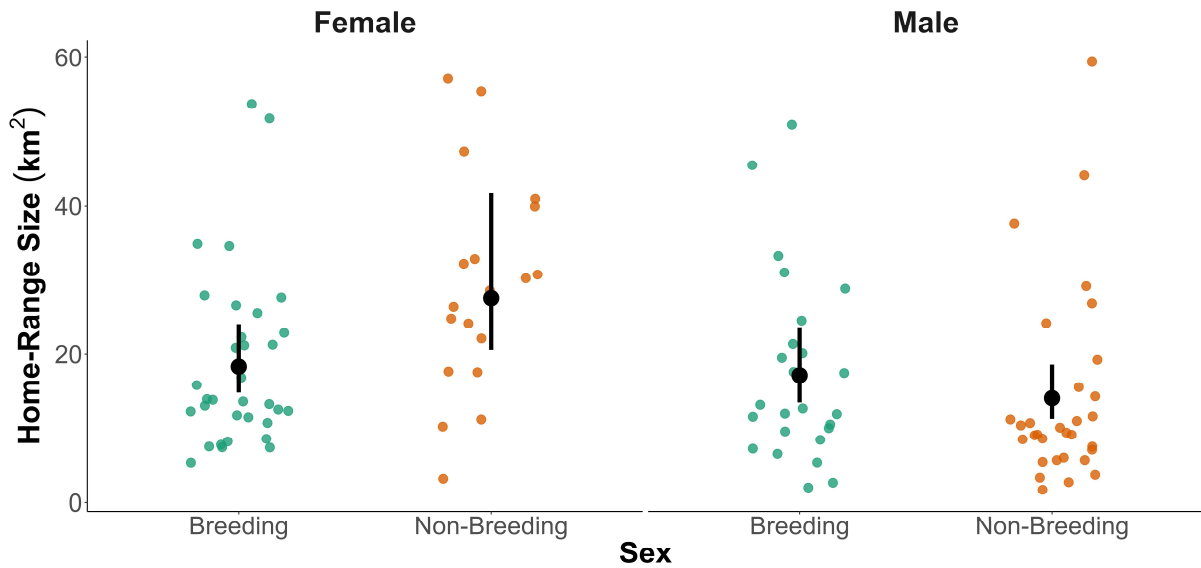
3675 *Figure 6.1. Locations of male (n individual = 7, n locations = 1698) and female (n individual =*
 3676 *12, n locations = 1606) mandrills during the breeding and non-breeding seasons. Each point*
 3677 *indicates the daily waking site of a collared individual. Some points lie outside of the habitat*
 3678 *areas, because on some occasions mandrills crossed the Offoué river (distinct from the Ogooué*
 3679 *to the north of the park), which is the eastern park boundary.*

3680

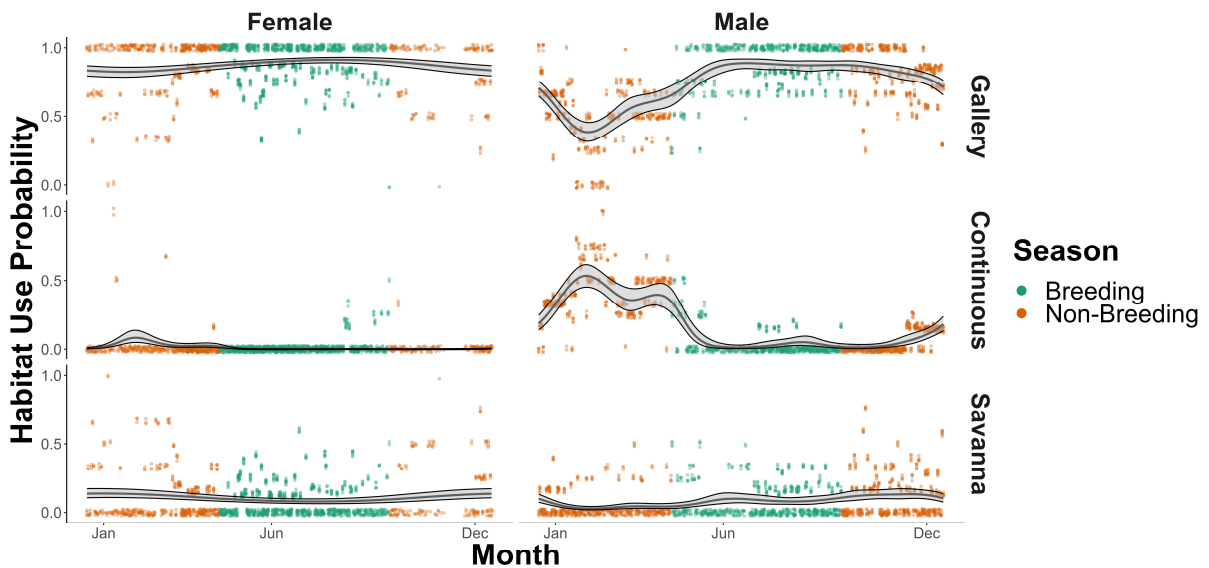
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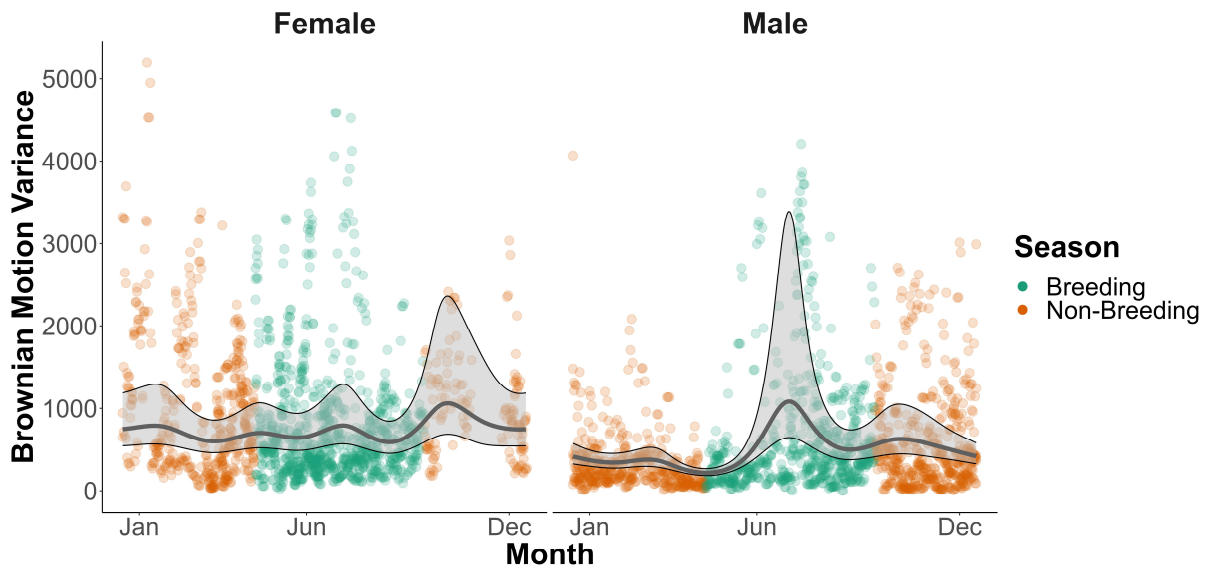
3683



3684 Figure 6.2. Male (n individuals = 7, n home-ranges = 56) and female (n = 12, n home-ranges =
 3685 52) home-range sizes during the breeding and non-breeding seasons. Each colored point
 3686 represents the home-range size of one individual, over one month. Black points represent the
 3687 mean monthly home-range size for a given group and vertical black lines show the 95%
 3688 confidence interval around the means.



3689 Figure 6.3. Seasonal changes in habitat selection by adult male (n individuals = 7, n locations
 3690 = 5094) and adult female (n individuals = 12, n locations = 4818) mandrills. Each point
 3691 represents the first location recorded each day, for a single animal. Grey lines display the
 3692 predicted likelihood of each sex being present in each habitat type on a given day of the year.
 3693 Grey ribbons display the 95% confidence interval around that prediction.



3695 *Figure 6.4. Seasonal changes in Brownian motion variance, a measure of the irregularity of*
 3696 *space use, of adult male ($n = 7$) and adult female ($n = 12$) mandrills. Points represent a single*
 3697 *measure of motion variance for a single animal over a three day period (n females = 1548, n*
 3698 *males = 1669). Grey lines display the predicted motion variance of each sex on each day of the*
 3699 *year. Grey ribbons display the 95% confidence intervals around the predicted motion variance*
 3700 *of each sex.*

3701

3702 6.4.2 The effect of sex on travel

3703 To investigate potential sexes differences in travel, we compared male and female daily travel
 3704 distance and speed during the breeding and non-breeding seasons. Generalized additive
 3705 models suggested that daily travel distance and speed fluctuate aseasonally in females,
 3706 peaking during both the breeding and non-breeding seasons. However, male daily distance
 3707 and speed, as for motion variance, showed major peaks during the breeding season (Figure
 3708 6.5; Figure 6.6).

3709 Our analysis of GPS collars is limited by uncertainty about the social status of collared males
 3710 during the period each individual was tracked. This uncertainty means that some males may
 3711 have been social during the non-breeding season and some solitary during the breeding
 3712 season. Thus, analyzing male spatial behavior in the context of breeding seasons, as opposed

3713 to known group status, is only a proxy measure following general patterns of observed
 3714 mandrill behavior (Abernethy *et al.*, 2002). A major consequence of this is that the seasonal
 3715 changes in male home-range size, motion variance, daily distance and speed are likely to be
 3716 conservative. Had we been able to categorize males as social or solitary, it is likely that
 3717 associations between group status and movement would be stronger than we have found
 3718 here.

3719 Because of uncertainty about the group status of males, we supplemented our analysis of GPS
 3720 data with direct observational data on known individuals, from an earlier time period.
 3721 Analyzing these focal follows using a linear mixed model (Table 6.2) showed that daily path
 3722 lengths were similar between permanent horde members (females and juveniles) and social
 3723 adult males. In contrast, solitary adult males traveled shorter distances than horde members
 3724 and social males during both the breeding and non-breeding seasons, on average (Figure 6.7).
 3725 AIC scores suggested that models including monthly fruit availability as a predictor of mandrill
 3726 daily path lengths were not preferable to models considering only group type (Table 6.3).

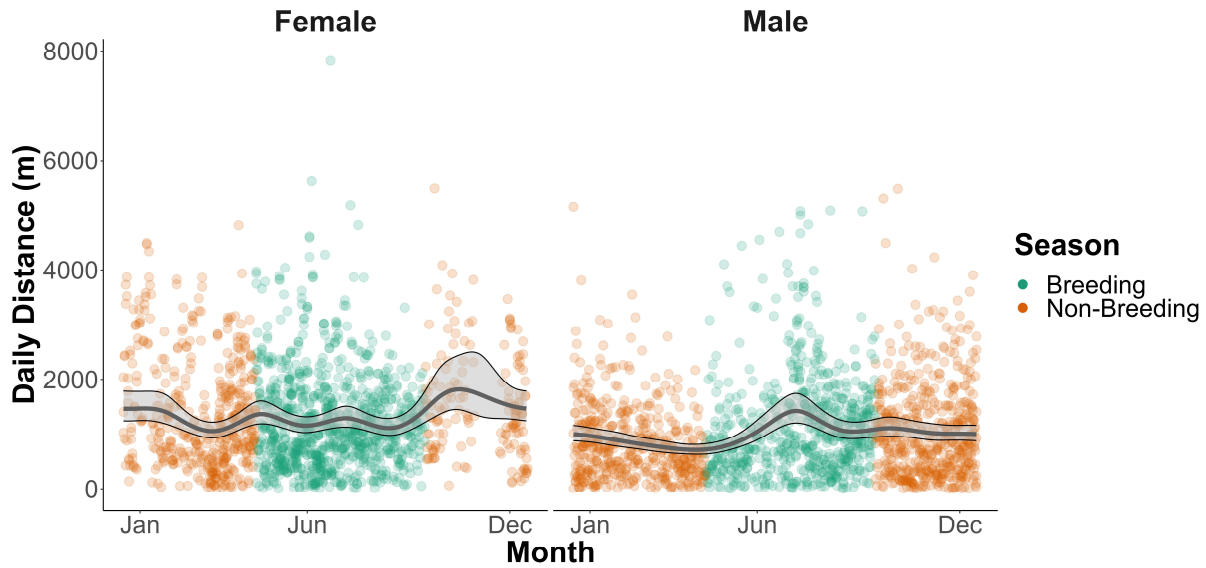
3727 *Table 6.2. Summary of a linear mixed model examining the effect of group type and season*
 3728 *on mandrill daily path lengths.*

	Estimate	Standard Error	<i>z</i>	<i>p</i>
Fixed Effects				
Intercept	3,760.35	314.07	11.97	<0.0001
Social Male	343.05	814.64	0.42	0.674
Solitary Male Breeding Season	-1,837.30	852.27	-2.16	0.031
Solitary Male Non-Breeding Season	-2,294.69	926.49	-2.48	0.013
Random Effects				
Individual	Intercept Variance	1206218		
Residual	Observation Variance	3964811		

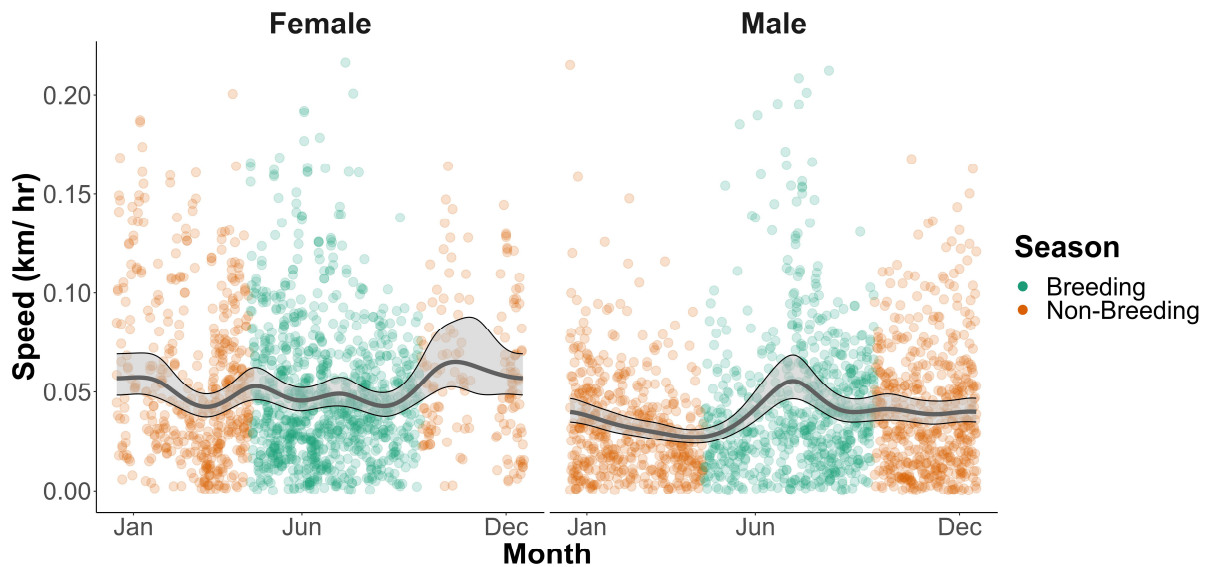
3729

3730 *Table 6.3. AIC scores of models examining the effect of group type and monthly fruit*
 3731 *availability on mandrill daily path length.*

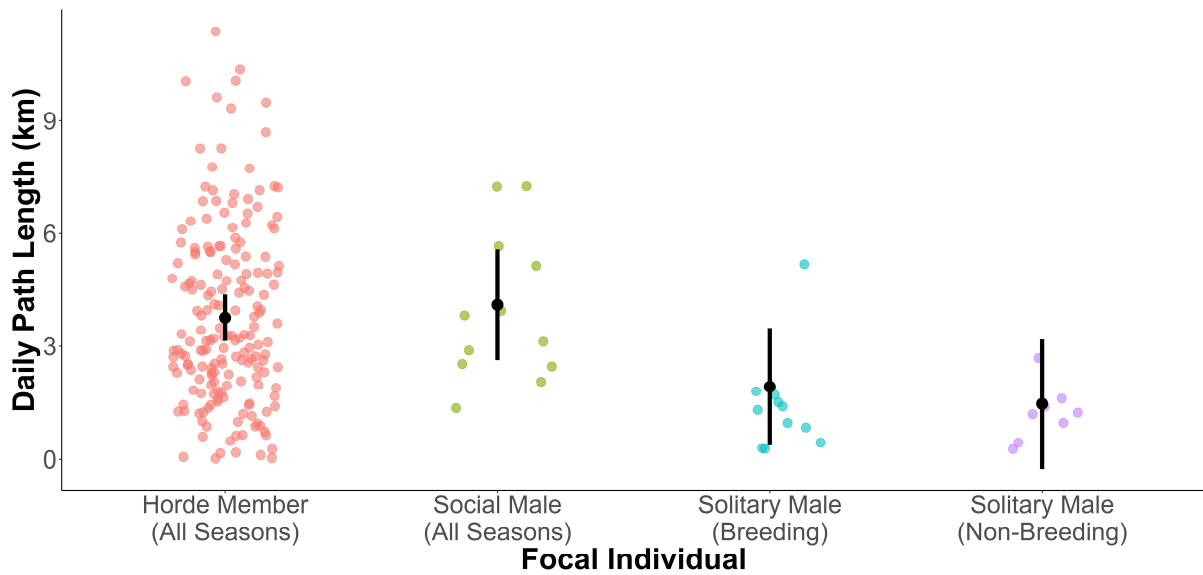
Fixed Factors	Random Factors	Degrees of Freedom	AIC
Group Type	1 Individual	6	3,916.8
Group Type + Fruit Availability	1 Individual	7	3,917.7
Group Type * Fruit Availability	1 Individual	10	3,922.2
Fruit Availability	1 Individual	4	3,920.2



3733 *Figure 6.5. Seasonal changes in daily distance, meaning straight line distance between waking*
 3734 *sites, of adult male ($n = 7$) and adult female ($n = 12$) mandrills. Points represent a single*
 3735 *measure of distance travelled in one day for a single animal (n females = 1594, n males =*
 3736 *1693). Grey lines display the predicted daily distance of each sex on each day of the year. Grey*
 3737 *ribbons display the 95% confidence intervals around the predicted daily distance of each sex.*



3738 *Figure 6.6. Seasonal changes in travel speed, meaning speed between daily waking sites, of*
 3739 *adult male ($n = 7$) and adult female ($n = 12$) mandrills. Points represent a single measure of*
 3740 *travel speed over one day for a single animal (n females = 1594, n males = 1693). Grey lines*
 3741 *display the predicted speeds of each sex on each day of the year. Grey ribbons display the 95%*
 3742 *confidence intervals around the predicted speed of each sex.*



3744 *Figure 6.7. Daily path length differences between horde members (n individuals = 23, n daily*
 3745 *paths = 184), social males (n individuals = 5, n daily paths = 12), and solitary males the*
 3746 *breeding (n individuals = 5, n daily paths = 11) and non-breeding seasons (n individuals = 4, n*
 3747 *daily paths = 8). Colored points represent a single measure of daily path length, obtained from*
 3748 *a direct focal follow of an individual mandrill from sleeping site to sleeping site. Black points*
 3749 *represent the mean daily path length of each group type and black vertical lines display the*
 3750 *95% confidence interval around each mean.*

3751

3752 **6.5 Discussion**

3753 Our study is the first to explore sex differences in mandrill spatial behavior using GPS collars,
 3754 allowing males to be tracked even when solitary. We reveal clear differences between males
 3755 and females in patterns of home-range size, habitat selection and space use during the non-
 3756 breeding season. Conversely, in the breeding season, male spatial behavior is similar to
 3757 females. During the non-breeding season, males, on average, exploited smaller home-ranges
 3758 than females, used distinct habitats, and displayed less variable movement. These results
 3759 imply that males in our study were not permanent members of the focal horde and further
 3760 evidence that male mandrill group presence varies depending on the availability of fertile
 3761 females. Furthermore, during the non-breeding season, males exhibited shorter average

3762 travel distances and lower speeds, suggesting that males expend less energy on travel during
3763 this period. These findings are a first indication that travel costs may influence the mandrill
3764 mating system and that energetic trade-offs may be important to the evolution of extreme
3765 sexual dimorphism in mandrills.

3766

3767 *6.5.1 The effect of sex on seasonal space use.*

3768 Information on unhabituated mandrill space use, and particularly by solitary males, has
3769 historically proven difficult to obtain because of the dense forests in which mandrills are
3770 found (Hongo *et al.*, 2016). We examined sex differences in seasonal habitat selection,
3771 monthly home-range size, and Brownian motion variance, using GPS collars, which allowed
3772 space use data to be obtained remotely, without carrying out focal follows (Figure 6.1).

3773 Male habitat selection differed from females on a seasonal basis. Throughout the year
3774 females were found largely in gallery forests, with only a small amount of continuous forest
3775 use in the non-breeding season. It is relevant to note that in the mosaic structure, some forest
3776 types may be briefly used simply as a travel corridor through a predominantly savanna area,
3777 rather than selected as a preferred choice of forest habitat. Conversely, males increased their
3778 use of continuous forest during the non-breeding season and, like females, almost exclusively
3779 used gallery forest in the breeding season. Much higher use of continuous forest by males is
3780 strong evidence that they, at least sometimes, were not travelling with the horde. It is possible
3781 that these males were alongside a sub-group of females containing no collared individuals.
3782 But the predicted use of continuous forest by females being minimal for the entire year makes
3783 the use of this habitat by uncollared females seem less likely.

3784 The generalized linear mixed model examining the relationship between sex, season and
3785 monthly home-range size suggested that male and female home-ranges are similar in size
3786 during the breeding season. However, as predicted, male home-ranges shrank during the non-
3787 breeding season, whereas female home-ranges became larger. Primate home-ranges are
3788 expected to increase linearly with group biomass, and mandrill hordes conform to this
3789 expectation (White *et al.*, 2010). So, the fact that male home-ranges were smaller during the

3790 non-breeding season suggests that they were in groups of smaller biomass than females
3791 during this time, if they were not solitary.

3792 The horde's fission-fusion behavior leads to subgroups containing fewer individuals and hence
3793 lower biomass (White, 2007). But fission-fusion by the horde cannot explain relatively small
3794 male home-ranges, otherwise the effect of season on male home-ranges would be
3795 indistinguishable from that of females. Males could leave the horde and form bachelor
3796 groups, with lower biomass and home-range requirements. However, large bachelor groups
3797 of multiple mature males have never been observed in mandrills, though a single solitary male
3798 may associate with up to three younger males (Abernethy *et al.*, 2002; SEGC, *unpublished*
3799 *data*). Thus, the smaller average home-ranges of males during the non-breeding season
3800 strongly suggests that at least some males were solitary, or associated with <5 others, and so
3801 required less space than social females.

3802 We also found sex differences in seasonal patterns of Brownian motion variance. A
3803 generalized additive model revealed that female Brownian motion variance fluctuated
3804 throughout the year, showing multiple peaks in the breeding and non-breeding seasons.
3805 However, males showed a clear peak in motion variance during the breeding season. Higher
3806 Brownian motion variance indicates fewer regular paths and increased movement, whereas
3807 lower values indicate more regular paths and lower movement (Silva *et al.*, 2018). Mandrills
3808 are omnivores, with a diet dominated by fruit (Nsi Akoué *et al.*, 2017; Chapter 3), and would
3809 therefore be expected to travel most when moving between fruiting trees (Goldsmith *et al.*,
3810 1999). As fruit patches (tree canopies) are of similar sizes larger primate groups should move
3811 more regularly between feeding patches than smaller groups (Di Bitetti, 2001). Conversely,
3812 solitary males would need to travel in search of new feeding patches less regularly than
3813 females. Longer patch residency times would result in less frequent behavioral state changes
3814 between feeding and travelling, explaining lower male motion variance in the non-breeding
3815 season. Thus, the fact that males exhibited an obvious peak in motion variance during the
3816 breeding season is further evidence that they were within larger groups during this time, and
3817 either solitary or in smaller groups during the non-breeding season.

3818 The similarity of male habitat selection, home-range size, and motion variance to females
3819 during the breeding season, and dissimilarity during the non-breeding season, all strongly

3820 suggest that collared males were only temporary horde members. Furthermore, the
3821 divergent temporal patterns exhibited by each sex indicates that their space use is governed
3822 by different factors. Female mandrill space use is known to be driven by food availability
3823 (White, 2007; Hongo *et al.*, 2018). The home-ranges of female mandrills were larger during
3824 the non-breeding season, which is also the time of year when fruit availability is highest (Tutin
3825 *et al.*, 1997). During periods of fruit abundance, the horde tends to remain as a larger unit,
3826 whereas during periods of scarcity, the horde may fission into sub-groups (White, 2007).
3827 Larger sub-groups likely explain why females used greater areas during the non-breeding
3828 season. The fact that males displayed the opposite pattern, with larger home-ranges in the
3829 breeding season, implies that male home-range use is not primarily driven by fruit availability
3830 and fission-fusion behavior. Our results, alongside those of prior studies (Abernethy *et al.*,
3831 2002; Hongo *et al.*, 2016), indicate that males join the horde during the breeding season. This
3832 is most clear when considering that males were often located in the continuous forest during
3833 the non-breeding season, a habitat rarely exploited by females. Male space use therefore
3834 appears to be driven by the seasonal availability of fertile females, as opposed to fruit
3835 availability (Abernethy *et al.*, 2002; Brockmeyer *et al.*, 2015; Hongo *et al.*, 2016).

3836

3837 *6.5.2 Males expend less energy on travel during the non-breeding season.*

3838 Group membership provides male primates with a number of benefits, including access to
3839 mating partners (Plavcan, 2001; Cassini, 2020b) and enhanced predator detection (van Shaik
3840 *et al.*, 1983; Matsumoto-Oda *et al.*, 2018). So, if male mandrills are leaving the horde, the loss
3841 of female access and predator detection must be offset by the benefits of being solitary. Two
3842 generalized additive models indicated that the average daily travel distance and speed of
3843 males was lower during the non-breeding season and increased during the breeding season.
3844 Energetic expenditure on movement rises with the distance an animal travels (Marcus
3845 Rowcliffe *et al.*, 2012) and with the speed at which an animal moves (Taylor *et al.*, 1982). Thus,
3846 the lower average daily distances and speeds that male mandrills exhibited during the non-
3847 breeding season suggest that they expend less energy on travel during this period.

3848 To confirm that the group status of males is the most likely determinant of their spatial
3849 behavior, we analyzed daily path length data for individuals whose social status was known.

3850 A linear mixed model showed that the daily path lengths of social males were similar to
3851 permanent members (females and juveniles) of the horde. In contrast, solitary males traveled
3852 significantly shorter distances than horde members. Reduced travel by solitary males was
3853 consistent whether individuals were tracked in the breeding or non-breeding season.

3854 Including monthly fruit availability in models explaining daily path lengths did not improve the
3855 predictive power of the models (Table 6.3). A limited effect of fruit availability on the daily
3856 path length of horde members was surprising, given that female space use is influenced by
3857 fruit availability in another study site (Hongo *et al.*, 2018; 2022). It could be that our measure
3858 of monthly fruit availability, corresponding to the entire study area, was too coarse in
3859 resolution to associate strongly with spatial behavior that occurs on more local scales.
3860 However, the same fruit availability measure was associated with fruit consumption by the
3861 horde, during the same time period that the focal follows were conducted, meaning our
3862 measure of fruit availability is associated with the hordes behavior (Chapter 3). Alternatively,
3863 flexible social behavior by the horde may explain why fruit availability is a poor predictor of
3864 daily path length. While the actual location of mandrill hordes is determined by the
3865 distribution of fruit, fission-fusion behavior might allow females to moderate the distances
3866 they travel by forming smaller sub-groups. Thus, flexible social behavior could allow hordes
3867 to mitigate lower fruit availability, by travelling shorter distances in smaller groups, that
3868 deplete patches more slowly. Low explanatory power of fruit availability for male movement,
3869 relative to group status, implies that changes in male social behavior is the primary
3870 determinant of male daily path length. This analysis of known individuals thus supports our
3871 conclusion that male spatial behavior is driven primarily by group membership, as opposed
3872 to other factors, such as food availability.

3873 Why should male mandrills need to forgo the benefits of group life to expend less energy on
3874 travel, when the relative cost of traveling further or faster tends to be lower for larger animals
3875 (Taylor *et al.*, 1982)? This question is especially pertinent given that, after controlling for
3876 group size, larger bodied primates tend to have larger home-ranges (Milton and May, 1976).
3877 The home-ranges of mandrills are exceptionally large compared to other primates, due to the
3878 total biomass of mandrill hordes (White *et al.*, 2010). It may be that the home-ranges of
3879 mandrill hordes are so large, and the size differences between males and females so extreme,

3880 that the energetic costs of maintenance for males are too high to remain with the horde all
3881 year.

3882 This idea is supported by investigations showing that larger animals exhibit higher basal
3883 metabolic rates (White *et al.*, 2019) and that these energetic costs may trade-off against the
3884 maximum home-range size that an animal can occupy (Boratyński, 2020). Furthermore, when
3885 male mandrills are courting females, they spend a substantial amount of time mate guarding
3886 and fighting other males (Setchell *et al.*, 2005; Setchell *et al.*, 2016). Stable isotope analysis
3887 suggests that time devoted to mate guarding produces nutritional stress in males during the
3888 breeding season (Oelze *et al.*, 2020). The most dominant males also experience greater
3889 oxidative stress than lower ranking males, or females, during the breeding season (Beaulieu
3890 *et al.*, 2014). Moreover, in semi-free ranging groups, fecal glucocorticoid metabolites and
3891 parasite load are elevated in dominant males (Setchell *et al.*, 2010). As a result, adult males
3892 may end the breeding season in poor body condition which, coupled with higher baseline
3893 costs of body maintenance, renders them unable to continue to travel alongside the horde.
3894 Abandoning the horde and expending less energy on travel may thus form part of a strategy
3895 to recoup sufficient body condition to compete for females in the following breeding season.

3896

3897 *6.5.3 A role for energetic trade-offs in the evolution of mandrill dimorphism?*

3898 We found that, outside of the breeding season, males selected different habitats, used
3899 smaller home-ranges, and displayed lower motion variance than females, all of which
3900 supports previous studies showing that most males leave mandrill hordes during the non-
3901 breeding season (Abernethy *et al.*, 2002; Hongo *et al.*, 2016). The lower average travel
3902 distance and speed of males we found in the non-breeding season suggest that lower travel
3903 costs may be one of the reasons male mandrills leave social groups. Sex differences in
3904 locomotion costs are not commonly invoked to explain the evolution of sexual dimorphism in
3905 primates, but have been proposed in other species (González-Solís *et al.*, 2000).

3906 Sexual dimorphism is instead generally explained by sexual selection (Andersson, 1994).
3907 There is strong evidence for sexual selection in mandrills, in which larger body size, longer
3908 canines and brighter coloration are all positively associated to male rank attainment, female

3909 preference and reproductive success (Setchell, 2005; Leigh *et al.*, 2008; Setchell *et al.*, 2008).
3910 However, if labile traits such as body size and coloration are related to male reproductive
3911 success, it is possible that the energy balance of individuals is also a relevant factor.
3912 Individuals are expected to devote energy towards reproductive effort only once their basic
3913 metabolic requirements have been met (Brown *et al.*, 1993). In social animals, an individuals'
3914 ability to meet these requirements may be impacted by within-group feeding competition
3915 and the costs of traveling with a group (Markham *et al.*, 2017).

3916 Theory suggests that resource competition can act to enhance sexual dimorphism produced
3917 by sexual selection (De Lisle, 2019). Analyzing the same individuals as in this study, Chapter 5
3918 showed that male mandrills are less impacted than females by resource competition.
3919 However, the stable isotope analysis used in Chapter 5 could not determine if males avoid
3920 resource competition by leaving the horde, or if their large body size allows males to
3921 dominate feeding patches within the horde. The spatial analyses presented here suggest that
3922 the same males often leave the horde outside of the breeding season, suggesting that the
3923 avoidance of feeding competition is what allowed males to exhibit more dietary consistency
3924 than females (Chapter 5).

3925 Taken together, the spatial and isotopic analyses of the same males imply that these
3926 individuals experience reduced feeding competition and travel costs, as a result of leaving the
3927 horde. Solitary males should therefore be able to maintain a more positive energy balance
3928 than social males, as result of greater food intake and reduced energy expenditure. Coupled
3929 with Oelze *et al.*'s (2020) finding of nutritional stress in actively breeding males, our isotopic
3930 and spatial analyses suggest that male energy balance has impacted the mandrill mating
3931 system, by driving males to seasonally leave the horde.

3932 The temporary nature of male group presence has two impacts on the mandrill mating
3933 system. Firstly, the dominance hierarchies among males are unstable (Smit *et al.*, 2022),
3934 meaning males must compete for dominance upon entering the horde. Secondly, males are
3935 unable to form long-term relationships with females (Abernethy *et al.*, 2002) and must
3936 therefore communicate their quality to females through other means, such as ornaments.
3937 Intra- and inter-sexual breeding competition, the two components of pre-copulatory sexual
3938 selection, are thus likely to be exaggerated by males entering and leaving the horde,

3939 compared to primate species that form more stable societies. Our present results, and those
3940 of Chapter 5, suggest resource competition and energetic trade-offs compel males to leave
3941 mandrill hordes. These results are supportive of the possibility that ecological factors,
3942 unrelated to reproduction, have impacted the mandrill mating system to produce stronger
3943 sexual selection and exaggerate sexual dimorphism in mandrills.

3944

Chapter 7: General Discussion

3945

3946

3947 **7.1 Background**

3948 Determining the origins of sexual dimorphism has been a major focus of evolutionary ecology,
3949 yet a complete explanation of how sexual dimorphism evolves remains elusive (Shine, 1989;
3950 Wyman *et al.*, 2013). The most widely accepted explanation is that anisogamy leads to sex
3951 differences in optimal reproductive strategy (Parker and Pizzari, 2015). These sex differences
3952 result in sexual or fecundity selection that may favor sexual dimorphism in morphological or
3953 behavioral traits that optimize the reproductive success of each sex (Lindenfors *et al.*, 2002;
3954 Mitoyen *et al.*, 2019). However, some sexually dimorphic traits, such as body size, may not
3955 only influence the reproductive success of males and females, but also their respective
3956 ecologies (Selander, 1966; De Lisle and Rowe, 2015). The existence of ecologically relevant
3957 sexual dimorphisms suggests that alternative evolutionary mechanisms, mediated by ecology
3958 instead of reproduction, could also produce sexual dimorphism (Slatkin, 1984; De Lisle, 2019).
3959 For example, sex differences in nutritional optima, energetic trade-offs, or response to
3960 resource competition could favor the evolution of ecological sexual dimorphisms (González-
3961 Solís *et al.*, 2000; De Lisle, 2023). The notion of sexual dimorphism driven by ecologically
3962 mediated selection remains controversial, however, and concrete examples are rare (Li and
3963 Kokko, 2021).

3964 This thesis adds to the body of evidence suggesting that selection mediated by ecology could
3965 contribute to the evolution of sexual dimorphism, both generally and in mandrills specifically.
3966 In the next paragraphs I summarize the sequence of questions and approaches before
3967 discussing each of the chapters in more detail. We first used a meta-analytic approach to
3968 model cross-species associations between sexual dimorphism and ecological sex differences,
3969 measured using stable isotopes. This meta-analysis included data from 158 studies, across
3970 163 vertebrate species with a global distribution, to investigate whether a relationship
3971 between sexual dimorphism and ecology could be more general than currently appreciated
3972 (Chapter 2; Bauld *et al.*, 2022).

3973 We next analyzed data from eight-years of mandrill fecal collections and fruiting tree
3974 phenology, to investigate how mandrill diets vary in response to seasonal variation in fruit
3975 production. The multi-year duration of the datasets allowed robust conclusions to be drawn
3976 about the feeding niche of mandrills, that provided a strong framework within which to
3977 interpret the stable isotope analysis of mandrill tissues (Chapter 3). We also conducted a diet-
3978 switch experiment to further strengthen our application of stable isotope analysis to
3979 mandrills, by establishing species-specific hair growth rates, isotopic tissue turnover times
3980 and hair-diet isotopic discrimination factors (Chapter 4).

3981 We then completed stable isotope analysis of wild, free-ranging mandrill hair segments, to
3982 examine sex differences in isotopic niche at the levels of individual and sex. These analyses
3983 allowed us to consider whether mandrill dimorphism is associated with niche divergence
3984 between wild males and females and whether individuals of each sex experience differing
3985 levels, or outcomes, of feeding competition (Chapter 5). Finally, we investigated sex
3986 differences in space use, with data obtained from focal follows and GPS collars. Here we
3987 explored whether males and females differ in seasonal patterns of home-range use and travel
3988 behavior, aiming to determine whether and why most male mandrills leave social groups
3989 outside of the breeding season (Chapter 6). Taken together, Chapters 5 and 6 allowed us to
3990 consider the extent to which resource competition or energetic trade-offs may have
3991 contributed to the evolution of extreme sexual dimorphism in mandrills.

3992 In the following discussion I approach several themes in turn, summarizing the results that
3993 are relevant to each theme, before considering the implications of my findings and future
3994 research that could be conducted based on my results. Firstly, I discuss the general
3995 importance of ecologically mediated selection to the evolution of sexual dimorphism. Next, I
3996 consider the consistency of my results with the hypothesis that ecologically mediated
3997 selection has influenced the evolution of sexual dimorphism in mandrills. I then examine how
3998 the results of the fecal analysis in Chapter 3, and the diet-switch experiment in Chapter 4,
3999 facilitated the application of stable isotope analysis to wild mandrills. Finally, I discuss some
4000 of the implications of my results for mandrill conservation.

4001

4002 **7.2 A general role for ecologically mediated selection in the evolution of sexual**
4003 **dimorphism?**

4004 In Chapter 2, we set out to conduct a meta-analysis examining cross-species associations
4005 between sexual size dimorphism and isotopic sex differences in vertebrates, to assess the
4006 strength of associations between sexual dimorphism and ecological sex differences. Because
4007 the nitrogen and carbon stable isotope ratios of animal tissues reflect an individuals' trophic
4008 level or basal carbon resource, respectively (Hobson and Welch, 1992; Codron *et al.*, 2007),
4009 associations between size dimorphism and isotopic differences between males and females
4010 would suggest that sexual dimorphism separately impacts, or has been impacted by, the
4011 feeding ecology of each sex.

4012

4013 *7.2.1 Summary of Results*

4014 Our meta-analytic models revealed that ecological sex differences are widespread, and that
4015 between-species variation could relate to explanatory variables, such as the traits of study
4016 subjects. We found a modest positive association between size dimorphism and sex
4017 differences in trophic level, implying that, across species, greater size dimorphism leads to
4018 greater trophic level sex differences, driven by the larger sex feeding at a higher trophic level
4019 (Figure 2.3). We also revealed the importance of ecological context on associations between
4020 dimorphism and ecological sex differences. Sex differences in trophic level were absent in
4021 herbivores, modest in carnivores and strongest in omnivores (Figure 2.4). Conversely, sex
4022 differences in basal carbon resources were present in herbivores, but not omnivores or
4023 carnivores, thus indicating greater dimorphism is associated with larger sex differences in
4024 plant consumption in herbivores (Figure 2.5). We further found that size dimorphism might
4025 have a greater impact on sex differences in trophic level in gape-limited predators, in which
4026 size may have a greater impact on maximum ingestible prey size and trophic level. However,
4027 controlling for phylogeny weakened the association between size dimorphism and trophic
4028 level sex differences in gape-limited carnivores (Figure 2.6).

4029 The meta-analysis in Chapter 2 considered sex differences in isotopic mean and variance, but
4030 not the isotopic niches of individuals of each sex. When applying stable isotope analysis to

4031 mandrills in Chapter 5, we revealed sex differences in isotopic niche at the individual level
4032 (Figure 5.7; 5.8), a pattern distinct from the sex level mean and variances analyzed in Chapter
4033 2.

4034

4035 *7.2.2 Interpretation and Future Research Directions*

4036 Experimental and modelling work has shown that ecological factors can proximately and
4037 ultimately affect sexual dimorphism (Slatkin, 1984; Bolnick and Doebeli, 2003; De Lisle and
4038 Rowe, 2015; De Lisle, 2023), but the general importance of ecology for explaining variation in
4039 dimorphism is a matter of ongoing debate (De Lisle, 2019; Li and Kokko, 2021). The
4040 associations we found between size dimorphism and isotopic sex differences indicate a cross-
4041 species relationship between sexual size dimorphism and ecological sex differences. Thus, the
4042 influence of ecologically mediated selection on the evolution of sexual dimorphism may be
4043 greater than currently appreciated. However, the modesty of associations between size
4044 dimorphism and trophic level sex differences is also consistent with previous assertions that
4045 the role of ecologically mediated selection in driving sexual dimorphism is relatively minor
4046 (Fairbairn, 1997; De Lisle and Rowe, 2015). Furthermore, our finding that ecological context
4047 modified associations between size dimorphism and ecological sex differences is consistent
4048 with arguments that ecological effects on sexual dimorphism are more likely under a limited
4049 range of circumstances (Li and Kokko, 2021).

4050 Chapter 2 considered isotopic sex differences, measured as sex differences in isotopic mean
4051 or standard deviation. In our analysis of mandrill isotopic sex differences, we found males and
4052 females to be very similar in terms of isotopic mean and standard deviation. Mandrill males
4053 are about three times larger than females (Abernethy and White, 2013), thus the results of
4054 Chapter 5 do not match predictions from Chapter 2, that greater size dimorphism is
4055 associated with larger trophic level differences between males and females. Chapter 5 does
4056 however, conform with our finding in Chapter 2 of high variation among taxa in the extent to
4057 which the sexes differ isotopically.

4058 Our stable isotope analysis of mandrill hairs revealed that, while males and females may be
4059 isotopically similar on average, the sexes can differ greatly at the individual level, with

4060 individual females more variable than individual males. At present, only a few studies have
4061 examined sex differences in individual level isotopic variance (Introduction, Chapter 5).
4062 However, as more studies are conducted on this topic, a meta-analysis of associations
4063 between sexual dimorphism and individual level isotopic sex differences could be an exciting
4064 approach for discovering covariation between sexual dimorphism and intersexual niche
4065 divergence.

4066

4067 **7.3 Ecologically mediated selection as a cause of mandrill dimorphism.**

4068 A great deal of evidence has accumulated suggesting that sexual selection has been an
4069 important driver of mandrill sexual dimorphism (Setchell, 2016). However, a number of
4070 dimorphic traits in mandrills, such as body size and dental morphology, are ecologically
4071 relevant and associated with dietary sex differences (Percher *et al.*, 2017; Nsi Akoué *et al.*,
4072 2017). Furthermore, most male mandrills leave social groups on a seasonal basis, which may
4073 intensify sexual selection for large male size and ornamentation (Abernethy *et al.* 2002;
4074 Hongo *et al.*, 2016). Thus, an interaction between ecologically mediated selection and sexual
4075 selection may best explain why sexual dimorphism is so extreme in mandrills. The primary
4076 aim of this thesis was therefore to investigate whether the sex differences in mandrill feeding
4077 and spatial behavior were consistent with a role for ecologically mediated selection in the
4078 evolution of mandrill dimorphism.

4079

4080 *7.3.1 Summary of results related to ecological sexual dimorphism in mandrills*

4081 The sexes were in similar positions within isotopic space, with $\delta^{13}\text{C}$ means of -24.6‰ and -
4082 24.7‰ and $\delta^{15}\text{N}$ means of 6.1‰ and 6.3‰ for females and males, respectively. The isotopic
4083 niche of male mandrills was 82.5% overlapped by the female isotopic niche and the niches of
4084 each sex were similar in size, with the female niche only slightly larger. However, the sexes
4085 displayed stark niche differences at the individual level. Regarding both $\delta^{15}\text{N}$ (trophic level)
4086 and $\delta^{13}\text{C}$ (plant diversity), random intercept (between individual) variance exceeded residual
4087 (within individual) variance in males. For both isotopes, the female residual (within individual)
4088 variance exceeded that of males. Somewhat surprisingly, female $\delta^{13}\text{C}$ random intercept

4089 (between individual) variance also greatly exceeded male random intercept variance (Chapter
4090 5). We found male habitat selection, home-range size, and Brownian motion variance to differ
4091 more from females during the non-breeding season, than during the breeding season.
4092 Furthermore, we found that males travelled shorter distances, at slower speeds, during the
4093 non-breeding season, compared to the breeding season (Chapter 6).

4094

4095 *7.3.2 The importance of ecologically mediated selection to the evolution of mandrill sexual*
4096 *dimorphism and future research directions*

4097 The mean position of females within isotopic space suggests that the diet of the females we
4098 sampled was omnivorous, as the ranges of both forest plant and forest animal foods
4099 contained the female mean, for both isotopes. Females were also positioned entirely within
4100 the ranges of forest plants and forest animals and did not overlap with savannah food items
4101 (Figure 5.1). Thus, the female isotope data from Chapter 5 is consistent with the fecal data
4102 from Chapter 3, indicating that females largely feed on plants and animal prey within C3 forest
4103 habitats. The mean position of males within isotopic space was very close to the female mean,
4104 for both isotopes (Figure 5.1). Therefore, the diet of males is likely to be similar to that of
4105 females.

4106 Fecal analysis also revealed that females are highly generalist feeders (Chapter 3). Thus, we
4107 can be confident that the isotopic niche ellipse for females (Figure 5.4) is representative of a
4108 generalist feeding niche. The similarity in size between the niche ellipses of both sexes (Figure
4109 5.4), alongside 82.5% overlap of the male ellipse by the female ellipse (Figure 5.5), strongly
4110 suggests that males are also generalist feeders, when considering all individuals combined
4111 (Chapter 5).

4112 However, when analyzing niche variation at the individual level, the results suggested
4113 differing contributions of within and between individual dietary variation to the total niche
4114 widths of each sex (Figure 5.7; Figure 5.8). For both isotopes, between individual variance
4115 exceeded within individual variance in males (Figure 5.8). Thus, the total niche width of males
4116 was explained more by dietary differences between males, than variation in food
4117 consumption within individual males. Dietary differences between males may be expected

4118 because they spend much of the year solitary (Abernethy *et al.*, 2002), meaning individual
4119 males are likely to forage in distinct locations. Thus, differences in food availability between
4120 the foraging locations of individual males could explain high between individual isotopic
4121 variation (Chapter 5).

4122 After accounting for niche variance explained by random intercepts for individuals (i.e.
4123 between individual niche variance), we took the remaining residual isotopic variance as
4124 representing within individual feeding variation (Chapter 5). Female within individual variance
4125 exceeded that of males for both isotopes (Figure 5.8), suggesting that individual females had
4126 more varied diets than individual males, on average. Optimal foraging theory posits that an
4127 animals' food choices should be based on the maximum nutritional return that can be made
4128 by choosing a given item (Pyke *et al.*, 1977). An animal should therefore opt to feed on
4129 preferred resources that provide the highest return on foraging energy investment, given free
4130 choice. Thus, higher dietary consistency for individual males implies that the males we
4131 sampled had more consistent access to preferred resources than individual females.
4132 Conversely, higher within individual variance in females suggests that individual females had
4133 more variable diets than individual males. Therefore, some ecological factor must have
4134 differed between males and females, to cause females to more frequently deviate from
4135 feeding on preferred resources, producing higher within individual isotopic variance in
4136 females.

4137 While male mandrills are often solitary during the non-breeding season, females live in hordes
4138 of several hundred animals throughout the year (Abernethy *et al.*, 2002). Feeding competition
4139 is expected to be higher for individual primates that live in larger groups (Sterck *et al.*, 1997).
4140 Food selection by female mandrills should therefore be impacted to a far greater extent by
4141 within-group feeding competition than solitary male food selection. Differences in the levels
4142 of resource competition experienced by each sex would explain why males exhibited lower
4143 within individual isotopic variance than females. Thus, abandoning social groups outside of
4144 the breeding season may be a male strategy to avoid feeding competition and gain more
4145 consistent access to preferred food sources (Chapter 5).

4146 However, stable isotope analysis alone cannot determine whether the males we sampled
4147 were feeding outside the horde, especially given that each male was within the horde when

4148 it was sampled. Dominant individuals may exclude subordinate individuals from feeding
4149 patches during within-group contest competition (Houle and Wrangham, 2021). It is therefore
4150 possible that the dietary consistency of individual males arose through large males
4151 dominating feeding patches whilst within the horde. Knowledge of each sex's spatial ecology
4152 was therefore needed to explore alternative explanations for the isotopic differences
4153 between male and female mandrills (Chapter 5).

4154 Data from GPS collars revealed that male and female mandrills exhibited differing seasonal
4155 patterns in spatial behavior (Chapter 6). Females were mostly present in gallery forests
4156 throughout the year, as were males during the breeding season. However, during the non-
4157 breeding season, male use of continuous forest increased, with a concomitant decrease in
4158 gallery forest use (Figure 6.3). The fact that males used continuous forest frequently during
4159 the non-breeding season, but females remained almost exclusively within gallery forests
4160 strongly suggests that males at least sometimes left the horde during the non-breeding
4161 season.

4162 Furthermore, male home-range size (Figure 6.2) and Brownian motions variance (Figure 6.4)
4163 were also similar to females during the breeding season, but fell during the non-breeding
4164 season. Groups containing more individuals should occupy larger home-ranges (White *et al.*
4165 2010) and should also more regularly switch between feeding in patches and travelling
4166 between patches (Stevenson and Castellanos, 2001), which would result in high Brownian
4167 motion variance. Therefore, the lower home-range size and motion variance of males during
4168 the non-breeding season is consistent with observations that male mandrills leave hordes
4169 outside of the breeding season and forage alone or in groups of less than five individuals
4170 (Abernethy *et al.*, 2002; Hongo *et al.*, 2016). Thus, the spatial behavior of collared males, the
4171 same individuals as stable isotope data were obtained from, suggests that the male dietary
4172 consistency revealed by the stable isotope analysis results from males leaving the horde, as
4173 opposed to dominating feeding patches within the horde (Chapter 5; Chapter 6).

4174 The GPS collar data also showed that males travelled shorter distances, and moved at slower
4175 speeds, during the non-breeding season, compared to both males in the non-breeding season
4176 and females in both seasons (Figure 6.5; Figure 6.6). Lower travel distances and speeds
4177 suggest that males expended less energy on travel during the non-breeding season (Taylor,

4178 1982; Marcus Rowcliffe *et al.*, 2012). Thus, lower energetic expenditure on travel may be
4179 another benefit to males of leaving mandrill hordes, in addition to having more consistent
4180 access to preferred resources (Chapter 6).

4181 Prior studies of wild mandrills suggest that males experience nutritional and oxidative stress
4182 during the breeding season (Beaulieu *et al.*, 2014; Oelze *et al.*, 2020), and evidence from semi-
4183 free ranging groups indicates that males also suffer higher physiological stress and parasite
4184 loads during this period (Setchell *et al.*, 2010). These results also suggest that males are in
4185 relatively poor body condition by the end of the breeding season. As mandrill hordes occupy
4186 exceptionally large home-ranges (White *et al.*, 2010), the large size of males may leave them
4187 unable to travel alongside the horde whilst in poor condition (Hongo *et al.*, 2016). Leaving
4188 hordes outside of the breeding season, when the availability of fertile females is low, may
4189 therefore allow males to recoup body condition, by expending less energy on travel and
4190 increasing their intake of preferred resources (Chapter 5; Chapter 6).

4191 The temporary nature of male group presence has two impacts on the mandrill mating
4192 system. Firstly, the dominance hierarchies among males are unstable (Smit *et al.*, 2022),
4193 meaning males must compete for dominance upon entering the horde. Secondly, males are
4194 unable to form long-term relationships with females (Abernethy *et al.*, 2002) and must
4195 therefore communicate their quality to females through other means, such as ornaments.
4196 Intra- and inter-sexual breeding competition, the two components of pre-copulatory sexual
4197 selection, are thus likely to be exaggerated by males entering and leaving the horde,
4198 compared to primate species that form more stable societies. Chapter 5 and Chapter 6
4199 suggest resource competition and energetic trade-offs compel males to leave mandrill
4200 hordes. These results imply that ecological factors, unrelated to reproduction, may have
4201 altered the mandrill mating system to raise the intensity of sexual selection on male
4202 morphology. Thus, this thesis provides novel evidence that is consistent with an interaction
4203 between selection mediated by ecology and reproduction explaining why mandrill
4204 dimorphism is so extreme, compared to other primates.

4205 One shortcoming of this conclusion, that mandrill dimorphism is enhanced by males re-
4206 entering the horde on a seasonal basis, is that it does not directly match the ecological
4207 character displacement theory outlined in the General Introduction (Chapter 1). In the

4208 General Introduction, scenarios were introduced in which ecologically mediated disruptive
4209 selection acts directly on morphological traits, such as body size, to produce sexual
4210 dimorphism in response to resource competition. Instead, the above conclusion states that
4211 resource competition and energetic constraints compel male mandrills to move and forage
4212 alone, in order to expend less energy on movement and feed under lower feeding
4213 competition. In this case, ecologically mediated disruptive selection acts on mandrill spatial
4214 and social behavior, rather than morphology, resulting in two alternative behavioral
4215 phenotypes: permanently social females and seasonally solitary males. These sex differences
4216 in behavioral phenotype then create the conditions necessary for amplified sexual selection,
4217 once males re-enter the horde to breed, that might explain the extremity of mandrill
4218 dimorphism. Thus, disruptive selection arising from resource competition and energetic
4219 constraints may be considered to indirectly impact the evolution of sexual dimorphism, via
4220 an effect on the social behavior of each sex, as opposed to directly acting on the morphology
4221 of males and females.

4222 A potential issue with concluding that the ecological sex differences found in this thesis are
4223 evidence of a role for ecologically driven dimorphism is the magnitude of the ecological sex
4224 differences found. The average isotopic and spatial sex differences in Chapters 5 and 6 are
4225 small compared to both the variation exhibited within each sex and to ecological sex
4226 differences found in other species (e.g. Jones et al. 2020). Small ecological sex differences in
4227 mandrills could be viewed as supporting assertions made for other species that the role of
4228 ecologically mediated selection in the evolution of sexual dimorphism is relatively minor (De
4229 Lisle and Rowe 2015; Bauld et al. 2022). Given the weight of evidence for sexual selection as
4230 a driver of mandrill sexual dimorphism (Setchell et al. 2016), a minor role for ecologically
4231 mediated selection would seem reasonable.

4232 Alternatively, it has been argued that documenting ecological sex differences (of any
4233 magnitude) may not provide evidence of the origins of dimorphism, as ecological differences
4234 between males and females could result from dimorphism arising from sexual or fecundity
4235 selection (Shine 1989). The existence of ecological sex differences in the present day might
4236 instead be taken as evidence of ecologically mediated selection acting to maintain, instead of
4237 produce, sexually dimorphic traits (Slatkin 1984). It could therefore be the case that strong
4238 sexual selection has driven the evolution of extreme sexual dimorphism in mandrills, with the

4239 relatively minor ecological sex differences documented in this thesis acting to maintain, as
4240 opposed to exaggerate, dimorphism. It is challenging to disentangle the effects of trait
4241 maintenance and evolution via observational field investigations of the sort conducted in this
4242 thesis. For example, predation pressure is often cited as a driver of larger primate groups, but
4243 present day mortality rates may be an inappropriate measure of predation pressure, as
4244 mortality may demonstrate the rate of predation after primates have executed the formation
4245 of large groups as an evolved anti-predation strategy (Hill and Lee, 1998). Further work is
4246 therefore needed to develop appropriate field based methods to differentiate the evolution
4247 and maintenance of ecological dimorphisms (and other traits) in primates.

4248 Whether viewed as a cause or consequence of mandrill sexual dimorphism, the ecological sex
4249 differences found in this thesis open a number of interesting avenues for future mandrill
4250 research. Firstly, the GPS data strongly implies that males expend less energy during the non-
4251 breeding season, but energetic expenditure could be more closely approximated by
4252 monitoring the activity levels of individuals. Past studies have shown that accelerometers can
4253 be fitted to primates with collars, to record the activity patterns of individuals (Kappeler and
4254 Erkert, 2003; Fernandez-Duque and Erkert, 2006). Like GPS collars, fitting accelerometers to
4255 primates allow behavioral data to be gathered remotely, meaning that activity pattern data
4256 could be gathered on males, even when they are solitary and difficult to track and observe. In
4257 baboons (*Papio ursinus*), accelerometers have successfully recorded the duration and timing
4258 of behaviors including locomotion, foraging, resting, and grooming (Fehlmann *et al.*, 2017).
4259 Accelerometers could therefore be used to determine whether males devote more time to
4260 locomotion during the breeding season and more time to resting and foraging during the non-
4261 breeding season. Seasonal differences in activity patterns could further support our
4262 conclusion that male energetic expenditure varies between the breeding and non-breeding
4263 seasons.

4264 Another open question concerns the traits under sexual selection in the context of males
4265 entering and remaining in mandrill hordes to compete for females. Results from semi-free
4266 ranging individuals indicate that higher body mass and brighter coloration are positively
4267 associated with rank attainment and reproductive success in male mandrills (Setchell, 2005;
4268 Setchell *et al.*, 2008). Though measuring animal color with camera traps presents a number
4269 of issues, including the need to control for variation in lighting conditions (Stevens *et al.*,

4270 2007), technological innovations are improving the application of digital photography to the
4271 measurement of animal color (Stevens *et al.*, 2009; Troscianko and Stevens, 2015). For
4272 example, the inclusion of color cards within photographs, alongside software for color
4273 standardization, can enhance the precision of color measurement under varying light
4274 conditions (Bergman and Beehner, 2008). Camera traps can also be used to estimate the body
4275 mass of wild animals (Tarugara *et al.*, 2019).

4276 Hongo *et al.* (2016) demonstrated that male presence in mandrill hordes, as well as mate-
4277 guarding behavior, can be recorded using camera traps. As technology develops, it would be
4278 interesting to investigate whether the brightness of red sexual skin or the body size of males
4279 are associated with joining the horde or mate-guarding behavior. As mate-guarding is
4280 associated with reproductive skew in semi free-ranging mandrills (Wickings *et al.*, 1993), an
4281 association between coloration or body size and mate-guarding behavior could provide
4282 indirect evidence of sexual selection on these traits in wild mandrills. Furthermore, if the
4283 average brightness of males declines over the course of the breeding season, it might indicate
4284 that the average body condition of males also declines during the breeding season. Such a
4285 result would be consistent with the conclusions drawn in this thesis that poor body condition,
4286 resulting from mating competition, is part of the explanation for male mandrills leaving
4287 hordes outside of the breeding season.

4288 Why mandrill females form such large groups is also an important knowledge gap, given that
4289 large group size likely leads to high resource competition and extensive travel (Emlen and
4290 Oring, 1977; Sterck *et al.*, 1997; White *et al.*, 2010; Cassini, 2020b). Groups of mandrills rarely
4291 encounter each other (White *et al.*, 2010), so between-group contest competition is perhaps
4292 unlikely to explain why mandrills form such large groups. Two other factors that may drive
4293 large group formation are inter-specific feeding competition and predation (Isbell, 1994;
4294 Sterck *et al.*, 1997). Mandrills co-occur with a number of other frugivorous species (Tutin *et*
4295 *al.*, 1997) and are preyed upon by leopards (*Panthera pardus*), African golden cats (*Profelis*
4296 *aurata*), birds of prey and African rock pythons (*Python sebae*) (Abernethy and White, 2013),
4297 though predation pressure in African tropical forests is presumed to be lower than for
4298 primates in open savanna habitats (Hongo, 2016). The influence of inter-specific competition
4299 and predation on mandrill behavior could be investigated using camera trap data to examine
4300 temporal and spatial patterns of co-occurrence of mandrills with competitor species and

4301 predators (Murphy *et al.*, 2019; Andrade-Ponce *et al.*, 2022). Comparative analysis which
4302 other semi-arboreal primate species would likely be needed to truly answer why mandrill
4303 groups are so large. But directly comparing the influence of competitor species and predators
4304 on mandrill space use could reveal which of these factors is more influential on mandrill
4305 behavior.

4306

4307 **7.4 Ensuring robust application of stable isotope analysis to wild mandrills.**

4308 Stable isotope analysis is best applied to wild animals, with knowledge of the feeding ecology
4309 of the focal species, the stable isotope routing within the species' tissues (Martinez del Rio *et al.*,
4310 2009), and of the sources of isotopic variation to which the focal animals may be exposed
4311 (Roberts *et al.*, 2017). The fecal analysis we conducted produced reliable information on the
4312 feeding ecology of the focal horde, thanks to which we could draw stronger dietary inferences
4313 from the stable isotope ratios of mandrill tissues (Chapter 3). The diet-switch experiment
4314 provided data on the growth rates of mandrill hairs and isotopic turnover and fractionation
4315 within mandrill hairs (Chapter 4). Finally, stable isotope analysis of mandrill food items
4316 revealed the sources of isotopic variation within the focal horde's home-range (Chapter 5).

4317

4318 *7.4.1 Summary of results impacting the application of stable isotope analysis*

4319 We found that the Lopé mandrill horde are highly generalist feeders that consume a diverse
4320 range of food types when fruit production falls (Chapter 3). From the diet-switch experiment,
4321 we found that mandrill hair growth rates were variable between body parts and individuals.
4322 We also found that mandrill hair stable isotope ratios reach equilibrium with a new diet after
4323 23 days for $\delta^{15}\text{N}$ and 38 days for $\delta^{13}\text{C}$. The isotopic discrimination factors of mandrill hairs
4324 were 3.1‰ and 2.8‰ for $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$, respectively (Chapter 4). Analyzing the stable isotope
4325 ratios of mandrill food items, we found high diversity within major food types and very little
4326 clustering between food types. Isotopic variation between mandrill food items did not appear
4327 to be strongly influenced by systematic spatial or temporal isotopic baseline variation
4328 (Chapter 5).

4329

4330 *7.4.2 Implications for stable isotope analysis of mandrills and future research directions*

4331 Analysis of mandrill fecal samples revealed that the horde are highly generalist feeders that
4332 consume a wide range of fallback foods when fruit availability declines (Tutin *et al.*, 1997;
4333 Hongo *et al.*, 2018). Because of the social structure of the horde, that often contains very few
4334 males (Abernethy *et al.*, 2002), the information from the fecal data largely applies to females
4335 and juveniles. However, the long-term, eight year duration of the study means that our
4336 inferences about the hordes' feeding niche can probably account for inter-annual changes in
4337 fruit supply. We can therefore be confident that females are generalist feeders when
4338 interpreting female stable isotope data. This knowledge of females then provides a robust
4339 context within which to interpret the stable isotope data of male mandrills, whose ecology is
4340 less well known.

4341 To further improve our application of stable isotope analysis to mandrills, we conducted a
4342 diet-switch experiment to quantify hair growth rates, isotopic tissue turnover times and hair-
4343 diet isotopic discrimination factors. Measuring hair growth rates showed substantial amounts
4344 of variation between the growth rates of different hair types and between the same hair type
4345 from different individuals. Inter-individual differences in hair growth rates are clearly difficult
4346 to account for in wild individuals. However, our results indicate that error can be minimized
4347 by analyzing samples of the same hair type from different individuals. We therefore analyzed
4348 only mustache hairs from wild individuals. Mustache hairs were also advantageous because
4349 they are relatively thick, meaning they contain more mass per unit of length. We found the
4350 minimum length we could use to achieve analytical precision was 2 mm, equivalent to an
4351 ecological window of approximately 5.5 days (Chapter 4).

4352 Tissue-diet isotopic discrimination factors describe the typical enrichment between the stable
4353 isotope ratios of animal tissues and animal diets. Discrimination factors must be accounted
4354 for to accurately reconstruct animal diets from stable isotope data. The need to account for
4355 discrimination factors can be seen in Figure 5.1 (Chapter 5) which displays the adjusted
4356 isotope ratios of male and female mustache hairs, relative to mandrill food items. Both the
4357 male and female adjusted $\delta^{15}\text{N}$ values are approximately 3‰, in the lower range of the forest
4358 plant $\delta^{15}\text{N}$ distribution. Without adjustment, the $\delta^{15}\text{N}$ position of both sexes would be above

4359 6 ‰, perhaps incorrectly suggesting a greater contribution of animal prey to the diet of both
4360 sexes, compared to the adjusted values.

4361 The importance of quantifying discrimination factors is greatest when applying stable isotope
4362 mixing models to animal stable isotope data (Kadye *et al.*, 2020). These models allow the
4363 relative contribution of different food types to an animals' diet to be estimated using stable
4364 isotope data (Parnell *et al.*, 2013). However, the extensive overlap we found between the
4365 isotope ratios of mandrill food items (Figure 5.1; Figure A5.1) made the application of stable
4366 isotope mixing models inappropriate in our case, as these models rely on isotopic distinction
4367 between food sources (Phillips *et al.*, 2005). Recent advances also allow tissue turnover times
4368 to be included into stable isotope mixing models (Ballutaud *et al.*, 2022). Despite the noise
4369 within the dataset (Figure 4.3; Figure 4.4), the estimates of mandrill isotopic discrimination
4370 factors and tissue turnover times were consistent with prior studies in primates and other
4371 mammals (Schwertl *et al.*, 2003; Caut *et al.*, 2008; Huelsemann *et al.*, 2009; Tsutaya *et al.*,
4372 2017; 2021). Our diet-switch experiment thus provides novel data that may be used in future
4373 applications of stable isotope analysis to mandrills or other African Papionines, where
4374 advanced stable isotope mixing models may be more appropriately utilized.

4375 Stable isotope analysis is also improved by an understanding of the isotopic variation within
4376 the environment of the animals under study. Stable isotope analysis of mandrill food items
4377 revealed that substantial isotopic variation was present within the habitat of the focal horde
4378 (Figure 5.1). However, this variation was continuous between different food types, with no
4379 obvious clustering of the isotopic signal of different foods. Grouping the food items into 11
4380 categories, including different C3 plant tissue types, did not appear to be a suitable framework
4381 within which to interpret the isotope ratios of mandrill hair samples (Figure A5.1). Numerous
4382 overlapping pairwise combinations of categories were present in the data; for example, the
4383 range of C3 plant seed isotope ratios was almost entirely overlapped by the ranges of C3 fruits
4384 and C3 leaves. Thus, we could not reliably draw inferences on the different plant tissue types
4385 that mandrills could have consumed, based on stable isotope data (Chapter 5).

4386 Grouping food types into five categories appeared to be a more suitable solution. C3 forest
4387 plants were relatively easily distinguished from C4 savannah plants along the $\delta^{13}\text{C}$ axis and
4388 animals were, on average, ^{15}N -enriched relative to plants (Figure 5.1). We were therefore

4389 more confident that we could draw inferences on mandrill diets using more general principles
4390 of stable isotope ecology, concluding that: greater diversity in plant consumption through
4391 time would lead to greater $\delta^{13}\text{C}$ variation along the length of mandrill hairs, and greater
4392 trophic level variation through time would increase $\delta^{15}\text{N}$ variation along hairs.

4393 We did not find systematic spatial or temporal variation in the stable isotope ratios of mandrill
4394 food items (Chapter 5). Thus, we did not find any reason, nor have any means, to adjust the
4395 stable isotope values of mandrill tissue samples to account for sampling season or location,
4396 before interpreting the data from wild mandrill hairs. One factor we did not consider is the
4397 'canopy effect', which describes vertical stratification of the isotope ratios of plant tissues,
4398 due to differences in transpiration rate between the canopy and forest floor (Lowry *et al.*
4399 2021). The canopy effect is not universal across forests, however, and depends on local
4400 conditions (Blumenthal *et al.*, 2015). As the home-range of the focal horde largely consisted
4401 of gallery forests, that are less dense relative to continuous forest, we assumed that
4402 differences between the canopy and forest floor in relative humidity, temperature, and light
4403 availability that impact transpiration rates would be low, compared to closed canopy
4404 continuous forest. We therefore did not account for vertical stratification of the isotope ratios
4405 of mandrill food items. However, Oelze *et al.* (2020) concluded that age related inter-
4406 individual differences in mandrill blood isotope ratios could have resulted from younger
4407 individuals feeding higher in the canopy. Given that other mandrill hordes also forage in the
4408 continuous forest in Lopé (Lahm, 1986), investigation of the canopy effect in this study site
4409 may be useful. If the canopy effect is found, it may be possible to use stable isotopes to reveal
4410 sex differences in foraging height, which could add a three dimensional component to
4411 analyses of sex differences in mandrill isotopic niche.

4412

4413 **7.5 Mandrill Conservation**

4414 Mandrill socioecology remains understudied, with potential adverse impacts for mandrill
4415 conservation. Mandrills are classified as Vulnerable to extinction on the IUCN Redlist, with
4416 habitat loss, climate change and subsistence hunting by humans constituting the principal
4417 threats to mandrill populations. Though this thesis does not directly address questions related
4418 to mandrill conservation, the results from Chapters 3, 4 and 5 contain dietary and spatial

4419 information that could inform the direction of future work to improve the management of
4420 these threats to mandrills.

4421 *7.5.1 Summary of results relevant to mandrill conservation*

4422 We found fruit to be the most frequently consumed resource by mandrills, with consumption
4423 covarying positively with availability. Fruit consumption exhibited clear peaks during the two
4424 wet seasons, when fruit abundance was high. During periods of low fruit availability, mandrills
4425 consumed a range of fallback foods, including animal prey, leaves, crushed seeds, and other,
4426 lower quality plant fibers (Chapter 3). Using GPS collars, we found that the focal hordes'
4427 home-range was almost entirely limited to the village enclave in the buffer zone of Lopé
4428 National Park, and that the horde sometimes crossed the boundary of the park into
4429 unprotected areas (Chapter 5).

4430

4431 *7.5.2 Implications for mandrill conservation and future research directions*

4432 Our results suggest that mandrills are omnivores, with a preference for fruit, which is
4433 consistent with previous studies (Hoshino, 1985; Lahm, 1986; Rogers *et al.*, 1996; Nsi Akoué
4434 *et al.*, 2017). Also following previous studies, we found that fruit consumption was seasonal,
4435 and followed patterns of fruiting tree phenology (Hongo *et al.*, 2018). Fruit production has
4436 dropped by around 30% in Lopé National Park over the previous three decades, as a result of
4437 ongoing climate change (Bush *et al.*, 2020a; 2020b), meaning availability of the mandrills'
4438 preferred resource has fallen. The available information on mandrill diets indicates dietary
4439 plasticity that could confer resilience under environmental change scenarios, as species which
4440 demonstrate flexibility in already varying environments may be adapted to cope with future
4441 change (Jones, 2005). However, the sensitivity of mandrills to tree phenology may, like other
4442 fruit dependent vertebrates, leave them vulnerable to climate change induced reductions in
4443 fruit supply (Butt *et al.*, 2015). Whether dietary flexibility will facilitate adaptation by mandrills
4444 to novel environmental conditions is unclear, however, there are several promising avenues
4445 for answering this question, as noted below.

4446 As outlined in Chapter 3, data from Lopé indicate that forest elephants are consuming lower
4447 quality diets in 2020 than in 1990 (Tejler *et al.*, *unpublished data*) and exhibit more frequent

4448 and severe emaciation (Bush *et al.*, 2020). Therefore, a useful next step regarding mandrills
4449 would be to investigate whether present day mandrill diets contain greater amounts of
4450 fallback foods or lower quality fruit, compared to the 1998 - 2004 period examined in Chapter
4451 3. A present day analysis of the focal hordes' diet may be especially useful if coupled with
4452 nutritional analysis to elucidate, for example, whether disrupted phenological patterns and
4453 warmer temperatures are adversely impacting mandrill macronutrient intake (Rothman *et al.*,
4454 2015), or associated with increasing levels of plant secondary metabolite consumption
4455 (Windley *et al.*, 2022).

4456 The change in the hordes' spatial behavior over the past twenty years may also be concerning.
4457 In Chapter 6 we found that the hordes home-range has shifted northwards and eastwards,
4458 since the horde was first tracked between 1998 and 2004 (White *et al.*, 2010). This new area
4459 of occupation contains greater amounts of oil palm (*Elaeis guinensis*) fruits (David Lehmann
4460 *personal communication*), which are a keystone resource to the horde (Chapter 3). However,
4461 this area partially falls outside the strictly protected area of Lopé National Park, and overlaps
4462 with village enclaves in the buffer zone. We even observed the horde crossing the Offoué
4463 river, moving outside of the park boundaries.

4464 If the home-range of the horde has shifted closer to human habitation due to falling fruit
4465 availability, other hordes throughout the mandrill range could be exhibiting similar behavioral
4466 changes. Thus, mandrills may be exposed to increasing levels of human contact, which could
4467 increase the hunting pressure on mandrill populations. Of greater concern is that mandrills
4468 are known to crop-raid in human plantations (Djoko *et al.*, 2022). Therefore, changing
4469 mandrill spatial behavior could potentially lead to increased human-wildlife conflict, through
4470 crop-raiding behavior that disrupts the livelihoods and food supplies of local people (Redpath
4471 *et al.*, 2013). Research to determine whether the rate of crop-raiding behavior by mandrills
4472 (and other species) is increasing in areas impacted by forest loss or climate change would
4473 therefore be useful. Investigations into the environmental factors and management
4474 interventions that could mitigate the costs of mandrill crop-raiding will also be of benefit,
4475 both to mandrill conservation and local stakeholders affected by human-wildlife interactions
4476 (Ngama *et al.*, 2019).

4477

4478 **7.6 Conclusion**

4479 This thesis contains new knowledge of mandrill ecology and makes novel contributions to the
4480 debate about how diversity in sexual dimorphism arises. We have found that mandrills are
4481 extreme generalists and this flexibility may confer some ability to adapt to ongoing
4482 environmental changes. However, the sensitivity of mandrills to fruiting tree phenology, high
4483 feeding competition among females and recent changes to the focal hordes' spatial behavior
4484 that we have documented could all suggest that mandrills are vulnerable to declining habitat
4485 quality. These results can therefore guide further research to enhance the knowledge base
4486 underpinning mandrill conservation policy.

4487 Though sexual selection is likely to be the primary cause, why mandrill dimorphism is so
4488 extreme is still an unanswered question. Our analysis of mandrill feeding and spatial ecology
4489 strongly suggest a role for ecologically mediated selection. Our results indicate that a
4490 combination of resource competition and energetic trade-offs compel males to leave mandrill
4491 hordes outside of the breeding season. This response by males to ecological pressures then
4492 leads to an influx of males to mandrill hordes during the breeding season, with these males
4493 unfamiliar to each other and to the females resident within hordes. The resulting need of
4494 males to establish dominance and attract females likely produces strong intra- and inter-
4495 sexual mating competition, triggering strong sexual selection for traits such as large size, long
4496 canines, and bright ornamentation. It therefore seems likely that a combination of selective
4497 pressures, mediated by both ecology and reproduction, are responsible for the evolution of
4498 extreme sexual dimorphism in mandrills.

4499 Our findings contribute to the growing body of evidence that sexual dimorphism can be driven
4500 by ecologically mediated selection. In addition to evidence of ecological drivers of dimorphism
4501 in mandrills, we also found cross-species associations between sexual size dimorphism and
4502 ecological sex differences. These associations suggest that an interplay between sexual
4503 dimorphism and ecology may be widespread. However, our meta-analysis also highlighted
4504 that ecological sex differences are pervasive and potentially explained by factors other than
4505 sex differences in size. We therefore hope that our results will stimulate new investigations
4506 into the drivers of ecological sex differences, that will generate a complete answer to the
4507 question of how sexual dimorphism evolved.

4508

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