

**Negative density dependence in the mortality and growth of tropical tree seedlings is strong, and primarily caused by fungal pathogens**

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1 **Negative density dependence in the mortality and growth of tropical tree**  
2 **seedlings is strong, and primarily caused by fungal pathogens**

3 ***Running Title***

4 Natural enemies contribute to tree diversity

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25 ***Statement of Authorship***

26 All authors formulated and designed the study, KH and HB established the  
27 experiment, all authors collected data, KH and CETP performed the data analysis,  
28 KH wrote the first draft and all authors contributed to the revisions.

29 ***Data Availability Statement***

30 Should this manuscript be accepted, all data supporting these results will be  
31 archived in Dryad.

32

**33 Abstract**

- 34 1. Natural enemies have been implicated as agents of negative density  
35 dependence (NDD) in tropical forests, but their relative contributions to NDD,  
36 and thus to the maintenance of diversity, are largely unknown.
- 37 2. We monitored the rates of survival and relative growth rates on seedlings for  
38 ten years in tropical moist forest in Manu National Park, Peru. We then  
39 experimentally manipulated the plots to exclude fungal pathogens, insects,  
40 small mammals, and large mammals for an additional 31 months to assess  
41 the influence of these natural enemies on density-dependent interactions  
42 among tropical seedlings.
- 43 3. Fungal pathogens made the most important contribution to negative density  
44 dependence. The application of fungicide led to lower mortality rates, faster  
45 growth rates, and decreased species diversity. Other taxa of natural enemies  
46 had at most minor effects on seedling performance.
- 47 4. *Synthesis.* We conclude that fungal pathogens are the strongest contributors  
48 to the widely observed NDD that occurs among seedlings. Moreover, the  
49 presence of fungal pathogens augments the species diversity of seedlings,  
50 indicating their critical contribution to the maintenance of species coexistence  
51 and the structure of tropical tree communities.

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## 53 **Introduction**

54 Negative density dependence (NDD) has received extensive empirical support as a  
55 key mechanism of species coexistence (Harms, Wright, Calderón, Hernández, &  
56 Herre, 2000; Liu et al., 2012; Wright, 2002; Zhu, Woodall, Monteiro, & Clark, 2015).  
57 NDD promotes coexistence by reducing individual performance at high conspecific  
58 density, thereby favouring rare species (Chesson, 2000).

59 Mechanisms that contribute NDD have been widely debated (Terborgh, 2012;  
60 Wright, 2002). NDD was once thought to be driven by competition among  
61 neighbouring plants for shared resources, but little evidence for competition among  
62 tropical rainforest seedlings has emerged (Paine, Harms, Schnitzer, & Carson, 2008;  
63 Svenning, Fabbro, & Wright, 2008). Multiple studies, on the other hand, have shown  
64 that natural enemies cause NDD by disproportionately preying upon locally common  
65 species (Bagchi et al., 2014; Bell, Freckleton, & Lewis, 2006; Jia et al., 2020; Packer  
66 & Clay, 2000; Paine, Beck, & Terborgh, 2016). Studies examining NDD focus on  
67 seedlings, as much of tree community structure is determined during the seedling  
68 stage, when mortality rates are high and non-random with respect to species (Green,  
69 Harms, & Connell, 2014).

70 We still know little about the relative importance of the taxa of natural enemies that  
71 contribute to NDD (Bagchi et al., 2014; Gripenberg et al., 2014; Jia et al., 2020;  
72 Paine et al., 2016). Bagchi et al. (2014) found both fungal pathogens and insect  
73 herbivores to cause NDD among seedlings in Belize, although only fungi affected  
74 species diversity. Their results are supported by those of Jia et al (2020), who  
75 additionally found that functional traits affected the severity of NDD in recruitment  
76 and survival. Paine et al. (2016) evaluated the contributions of mammals to NDD

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77 during seedling recruitment in Peru, finding that small and medium mammals, but not  
78 large mammals, affected mortality and diversity. We build upon these studies by  
79 evaluating the relative contributions of fungi, insects, small mammals, and large  
80 mammals to NDD in seedling mortality, growth, and species diversity. Each of the  
81 four taxa can contribute to NDD, as all consume seedlings (Packer & Clay, 2003;  
82 Paine et al., 2016).

83 The natural enemies most commonly implicated as drivers of NDD are pathogenic  
84 fungi or oomycota (referred to as fungal pathogens hereafter; Bell et al., 2006;  
85 Mangan et al., 2010). They are commonly highly host specific or have a limited host  
86 range (Gilbert, Magarey, Suiter, & Webb, 2012). Herbivorous insects can also cause  
87 NDD (Bagchi et al., 2014; Fricke, Tewksbury, & Rogers, 2014), though their  
88 contributions have been debated (Bagchi, Press, & Scholes, 2010; Gripenberg et al.,  
89 2014). Furthermore, small and large terrestrial mammals can also cause NDD (Beck,  
90 Snodgrass, & Thebpanya, 2013; Theimer, Gehring, Green, & Connell, 2011),  
91 although they appear to have limited consequences for diversity (Paine et al., 2016).

92 This study aims to determine the degree to which different taxa (fungi, insects and  
93 mammals) contribute to NDD and consequently maintain tree species diversity. We  
94 hypothesize that smaller natural enemies will make larger contributions to NDD  
95 (Bagchi et al., 2014), because they are far more abundant, and are more likely to be  
96 host specific (Gilbert et al., 2012), than are the larger-bodied taxa. We also  
97 hypothesize that the contributions of differing taxa of natural enemies to NDD will  
98 interact. For example, García-Guzman and Dirzo (2001) found that insect herbivores  
99 accelerated rates of pathogen infection by creating wounds through which pathogens  
100 attacked seedlings. To test this hypothesis, we experimentally combined exclusion

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101 treatments, expecting to find stronger effects on NDD in plots from which fungi,  
102 insects, or mammals were jointly excluded. We assess two aspects of seedling  
103 performance: mortality and relative growth rate (RGR). Because generalist natural  
104 enemies impact multiple species relatively mildly, we expect seedling RGR to be  
105 more strongly affected by generalist natural enemies such as insects and mammals  
106 (Beck et al., 2013; Novotny et al., 2002). We expect mortality, in contrast, to be more  
107 strongly driven by host specific enemies (Comita, Muller-Landau, Aguilar, & Hubbell,  
108 2010), and we therefore expect fungi to contribute more to conspecific NDD in  
109 mortality (Bagchi et al., 2014; Gilbert et al., 2012).

## 110 **Methods**

### 111 ***Study site***

112 This study was carried out at the Cocha Cashu Biological Station (CCBS). CCBS is  
113 located in Amazonian South-East Peru in lowland tropical rain forest, at 11°51'S,  
114 71°19'W, 350 m elevation. This seasonal forest receives a mean of 2167 mm of rain  
115 annually, and mean daily temperatures vary between 21.8°C and 24.2° over the  
116 course of a year (Paine, 2007). The site is in a highly diverse and remote area of  
117 Manu National Park, with over 350 tree species with a diameter  $\geq 10$  cm DBH. It has  
118 experienced minimal hunting, and no logging or mining, during the last century  
119 (Hazelwood et al., 2020).

### 120 ***Experimental design***

121 Circular 1 m<sup>2</sup> experimental plots were established in a random blocked design  
122 throughout a 4 km<sup>2</sup> area of mature floodplain rain forest. 24 plots were spaced  
123 between 5 and 10 meters apart in each of 24 blocks, avoiding trails and newly fallen

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124 trees, for a total of 576 plots. Within each plot, all woody seedlings  $\geq 10$  cm and  $< 100$   
125 cm in stem height were identified and tagged over eight censuses between 2003 and  
126 2017. Height was measured on all seedlings as the vertical distance from the soil to  
127 the apical meristem. All understory shrubs and lianas were excluded. Owing to the  
128 blocked design of the seedling plots, it was not feasible to identify the adult trees  
129 neighbouring them. Unfortunately, this precluded the assessment of adult  
130 competition on seedlings. Censuses were carried out 269 to 1566 days apart (see  
131 Paine & Harms, 2009 for details).

132 The experimental phase of the study began in October 2014, when we applied  
133 treatments to exclude fungi, insects, and mammals. Within each block, eight plots  
134 were randomly selected for the application of one of eight treatments: none (a  
135 control), fungi, insects, large mammals, all mammals, fungi and insects, all mammals  
136 and fungi, and all mammals and insects. The fungicide Amistar (Syngenta Ltd, active  
137 ingredient: azoxystrobin) provides a broad spectrum of protection against fungal  
138 attack, has low toxicity in non-target organisms, and was found to be effective by  
139 Bagchi et al. (2014). The insecticide Karate (Syngenta Ltd. active ingredient: lambda  
140 cyhalothrin), provides protection against a broad spectrum of insect herbivores,  
141 leaving low rates of residue and has low impact on non-target organisms. Pesticides  
142 were applied according to manufacturer's instructions, mixing 1.25 ml of pesticide  
143 with 1 litre of water, and applying 50 ml of the mixture to 1 m<sup>2</sup> plot with spray bottles.  
144 Pesticides were applied to treatment plots every 10 to 14 days, in equal amounts  
145 over 31 months, with some treatment breaks when it was logistically impossible to  
146 apply treatments (max 1 month). Control plots were misted with an amount of water  
147 equivalent to that applied to pesticide plots.

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148 We excluded mammals from the study plots using 2×2 m wire mesh exclosures.  
149 These were 150 cm high and included a 50 cm buffer around each plot to reduce  
150 potential germination bias from perching birds. The 'Large mammal' exclosures  
151 allowed the entry of small mammals through 15×15 cm holes cut into the base of the  
152 mesh. These were large enough to allow agoutis (*Dasyprocta* sp.) or smaller rodents  
153 to enter, but were too small for peccaries (*Pecari* and *Tayassu* spp.), deer (*Mazama*  
154 *americana*), or tapir (*Tapirus terrestris*). The 'All mammal' exclosures, on the other  
155 hand, were constructed flush to the ground and excluded all terrestrial mammals.  
156 Previous studies at the same site have shown this design of exclosures to be  
157 extremely effective at excluding terrestrial mammals (Beck et al., 2013; Paine et al.,  
158 2016).

159 No treatment was applied to the remaining 16 plots in each block. Seedling mortality  
160 and growth were modelled against conspecific and heterospecific neighbourhood  
161 density in control plots (where water was applied) and non-treatment plots (where no  
162 water was applied). No significant differences were found between control and non-  
163 treatment plots in any models ( $P \geq 0.16$ ). Therefore, we combined control and non-  
164 treatment plots for all subsequent analyses.

### 165 **Data analysis**

166 In eight censuses conducted over 153 months, we monitored 10,557 seedlings from  
167 638 unique species or morpho-species. Of these, 1,317 individuals were unidentified  
168 or identifiable only to family; they were excluded from all analyses. Evaluating the  
169 effects of conspecific crowding on very rare species was not possible. We therefore  
170 also excluded species that were represented by fewer than 10 seedlings, even if  
171 those species may be relatively abundant as adults. The resulting dataset consisted



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172 of 8,018 individuals representing 149 unique species and morpho-species, with  
173 76.8% identified to species level and the remainder identified to genus level.

174 Seedlings excluded from the dataset were counted among heterospecific  
175 neighbours.

176 Conspecific and heterospecific neighbourhood density, and the sizes of neighbouring  
177 seedlings, can impact mortality or RGR. We therefore calculated neighbourhood

178 crowding indices (NCI) for each seedling per plot and census as:  $NCI_k = \sum^N \frac{height_n}{height_k}$ ,

179 where k is the focal individual, and n indexes over the N neighbouring seedlings.

180 Because individual seedlings were not mapped within plots, distances among  
181 seedlings were not available. Indices were calculated separately for conspecific  
182 neighbours and heterospecific neighbours.

183 We first evaluated the extent to which density dependence shapes the rates of  
184 mortality and growth in this community. Mortality was predicted by heterospecific and  
185 conspecific crowding indices using a generalized linear mixed effect model with  
186 binomial errors. We included an offset of the log-transformed time between censuses  
187 to account for differing census intervals and to yield parameter estimates in units of  
188 years. We assessed the effects of crowding on RGR using a similarly structured  
189 linear mixed-effect model, in which relative growth rate was predicted by conspecific  
190 and heterospecific crowding indices. Seedling RGR was assumed to be exponential  
191 since growth rates do not slow until trees reach a height far exceeding 1 m.

192 In all models, species was included as a random intercept to account for differing  
193 rates of mortality or growth among species. We evaluated whether the effects of  
194 conspecific or heterospecific crowding differed among species by testing the support

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195 for models that included random slopes among species. Larger seedlings experience  
196 lower mortality risk (Green et al., 2014; Paine et al., 2012), therefore log-transformed  
197 seedling height was included as a fixed effect in all models. We evaluated the  
198 support for size dependent responses in models that included interactions between  
199 height and both crowding indices. Finally, location, coded as plots nested within  
200 blocks, was included as a random intercept, to account for spatial heterogeneity in  
201 the rates of mortality and growth across the study site.

202 Secondly, we assessed the effects of fungi, insects, and mammals as contributors of  
203 density-dependent mortality and growth. To do so, we included interactions between  
204 conspecific crowding and treatment, and between heterospecific crowding and  
205 treatment. The first assessment evaluated the effects of fungi, insects, large  
206 mammals and small mammals. The effects of fungi, insects, and large mammals  
207 were evaluated by comparing the mortality and growth of seedlings in control plots to  
208 their performance from which each taxon was excluded. The effects of small  
209 mammals was evaluated by comparing mortality and growth of seedling in plots from  
210 which large mammals were excluded against plots from which all mammals were  
211 excluded. These effects were tested using one-degree of freedom orthogonal  
212 contrasts.

213 To evaluate whether fungi, insects, and mammals make interactive contributions to  
214 density dependence, we built three further models: 1) fungi and insects: This model  
215 included the treatments fungicide, insecticide, fungicide and insecticide, and control;  
216 2) fungi and mammals: including the treatments fungicide, all mammals, fungicide  
217 and all mammals, and control; and 3) mammals and insects, including the treatments  
218 all mammals, insecticide, all mammals and insecticide, and control. In each of these

219 tests, we evaluated the support for a statistical interaction between the two main  
220 effects on seedling mortality and growth.

221 We assessed the effects of each experimental treatment on diversity using the  
222 exponent of the Shannon-Weiner diversity index for each plot at each census, which  
223 can be interpreted as the number of equally abundant species in a community (Jost,  
224 2006). To assess the effects of each exclusion treatment on species diversity, and  
225 how the effects may change through time, diversity was predicted by the interaction  
226 of treatment and census period.

227 All analyses were performed in R 4.0.2 (R Core Team, 2020), using package lme4  
228 (Bates, Mächler, Bolker, & Walker, 2012). All models were compared on an AIC  
229 basis. Model residuals were examined for overdispersion and heteroscedasticity  
230 using the DHARMA package (Hartig, 2020). No issues were identified. P values and  
231 confidence intervals were derived from non-parametric bootstrapping using 10,000  
232 replicates.

## 233 **Results**

234 We estimated the effects of conspecific and heterospecific density on mortality and  
235 relative growth rate over the entire study period (2003-17). Across all seedlings, the  
236 estimated annual probability of mortality was 25.3% per year in the absence of  
237 crowding from neighbours. Mortality increased with conspecific crowding. An order-  
238 of-magnitude increase in conspecific crowding increased the mortality rate from  
239 27.1% to 32.1% ( $P < 0.001$ , Fig. 1A). In contrast, heterospecific crowding did not  
240 affect the risk of mortality ( $P = 0.10$ , Fig. 1B).

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241 Increasing seedling height consistently and significantly reduced mortality rates, but  
242 height did not interact significantly with crowding: large and small seedlings were  
243 similarly affected by crowding from neighbours. The estimated annual relative growth  
244 rate was  $0.61 \text{ mm}\cdot\text{cm}^{-1}\cdot\text{y}^{-1}$  for median-sized seedlings (15 cm in height) in the  
245 absence of crowding from neighbours (Fig. 1C). However, there was great variation  
246 in growth rates, with many seedlings growing or shrinking, even in the absence of  
247 crowding from neighbours. There was a strong interaction between height and both  
248 conspecific and heterospecific crowding in RGR ( $P < 0.0001$ , Fig 1C &D). Growth  
249 among smaller seedlings increased with increasing conspecific and heterospecific  
250 crowding, whereas the opposite pattern was observed for larger seedlings, though  
251 the decline was stronger with heterospecific crowding (Fig. 1D).

## 252 ***Mortality***

253 We assessed the contribution of each taxon of natural enemies to density dependent  
254 mortality during the experimental period of the study (2014-2017). The effect of  
255 conspecific crowding on mortality was significantly reduced by fungicide application  
256 ( $P = 0.0476$ , Fig. 2A). Fungicide application reduced mortality rate by 19% at a  
257 conspecific neighbourhood crowding index of 10, corresponding to a focal seedling  
258 in a plot with 10 conspecific seedlings of the same size as the focal. This result  
259 suggests that fungi are a major contributor to conspecific density dependence. In  
260 contrast, the application of insecticide and the use of mammal-excluding cages did  
261 not significantly interact with conspecific crowding ( $P \geq 0.13$ ). The exclusion of small  
262 mammals led to a minor, though significant, decrease in mortality under high  
263 heterospecific crowding, whereas the application of fungicide increased seedling  
264 mortality under conditions of high heterospecific crowding (Fungicide:  $P = 0.0001$ ;

265 Small mammals:  $P = 0.046$ , Fig 2B). The other treatments had no significant effects  
266 ( $P \geq 0.13$ ). There was no significant interaction between the exclusion of insects,  
267 large mammals, or small mammals and conspecific crowding, suggesting that these  
268 taxa do not contribute in complementary manners to the effects of conspecific  
269 crowding on seedling mortality ( $P \geq 0.30$ , Figs. 2C, E, & G). The interactive  
270 contribution of fungicide and insecticide to heterospecific crowding were significant  
271 ( $P = 0.0212$ ), but were intermediate between their independent effects (Fig. 2D). The  
272 combined application of fungicide and the exclusion of all mammals led to a  
273 significant increase in mortality in conditions of high heterospecific crowding ( $P =$   
274  $0.0001$ , Fig 2F). Insecticide and the exclusion of all mammals had no interactive  
275 effects on mortality (Fig. 2H).

### 276 ***Relative growth rate***

277 The application of fungicide led significantly increased relative growth rates under  
278 conditions of high conspecific crowding ( $P = 0.0172$ , Fig. 3A). In contrast, the  
279 application of insecticide significantly reduced growth rates ( $P < 0.0001$ ), whereas  
280 the exclusion of large and small mammals had no significant effects. The interactive  
281 contributions of fungicide and insecticide to conspecific crowding significantly  
282 increased rates of growth in conditions of high conspecific crowding ( $P = 0.0091$ , Fig.  
283 3C). Similarly, the interactive contributions of fungicide and the exclusion of all  
284 mammals also slightly but significantly, increased growth rates ( $P = 0.0250$ , Fig. 3E).  
285 Insecticide and the exclusion of all mammals had no interactive effects on growth  
286 (Fig. 3G). In contrast to the effects on conspecific crowding, no experimental  
287 treatment, or combination of treatments, influenced the growth response to  
288 heterospecific crowding ( $P \geq 0.13$ , Figs. 3B, D, F, H).

**289 Diversity**

290 The effect of exclusion treatments on diversity was assessed by monitoring the  
291 change in the exponent of Shannon-Weiner diversity since the experimental  
292 treatments were first applied in October 2014. The application of fungicide caused a  
293 rapid and significant decrease in diversity, which persisted over the experimental  
294 period (Fig. 4A). The application of insecticide, and the exclusion of large mammals,  
295 also reduced diversity, though these effects only became significant at the final  
296 census ( $P \leq 0.006$ , Fig. 4A). The exclusion of small mammals had no effect at any  
297 time ( $P \geq 0.055$ ). When treatments were combined, their joint effects reduced  
298 diversity, at least by the final census, but their effects were intermediate between the  
299 relevant main effects in all cases (Fig. 4B, C, D).

300 Digging a bit deeper into the diversity data, we dissected the effects of each taxon of  
301 natural enemies on seedling species richness and evenness. The patterns observed  
302 in species diversity were mirrored by those of species richness. Patterns in species  
303 evenness, in contrast, were extremely minor, and divorced from those of diversity  
304 (Supplemental Figure 1). This indicates that natural enemies, particularly fungi, affect  
305 seedling diversity primarily through the local exclusion of species, rather than through  
306 effects on their relative abundance.

**307 Discussion**

308 The extent and causes of negative density dependence (NDD) in plant performance  
309 have long been investigated, owing to the fundamental importance of NDD in  
310 promoting the coexistence of species (Chesson, 2000; Wright, 2002). We examined  
311 the effects of NDD on mortality and growth in tropical tree seedlings using data from  
312 a 14-year period. The extensive dataset allows us to robustly conclude that

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313 conspecific crowding increases mortality rates, whereas crowding from  
314 heterospecific neighbours has much weaker effects (Fig. 1A, B). These patterns are  
315 broadly in accord with previous studies (Jia et al., 2020; Piao, Comita, Jin, & Kim,  
316 2013; Terborgh, 2012; Wills et al., 2006). Conspecific crowding, and, to a lesser  
317 extent, heterospecific crowding increased the growth of small seedlings, but reduced  
318 the growth of larger ones (Fig. 1C, D). This result stands in contrast with previous  
319 studies, which found a purely negative effect on growth from conspecific neighbours.  
320 This delayed negative effect could arise through interspecific competition (Tanner,  
321 Teo, Coomes, & Midgley, 2005). However, competition among seedlings is unlikely,  
322 given their relative scarcity in the understory of closed-canopy tropical forests (Paine  
323 et al., 2008; Svenning et al., 2008). It should be noted that the stem density of  
324 seedlings at our study site ( $6.4 \pm 4.8$  stems  $m^{-2}$ ; Harms, Powers, & Montgomery,  
325 2004) is comparable to that found in other Neotropical rain forests (Paine et al.,  
326 2008). A stronger possibility is that the primary drivers of density dependent mortality  
327 and growth are natural enemies (Paine et al. 2016). Seedling germination and the  
328 initial survival of seedlings can be promoted in favourable microsites (Paine &  
329 Harms, 2009). Thereafter, natural enemies, including fungal pathogens, insects and  
330 mammals (Sedio & Ostling, 2013), would be attracted to high densities of host  
331 species, and their effects can be expected to accumulate over time to impact larger  
332 seedlings.

333 Studies of NDD have recently come under scrutiny for bias (Detto, Visser, Wright, &  
334 Pacala, 2019). Previous studies have over- and under-estimated the strengths of  
335 NDD because of their use of error-prone proxies. We believe that our conclusions  
336 are largely insulated from these biases. First, we assessed longitudinal data within a

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337 single life stage (seedlings), rather than transitions between life stages. This  
338 approach is not biased to detect NDD when none is present (Detto et al., 2019).  
339 Second, we used a similar analytical technique throughout the study. In other words,  
340 if bias taints our results, they should all be biased in a similar fashion. Thus, we are  
341 able to accurately assess the contribution of each taxon of natural enemies to NDD.  
342 Therefore, we are confident in the conclusions of this study.

343 Our primary conclusion, that fungi are the predominant contributor to NDD in  
344 seedling mortality and growth, confirms that of previous studies (Bagchi et al., 2014;  
345 Gripenberg et al., 2014; Packer & Clay, 2000; Paine et al., 2016). Our study builds  
346 upon this body of work in four ways: 1) We establish the baseline level of NDD over  
347 14 years of observation, 2) We investigate the contributions of four key taxa of  
348 natural enemies to density-dependent performance using a consistent framework, 3)  
349 we examine the degree to which various taxa of natural enemies interact, and 4) we  
350 monitor both growth and survival. Despite the clear support for the leading role of  
351 fungi, all four taxa we investigated could plausibly have contributed to the NDD  
352 observed at this site (Fig. 1). Fungal pathogens and oomycota can negatively affect  
353 the performance of tropical tree seedlings (Augspurger, 1983; Bagchi, Swinfield, et  
354 al., 2010), even as mycorrhizal fungi are key to seedling establishment and growth.  
355 Similarly, many insect herbivores are host-specific or clade-specific (Forister et al.,  
356 2015; Novotny et al., 2002), and could therefore contribute to NDD.

357 We found that fungal pathogens contributed more to negative density dependence in  
358 mortality and growth than did insects or mammals, as the application of fungicide  
359 reduced rates of mortality and increased relative growth rates in conditions of high  
360 conspecific crowding (Figs. 2A & 3A). These results are consistent with studies that



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361 found strong associations between NDD and fungal pathogens (Bagchi et al., 2014;  
362 Bell et al., 2006; Jia et al., 2020; Packer & Clay, 2000), and validate a mechanism  
363 originally proposed by Janzen (1970) and Connell (1971). Given that the application  
364 of fungicide is likely to have reduced the abundance of both pathogenic and  
365 beneficial fungi, the strong positive effects of fungicide application on seedling  
366 performance lend further credence to our interpretation that pathogenic fungi are the  
367 key contributors to negative density dependence in this system, as they overcame  
368 countervailing positive effects from mycorrhizal fungi (Bagchi et al., 2014). Fungal  
369 pathogens are often highly host specific (Gilbert & Webb, 2007); it is this host-  
370 specificity that allows pathogens to pass between seedlings rapidly under high  
371 conspecific crowding. When we excluded fungi, mortality decreased and growth  
372 rates increased with conspecific density (Figs 2A & 3A). This indicates that, once  
373 released from the negative effects of fungal pathogens, the environmental conditions  
374 were favourable enough for high densities of conspecifics to thrive. This further  
375 supports the argument that intra-specific competition among seedlings is weak  
376 (Paine et al., 2008).

377 Moreover, the application of fungicide also reduced species diversity, suggesting that  
378 the depredations of fungal pathogens play a key role in maintaining diversity in this  
379 community (Fig. 4A). To a lesser degree, diversity was also promoted by insects and  
380 large mammals, but they appear to do so through non-NDD processes. Notably,  
381 fungi more strongly affected species richness than evenness (Supplemental Figure  
382 1). This further suggests the role pathogenic fungi play in excluding seedling  
383 species, and thereby structuring the tree community.

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384 In addition to fungi and insects, small mammals have been found to contribute to  
385 NDD at this site and in Corcovado National Park (CNP), Costa Rica (Demattia,  
386 Rathcke, & Curran, 2006; Paine et al., 2016). These two studies, however, focused  
387 on seed removal, rather than seedling growth or mortality, which may have led us to  
388 underestimate the contribution of small mammals to NDD. Beck et al. (2013) found,  
389 at the same study site, that large mammals also contributed to seedling mortality and  
390 abundance. However, we observed that White-lipped Peccaries (*Tayassu pecari*),  
391 which are major consumers of seeds and seedlings, occurred at unusually low  
392 densities at the site during the experimental phase of this study. We speculate that  
393 this may account for the contrast between our results and those of Beck et al.  
394 (2013). Moreover, none of the previous mammal-focused studies observed fungal  
395 pathogens or insects. Mammals can contribute to NDD in seedling recruitment, but  
396 their effects appear to occur during an earlier ontogenetic phase than that which we  
397 examined here.

398 If one taxon facilitates the consumption of another, then they could interactively  
399 affect the growth and eventually the mortality of seedlings. For example, insect  
400 herbivores can accelerate rates of pathogen infection (García-Guzmán & Dirzo,  
401 2001). In the current study, interactions among taxa of natural enemies that  
402 intensified NDD were uncommon. Only the joint exclusions of fungi and insects, and  
403 fungi and mammals, increased growth rates in conditions of conspecific crowding  
404 (Fig. 3C, E). More frequently, the joint effect of excluding multiple taxa was  
405 intermediate between their independent effects, or was opposite to the pattern  
406 expected under NDD. This may be because perturbing multiple taxa of natural  
407 enemies simultaneously made the rates of mortality and growth more stochastic.

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408 Overall, our results suggest that natural enemies are affecting NDD interactively, but  
409 rather than their effects are additive on the seedling community. This suggests that  
410 they are impacting different sections of the seedling community (Forister et al., 2015;  
411 Novotny et al., 2002), and provides evidence for differences among species in  
412 vulnerability to different types of natural enemies.

413 Multiple studies have examined mortality as an indicator of NDD in seedlings.  
414 Mortality, however, provides only a coarse indication of the strength of NDD, as it is  
415 a binary response, and can occur only once to any organism. RGR is a more  
416 sensitive detector of NDD, as it can vary at sub-lethal scales. RGR is also a noisy  
417 response variable, owing to the many factors unrelated to NDD that can cause it to  
418 vary (Zhu et al., 2015). Effects of heterospecific crowding on RGR, but not on  
419 mortality, implicate mechanisms that impact plant health but do not necessarily  
420 cause mortality, although a decline in health can result in death. Both competition  
421 and generalist natural enemies may have a slow but non-fatal impact on seedlings  
422 (Murrell, 2009; Theimer et al., 2011), and it is possible that these mechanisms  
423 reduced growth rates. In the current study, the insights provided by mortality, growth,  
424 and diversity were largely concurrent, together indicating that fungi are the  
425 predominant contributors to NDD in this system.

## 426 **Conclusions**

427 Our study demonstrates that fungal pathogens increase mortality rates, decrease  
428 growth rates, and promote species diversity in a highly diverse tropical moist forest.  
429 Insects and large mammals also promote diversity, but they appear to do so through  
430 non-NDD processes. Given the outsized importance of seedling recruitment in  
431 structuring tropical tree communities (Green et al., 2014; Harms et al., 2000), it is

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432 likely that the effects of fungi on seedlings persist throughout ontogeny. Although our  
433 results suggest that diversity is maintained by multiple drivers and should not be  
434 assigned a unique mechanism, we present strong evidence that fungal pathogens  
435 are uniquely important in shaping tropical tree communities, and are  
436 disproportionately important drivers of diversity in tropical forests.

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- 600



**601 Figures****602 Figure 1**

603 Predicted effects of conspecific and heterospecific neighbourhood crowding on  
604 mortality and relative growth rate between 2003 and 2017. Panels A) and B) display  
605 mortality, whereas C) and D) show relative growth rate. The three fitted lines on each  
606 panel represent the rates of mortality (A & B) or growth (C & D) for seedlings that  
607 began each census interval with heights of either 10, 20, or 50cm. Shaded areas  
608 represent 95% confidence intervals, derived from non-parametric bootstrapping.  
609 Histograms and right-hand vertical axes in panels A) and B) show the distribution of  
610 crowding experienced by seedlings that died (at bottom of panels) or survived  
611 between censuses (at top). Points in panels C) and D) represent the growth rates of  
612 seedlings. Many points are over-printed, either because crowding was zero, or  
613 because the observed rate of growth was zero. Note that all x-axes are log-  
614 transformed.

**615 Figure 2**

616 Probability of mortality predicted by conspecific crowding index (left column) or  
617 heterospecific crowding index (right column). Panels A) and B) show the change in  
618 mortality rates incurred by the exclusion of fungal pathogens, insects, large  
619 mammals, and small mammals. Panels C) and D) show the interacting effects of  
620 fungicide and insecticide with crowding. Panels E) and F) show the interacting  
621 effects of fungicide and all-mammal exclosures with crowding. Panels G) and H)  
622 show the interacting effects of all-mammal exclosures and insecticide with crowding.  
623 The P-values shown on each panel represent the significance of the interaction of  
624 each term with the crowding index (See Appendix 1 for complete statistical results).

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625 Significant P-values represent effects that differ from the effects of crowding in  
626 control plots. Shaded areas represent 95% confidence intervals, derived from non-  
627 parametric bootstrapping. ns: not significant. Note that all x-axes are log-  
628 transformed.

629 **Figure 3**

630 Relative growth rate predicted by conspecific crowding index (left column) or  
631 heterospecific crowding index (right column). Panels A) and B) show the change in  
632 growth rate incurred by the exclusion of fungal pathogens, insects, large mammals,  
633 and small mammals. Panels C) and D) show the interacting effects of fungicide and  
634 insecticide with crowding. Panels E) and F) show the interacting effects of fungicide  
635 and all-mammal exclosures with crowding. Panels G) and H) show the interacting  
636 effects of all-mammal exclosures and insecticide with crowding. Formatting  
637 otherwise follows that of Figure 2.

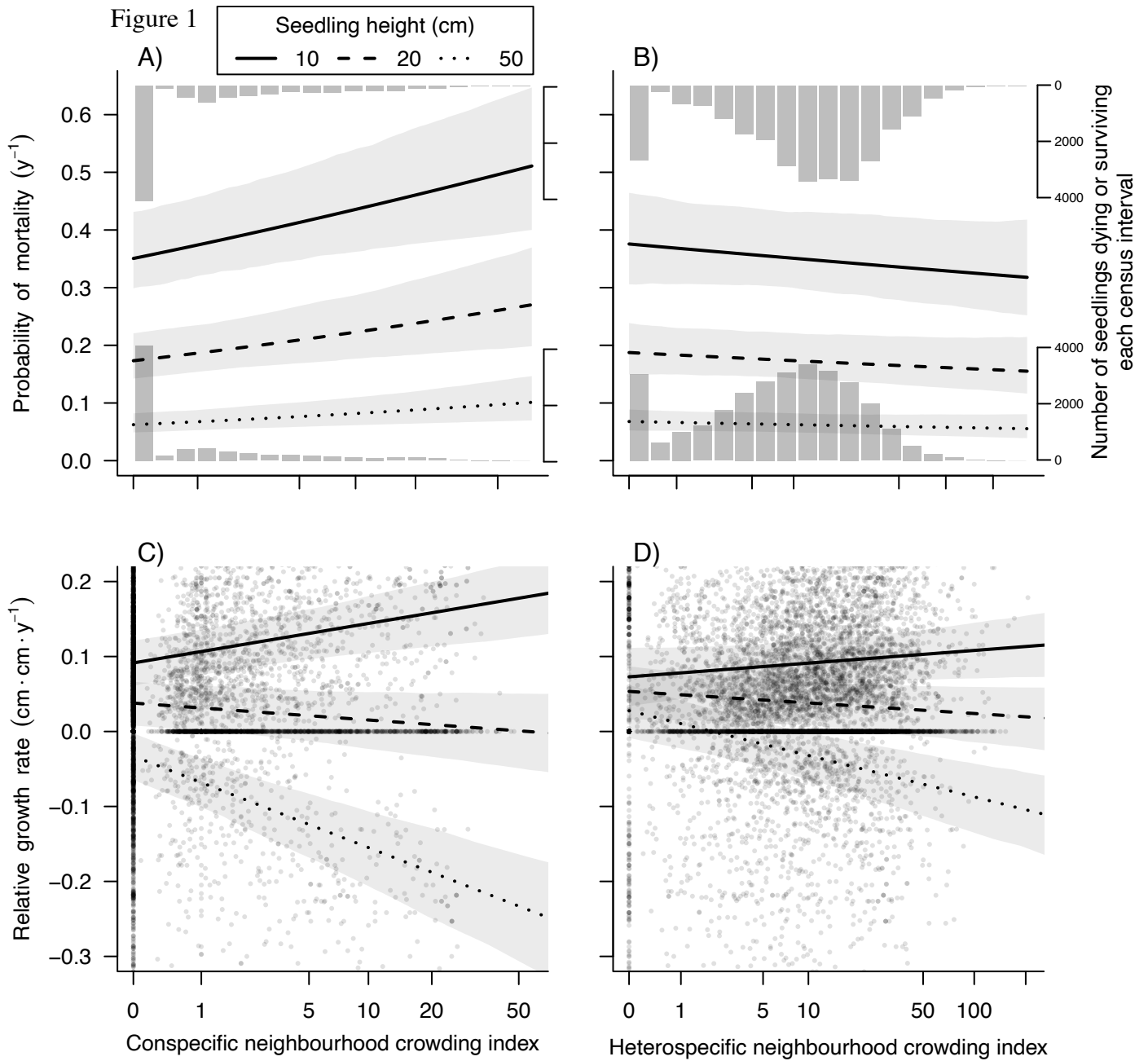
638 **Figure 4**

639 Change in Shannon-Weiner Diversity index under the exclusion of A) fungal  
640 pathogens, insects, small mammals and large mammals. Panel B) shows the  
641 interacting effects of fungicide and insecticide on diversity, panel C) shows the  
642 interacting effects of fungicide and all-mammal exclosures, and panel D) shows the  
643 interacting effects of all-mammal exclosures and insecticide. Significant changes  
644 from the diversity level in September 2014 (just prior to the beginning of treatment  
645 application) are shown as filled dots. Error bars represent 95% confidence intervals,  
646 derived from non-parametric bootstrapping.

647

**648 Supplemental Figures****649 *Supplemental Figure 1***

650 Change in species richness (A-D) and evenness (E-H) under the exclusion of A & E)  
651 fungal pathogens, insects, small mammals and large mammals. Panels B & F show  
652 the interacting effects of fungicide and insecticide on richness and evenness, panels  
653 C & G show the interacting effects of fungicide and all-mammal exclosures, and  
654 panels D & H show the interacting effects of all-mammal exclosures and insecticide.  
655 Significant changes from the levels of richness and evenness in September 2014  
656 (just prior to the beginning of treatment application) are shown as filled dots. Error  
657 bars represent 95% confidence intervals, derived from non-parametric bootstrapping.  
658 Note the difference in scale on the Y axes of panels A-D and E-H.



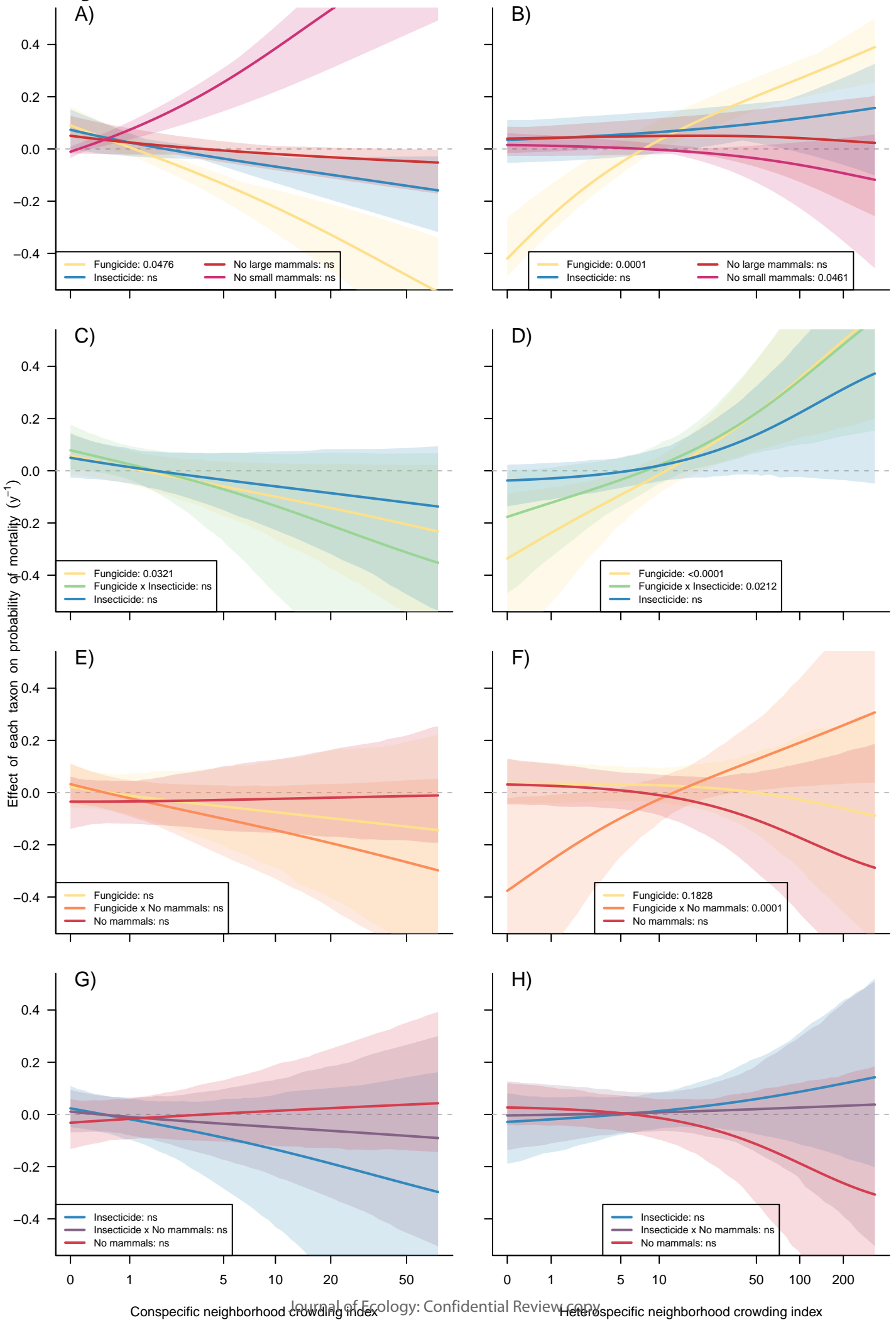
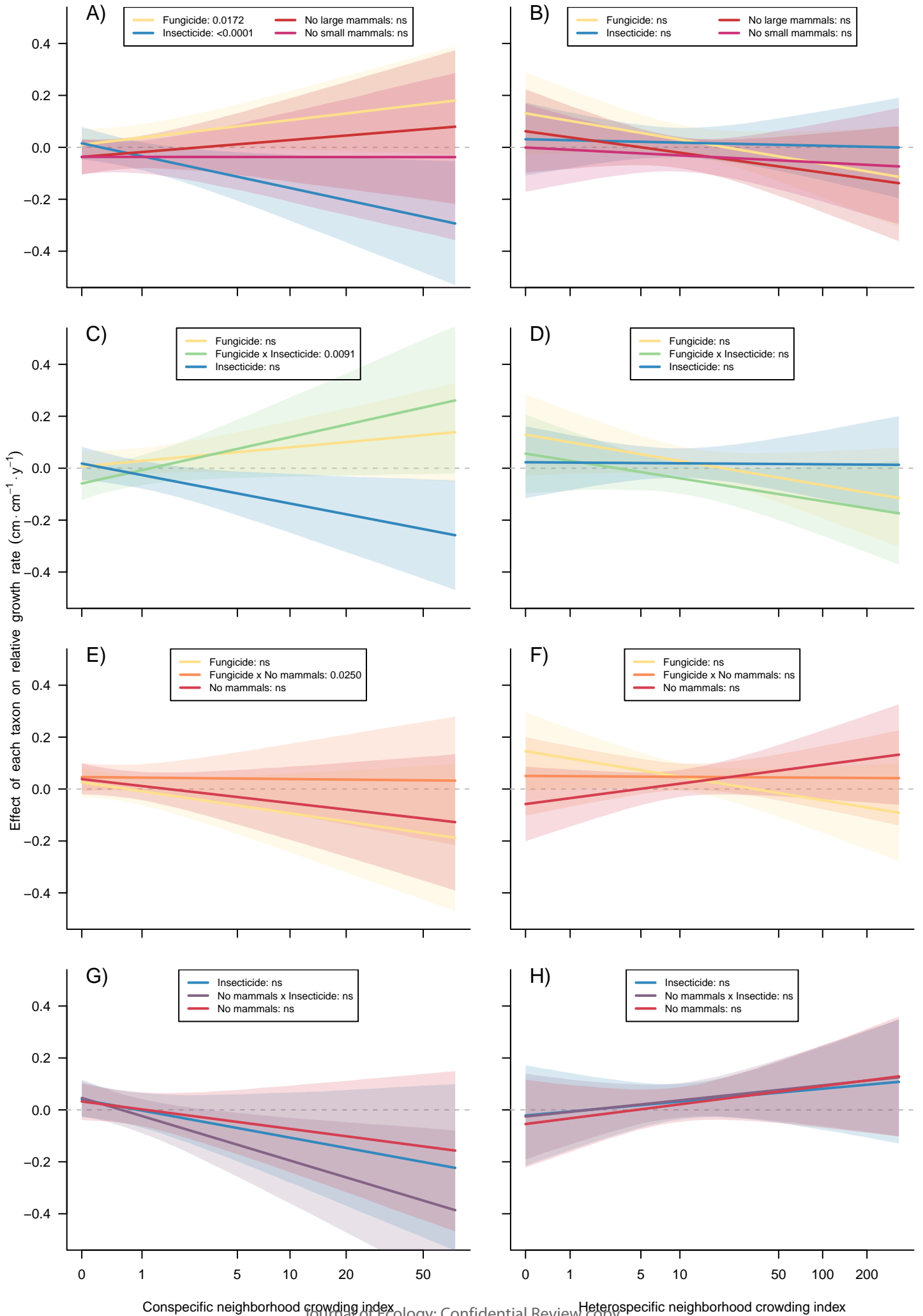
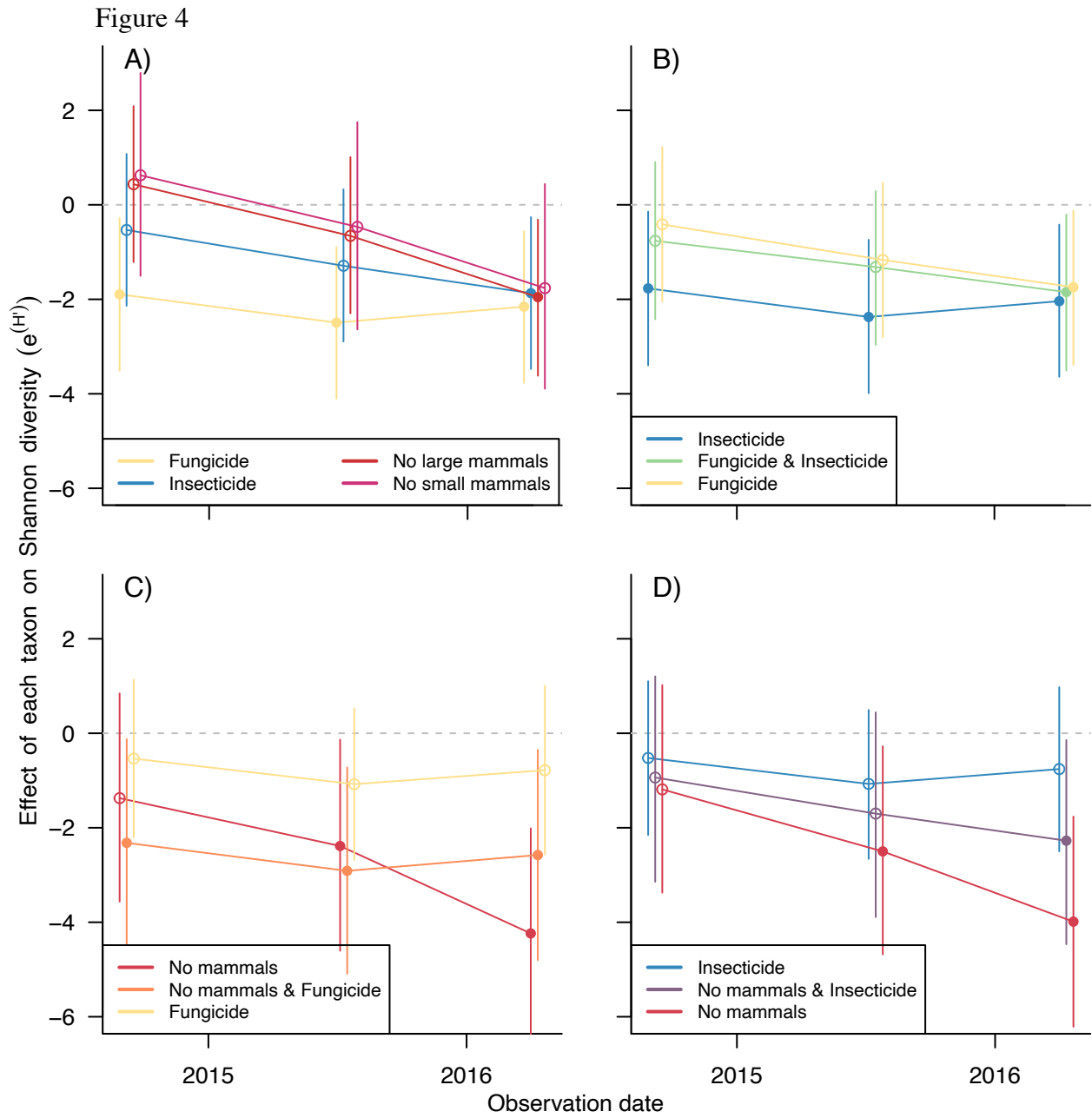
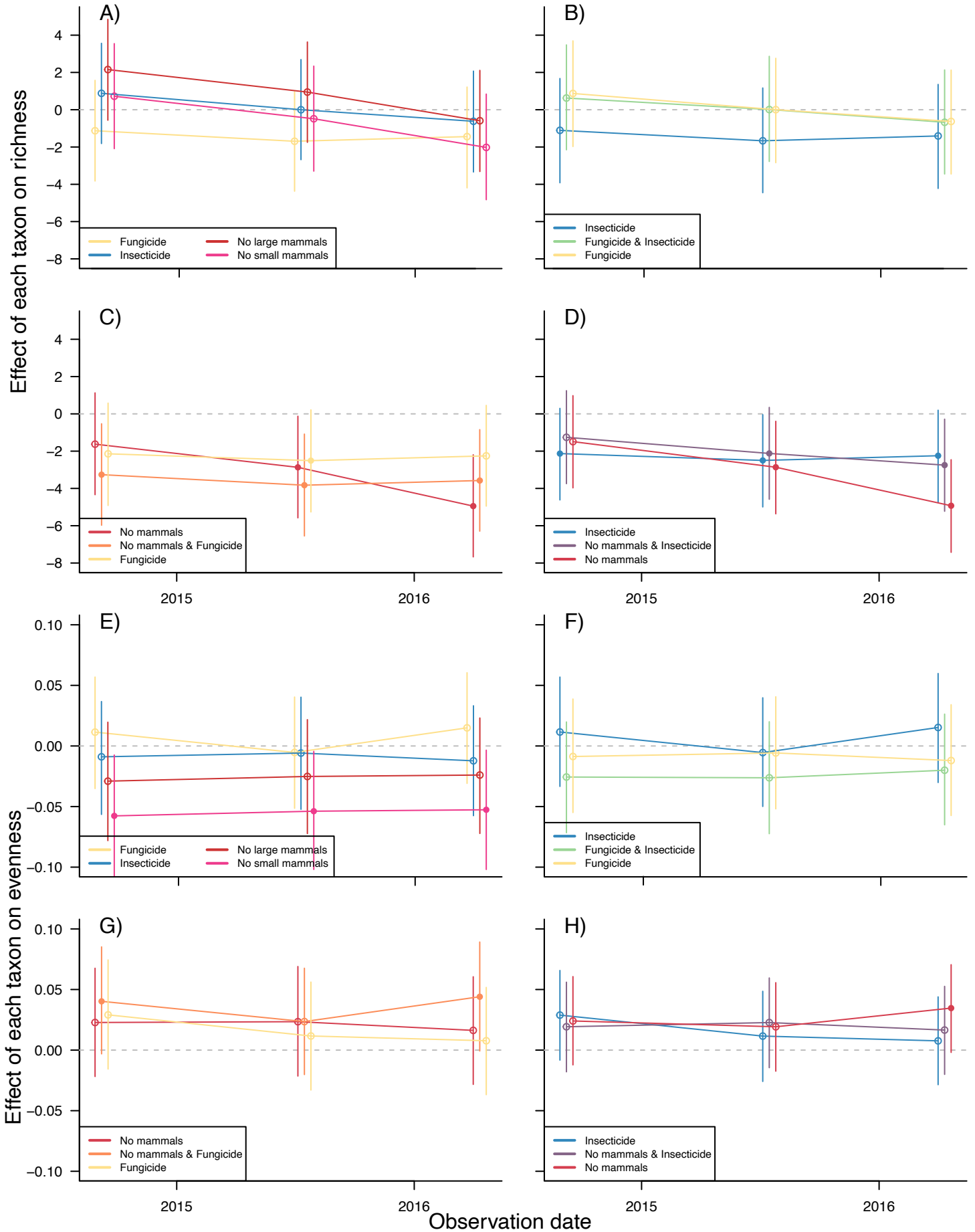


Figure 3





Supplemental Figure 1





**Appendix 1: Complete statistical output for all analyses**

This document contains all the statistical output for the results presented in Hazelwood et al “Negative density dependence in the mortality and growth of tropical tree seedlings is strong, and primarily caused by fungal pathogens”

**DENSITY DEPENDENCE (Figure 1)****Figure 1A & 1B**

Generalized linear mixed model fit by maximum likelihood (Laplace Approximation) ['glmerMod']

Offset: offset\_time, Family: binomial (cloglog)

event ~ consp.crowd.l + hetsp.crowd.l + ht.l + (1|trans/plot) + (1|Census) + (consp.crowd.l|Name) + (hetsp.crowd.l|Name)

AIC	BIC	logLik	deviance	df.resid
15241.4	15340.1	-7607.7	15215.4	14594

Random effects:

Groups	Name	Variance	Std.Dev.	Corr
plot.trans	(Intercept)	0.131053	0.36201	
Name	(Intercept)	0.255901	0.50587	
	hetsp.crowd.l	0.012195	0.11043	-0.19
Name.1	(Intercept)	0.170577	0.41301	
	consp.crowd.l	0.031816	0.17837	-0.04
trans	(Intercept)	0.007364	0.08581	
Census	(Intercept)	0.091562	0.30259	

Number of obs: 14607, groups: plot:trans, 516; Name, 143; trans, 24; Census, 11

Fixed effects:

	Estimate	Std. Error	z value	P value
(Intercept)	1.96983	0.19904	9.897	0.000
consp.crowd.l	0.11740	0.03847	3.052	0.000
hetsp.crowd.l	-0.03591	0.02981	-1.205	0.102
ht.l	-1.18234	0.04477	-26.412	0.000

R<sup>2</sup>: Marginal 0.1393731, Conditional: 0.2997038

**Figure 1C & 1D**

Linear mixed model fit by maximum likelihood ['lmerMod']

$$\text{rgr} \sim (\text{consp.crowd.l} + \text{hetsp.crowd.l}) * \text{ht.l} + (1|\text{trans/plot}) + (1|\text{Census}) + (\text{consp.crowd.l}|\text{Name}) + (\text{hetsp.crowd.l}|\text{Name})$$

AIC	BIC	logLik	deviance	df.resid
-2038.3	-1923.4	1035.2	-2070.3	9697

Random effects:

Groups	Name	Variance	Std.Dev.	Corr
plot.trans	(Intercept)	0.0006340	0.02518	
Name	(Intercept)	0.0022570	0.04751	
	hetsp.crowd.l	0.0004096	0.02024	-0.97
Name.1	(Intercept)	0.0001903	0.01379	
	consp.crowd.l	0.0005975	0.02444	-0.36
trans	(Intercept)	0.0002024	0.01423	
Census	(Intercept)	0.0023253	0.04822	
Residual		0.0459208	0.21429	

Number of obs: 9713, groups: plot:trans, 503; Name, 143; trans, 24; Census, 11

Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	0.137514	0.034871	3.944	0.0000
consp.crowd.l	0.126172	0.017359	7.268	0.0000
hetsp.crowd.l	0.054187	0.012259	4.420	0.0000
ht.l	-0.028064	0.009350	-3.001	0.0016
consp.crowd.l:ht.l	-0.045262	0.006176	-7.329	0.0000
hetsp.crowd.l:ht.l	-0.020219	0.004053	-4.988	0.0000

R<sup>2</sup>: Marginal 0.07440856, Conditional: 0.1582671

**MORTALITY (Figure 2)****Figure 2A & 2B**

Generalized linear mixed model fit by maximum likelihood (Laplace Approximation) ['glmerMod']

Offset: offset\_time, Family: binomial (cloglog)

event ~ (consp.crowd.l + hetsp.crowd.l)\*Trt + ht.l + (1|trans/plot) + (1|Census) + (1|Name)

AIC	BIC	logLik	deviance	df.resid
1117.1	1223.5	-538.6	1077.1	1488

Random effects:

Groups	Name	Variance	Std.Dev.
Name	(Intercept)	7.722e-01	0.878725
plot:trans	(Intercept)	4.731e-05	0.006878
trans	(Intercept)	2.084e-01	0.456497
Census	(Intercept)	2.128e-01	0.461341

Number of obs: 1508, groups: Name, 105; plot:trans, 78; trans, 16; Census, 3

Fixed effects:

	Estimate	Std. Error	z value	P value
(Intercept)	1.83847	0.98791	1.861	0.0030
consp.crowd.l	0.14057	0.11424	1.230	0.1347
hetsp.crowd.l	0.07311	0.14046	0.521	0.2513
TrtFungi	-1.78587	0.65550	-2.724	0.0005
TrtInsects	0.10104	0.70482	0.143	0.3760
TrtSmall Mammals	0.40799	0.69468	0.587	0.1252
TrtLarge Mammals	0.67241	0.87550	0.768	0.0429
ht.l	-1.40947	0.23993	-5.874	0.0000
consp.crowd.l:TrtFungi	-0.27287	0.16668	-1.637	0.0476
consp.crowd.l:TrtInsects	0.11193	0.20983	0.533	0.3085
consp.crowd.l:TrtSmall Mammals	0.23218	0.31483	0.737	0.1663
consp.crowd.l:TrtLarge Mammals	-0.31912	0.29724	-1.074	0.1289
hetsp.crowd.l:TrtFungi	0.68258	0.21693	3.147	0.0001
hetsp.crowd.l:TrtInsects	0.01507	0.23814	0.063	0.4715
hetsp.crowd.l:TrtSmall Mammals	-0.15282	0.24061	-0.635	0.1342
hetsp.crowd.l:TrtLarge Mammals	-0.24897	0.29509	-0.844	0.0461

R<sup>2</sup>: Marginal 0.1297976, Conditional: 0.2760001

**Figure 2C & 2D**

Generalized linear mixed model fit by maximum likelihood (Laplace Approximation) ['glmerMod']

event ~ (consp.crowd.l + hetsp.crowd.l)\*Insecticide\_trt\*Fungicide\_trt + ht.l + (1|trans/plot) + (1|Census) + (1|Name)

Offset: offset\_time, Family: binomial (cloglog)

AIC	BIC	logLik	deviance	df.resid
919.7	1007.3	-442.8	885.7	1262

Random effects:

Groups	Name	Variance	Std.Dev.
Name	(Intercept)	5.967e-01	0.772472
plot:trans	(Intercept)	1.286e-06	0.001134
trans	(Intercept)	1.925e-01	0.438730
Census	(Intercept)	1.378e-01	0.371214

Number of obs: 1279, groups: Name, 104; plot:trans, 64; trans, 16; Census, 3

Fixed effects:

	Estimate	Std. Error	z value	P value
(Intercept)	0.4691	1.3745	0.341	0.1735
consp.crowd.l	-0.2561	0.2229	-1.149	0.1002
hetsp.crowd.l	0.5824	0.3063	1.901	0.0016
Insecticide_trtInsecticide	0.6283	1.1860	0.530	0.1187
Fungicide_trtFungicide	2.4639	1.1185	2.203	0.0003
ht.l	-1.3378	0.2644	-5.059	0.0000
consp.crowd.l:Insecticide_trtInsecticide	0.3515	0.3190	1.102	0.1187
hetsp.crowd.l:Insecticide_trtInsecticide	-0.3378	0.3905	-0.865	0.0514
consp.crowd.l:Fungicide_trtFungicide	0.4943	0.2680	1.845	0.0321
hetsp.crowd.l:Fungicide_trtFungicide	-0.9904	0.3639	-2.722	0.0000
Insecticide_trtInsecticide:Fungicide_trtFungicide	-1.4345	1.6066	-0.893	0.0217
consp.crowd.l:Insecticide_trtInsecticide:Fungicide_trtFungicide	-0.2623	0.4970	-0.528	0.3074
hetsp.crowd.l:Insecticide_trtInsecticide:Fungicide_trtFungicide	0.5513	0.5395	1.022	0.0212

R<sup>2</sup>: Marginal 0.09758026, Conditional: 0.2120702

**Figure 2E & 2F**

Generalized linear mixed model fit by maximum likelihood (Laplace Approximation) ['glmerMod']

**event ~ (consp.crowd.l + hetsp.crowd.l)\*Mammal\_trt\*Fungicide\_trt + ht.l + (1|trans/plot) + (1|Census) + (1|Name)**

Offset: offset\_time, Family: binomial (cloglog)

AIC	BIC	logLik	deviance	df.resid
935.0	1022.1	-450.5	901.0	1226

Random effects:

Groups	Name	Variance	Std.Dev.
Name	(Intercept)	0.8038940	0.89660
plot:trans	(Intercept)	0.0002498	0.01581
trans	(Intercept)	0.0879335	0.29654
Census	(Intercept)	0.1566762	0.39582

Number of obs: 1243, groups: Name, 97; plot:trans, 63; trans, 16; Census, 3

Fixed effects:

	Estimate	Std. Error	z value	P value
(Intercept)	0.8503	1.2153	0.700	0.0936
consp.crowd.l	-0.3092	0.3858	-0.801	0.1320
hetsp.crowd.l	0.2676	0.2780	0.963	0.0732
Mammal_trtControl	-0.6413	1.1479	-0.559	0.1611
Fungicide_trtFungicide	-0.7333	1.2612	-0.581	0.1299
ht.l	-1.2226	0.2652	-4.609	0.0000
consp.crowd.l:Mammal_trtControl	0.0479	0.4419	0.108	0.4314
hetsp.crowd.l:Mammal_trtControl	0.2895	0.3861	0.750	0.1063
consp.crowd.l:Fungicide_trtFungicide	0.3461	0.5019	0.689	0.1828
hetsp.crowd.l:Fungicide_trtFungicide	0.1946	0.4144	0.470	0.2003
Mammal_trtControl:Fungicide_trtFungicide	3.2729	1.6584	1.974	0.0002
consp.crowd.l:Mammal_trtControl:Fungicide_trtFungicide	0.1991	0.5636	0.353	0.3248
hetsp.crowd.l:Mammal_trtControl:Fungicide_trtFungicide	-1.2164	0.5397	-2.254	0.0001

R<sup>2</sup>: Marginal 0.0937560, Conditional: 0.2253659

**Figure 2G & 2H**

Generalized linear mixed model fit by maximum likelihood (Laplace Approximation) ['glmerMod']

**event ~ (consp.crowd.l + hetsp.crowd.l)\*Mammal\_trt\*Insecticide\_trt + ht.l + (1|trans/plot) + (1|Census) + (1|Name)**

Offset: offset\_time, Family: binomial (cloglog)

AIC	BIC	logLik	deviance	df.resid
827.6	913.2	-396.8	793.6	1119

Random effects:

Groups	Name	Variance	Std.Dev.
Name	(Intercept)	0.65694	0.8105
plot:trans	(Intercept)	0.07534	0.2745
trans	(Intercept)	0.05661	0.2379
Census	(Intercept)	0.14863	0.3855

Number of obs: 1136, groups: Name, 96; plot:trans, 63; trans, 16; Census, 3

Fixed effects:

	Estimate	Std. Error	z value	P value
(Intercept)	0.61495	1.20914	0.509	0.1599
consp.crowd.l	-0.16357	0.37649	-0.434	0.2620
hetsp.crowd.l	0.29223	0.28336	1.031	0.0933
Mammal_trtControl	-0.63457	1.18440	-0.536	0.1627
Insecticide_trtInsecticide	0.22666	1.28109	0.177	0.3172
ht.l	-1.17708	0.26355	-4.466	0.0088
consp.crowd.l:Mammal_trtControl	-0.15141	0.43841	-0.345	0.3442
hetsp.crowd.l:Mammal_trtControl	0.30470	0.39639	0.769	0.1207
consp.crowd.l:Insecticide_trtInsecticide	0.47073	0.45272	1.040	0.1139
hetsp.crowd.l:Insecticide_trtInsecticide	-0.14945	0.42451	-0.352	0.2145
Mammal_trtControl:Insecticide_trtInsecticide	0.26369	1.75745	0.150	0.3012
consp.crowd.l:Mammal_trtControl:Insecticide_trtInsecticide	-0.08393	0.55829	-0.150	0.4080
hetsp.crowd.l:Mammal_trtControl:Insecticide_trtInsecticide	-0.13909	0.58052	-0.240	0.2510

R<sup>2</sup>: Marginal 0.1023544, Conditional: 0.2161587

## GROWTH (Figure 3)

## Figure 3A &amp; 3B

rgr ~ (consp.crowd.l + hetsp.crowd.l)\*Trt + consp.crowd.l\*ht.l + (1|trans/plot) + (1|Census) + (1|Name)

AIC	BIC	logLik	deviance	df.resid
-52.6	60.5	48.3	-96.6	1242

## Random effects:

Groups	Name	Variance	Std.Dev.
Name	(Intercept)	0.0015318	0.03914
plot:trans	(Intercept)	0.0012265	0.03502
trans	(Intercept)	0.0002194	0.01481
Census	(Intercept)	0.0003325	0.01823
Residual		0.0519215	0.22786

Number of obs: 1264, groups: Name, 102; plot:trans, 76; trans, 16; Census, 3

## Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	0.244355	0.081453	3.000	0.0017
consp.crowd.l	0.101088	0.062005	1.630	0.0537
hetsp.crowd.l	-0.012587	0.012821	-0.982	0.1602
TrtFungi	0.073079	0.053215	1.373	0.0890
TrtInsects	-0.026165	0.045902	-0.570	0.2908
TrtSmall Mammals	0.004836	0.053745	0.090	0.4683
TrtLarge Mammals	0.010640	0.063735	0.167	0.4344
ht.l	-0.049128	0.017825	-2.756	0.0037
consp.crowd.l:TrtFungi	0.035116	0.016359	2.147	0.0172
consp.crowd.l:TrtInsects	-0.076360	0.019952	-3.827	0.0000
consp.crowd.l:TrtSmall Mammals	0.022800	0.026255	0.868	0.1953
consp.crowd.l:TrtLarge Mammals	0.045617	0.028913	1.578	0.0615
hetsp.crowd.l:TrtFungi	-0.021176	0.018445	-1.148	0.1283
hetsp.crowd.l:TrtInsects	0.015326	0.017328	0.884	0.1940
hetsp.crowd.l:TrtSmall Mammals	-0.013702	0.020557	-0.667	0.2542
hetsp.crowd.l:TrtLarge Mammals	-0.015021	0.023958	-0.627	0.2642
consp.crowd.l:ht.l	-0.036686	0.022349	-1.642	0.0514

R<sup>2</sup>: Marginal 0.04461308, Conditional: 0.1018705



**Figure 3C & 3D**

rgr ~ (consp.crowd.l+hetsp.crowd.l)\*Insecticide\_trt\*Fungicide\_trt + (1|trans/plot) + (1|Census)

AIC	BIC	logLik	deviance	df.resid
-72.3	7.8	52.2	-104.3	1088

## Random effects:

Groups	Name	Variance	Std.Dev.
plot:trans	(Intercept)	1.322e-03	0.036361
trans	(Intercept)	0.000e+00	0.000000
Census	(Intercept)	5.363e-05	0.007323
Residual		5.220e-02	0.228484

Number of obs: 1104, groups: plot:trans, 63; trans, 16; Census, 3

## Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	0.080709	0.051443	1.569	0.0603
consp.crowd.l	0.007668	0.020344	0.377	0.3526
hetsp.crowd.l	-0.009346	0.019814	-0.472	0.3136
Insecticide_trtInsecticide	-0.022885	0.070163	-0.326	0.3775
Fungicide_trtFungicide	-0.129339	0.078020	-1.658	0.0474
consp.crowd.l:Insecticide_trtInsecticide	0.064706	0.028717	2.253	0.0118
hetsp.crowd.l:Insecticide_trtInsecticide	0.001484	0.026950	0.055	0.4765
consp.crowd.l:Fungicide_trtFungicide	-0.030602	0.024878	-1.230	0.1113
hetsp.crowd.l:Fungicide_trtFungicide	0.041983	0.028113	1.493	0.0694
Insecticide_trtInsecticide:Fungicide_trtFungicide	0.095869	0.107146	0.895	0.1827
consp.crowd.l:Insecticide_trtInsecticide:Fungicide_trtFungicide	-0.107962	0.044959	-2.401	0.0091
hetsp.crowd.l:Insecticide_trtInsecticide:Fungicide_trtFungicide	-0.003915	0.039177	-0.100	0.4534

R<sup>2</sup>: Marginal 0.03257394, Conditional: 0.05741377

**Figure 3E & 3F**

rgr ~ (consp.crowd.l + hetsp.crowd.l)\*Mammal\_trt\*Fungicide\_trt + consp.crowd.l:ht.l + (1|trans/plot) + (1|Census)

AIC	BIC	logLik	deviance	df.resid
-177.2	-87.8	106.6	-213.2	1041

Random effects:

Groups	Name	Variance	Std.Dev.
plot:trans	(Intercept)	6.587e-04	2.567e-02
trans	(Intercept)	1.942e-19	4.407e-10
Census	(Intercept)	0.000e+00	0.000e+00
Residual		4.731e-02	2.175e-01

Number of obs: 1059, groups: plot:trans, 62; trans, 16; Census, 3

Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	0.227055	0.093625	2.425	0.0078
consp.crowd.l	0.023428	0.068219	0.343	0.3696
hetsp.crowd.l	-0.001059	0.022765	-0.047	0.4854
Mammal_trtControl	0.058429	0.072988	0.801	0.2162
Fungicide_trtFungicide	-0.144450	0.077651	-1.860	0.0311
ht.l	-0.044654	0.017977	-2.484	0.0058
consp.crowd.l:Mammal_trtControl	0.038316	0.034057	1.125	0.1337
hetsp.crowd.l:Mammal_trtControl	-0.032813	0.028098	-1.168	0.1225
consp.crowd.l:Fungicide_trtFungicide	0.049489	0.035786	1.383	0.0838
hetsp.crowd.l:Fungicide_trtFungicide	0.040594	0.028648	1.417	0.0800
Mammal_trtControl:Fungicide_trtFungicide	0.036149	0.105986	0.341	0.3584
consp.crowd.l:ht.l	-0.020407	0.021932	-0.930	0.1797
consp.crowd.l:Mammal_trtControl:Fungicide_trtFungicide	-0.084653	0.042521	-1.991	0.0250
hetsp.crowd.l:Mammal_trtControl:Fungicide_trtFungicide	-0.006426	0.038640	-0.166	0.4268

R<sup>2</sup>: Marginal 0.03377294, Conditional: 0.04704284

**Figure 3G & 3H**

rgr ~ (consp.crowd.l + hetsp.crowd.l)\*Mammal\_trt\*Isecticide\_trt + consp.crowd.l:ht.l + (1|trans/plot) + (1|Census)

AIC	BIC	logLik	deviance	df.resid
77.8	165.8	-20.9	41.8	962

## Random effects:

Groups	Name	Variance	Std.Dev.
plot:trans	(Intercept)	0.002101	0.04584
trans	(Intercept)	0.000000	0.00000
Census	(Intercept)	0.000000	0.00000
Residual		0.059523	0.24397

Number of obs: 980, groups: plot:trans, 62; trans, 16; Census, 3

## Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	0.267508	0.108024	2.476	0.0092
consp.crowd.l	-0.015406	0.080412	-0.192	0.4277
hetsp.crowd.l	-0.007207	0.026495	-0.272	0.4052
Mammal_trtControl	0.055085	0.085367	0.645	0.2505
Isecticide_trtIsecticide	0.021851	0.099334	0.220	0.4142
ht.l	-0.053463	0.020602	-2.595	0.0073
consp.crowd.l:Mammal_trtControl	0.044414	0.039175	1.134	0.1360
hetsp.crowd.l:Mammal_trtControl	-0.031907	0.032913	-0.969	0.1625
consp.crowd.l:Isecticide_trtIsecticide	0.061138	0.041196	1.484	0.0731
hetsp.crowd.l:Isecticide_trtIsecticide	-0.022430	0.036204	-0.620	0.2670
Mammal_trtControl:Isecticide_trtIsecticide	-0.051828	0.125568	-0.413	0.3371
consp.crowd.l:ht.l	-0.007835	0.026019	-0.301	0.3767
consp.crowd.l:Mammal_trtControl:Isecticide_trtIsecticide	-0.004998	0.051990	-0.096	0.4680
hetsp.crowd.l:Mammal_trtControl:Isecticide_trtIsecticide	0.028432	0.046730	0.608	0.2714

R<sup>2</sup>: Marginal 0.0289217, Conditional: 0.06202766

## DIVERSITY (Figure 4)

## Figure 4A

Linear mixed model fit by maximum likelihood ['lmerMod']

**diversity.shannon.exp ~ Trt \* Census + (1|location)**

AIC	BIC	logLik	deviance	df.resid
1013.9	1072.4	-490.0	979.9	213

Random effects:

Groups	Name	Variance	Std.Dev.
location	(Intercept)	10.34	3.215
	Residual	1.46	1.208

Number of obs: 230, groups: location, 78

Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	5.64550	0.39082	14.445	0.0000
TrtFungi	-1.35370	0.77144	-1.755	0.0379
TrtInsects	0.01862	0.77144	0.024	0.4876
TrtSmall Mammals	-0.59626	0.77490	-0.769	0.2183
TrtLarge Mammals	0.36743	0.96045	0.383	0.3574
Census10	0.83778	0.19714	4.250	0.0001
Census11	1.31873	0.19705	6.692	0.0000
TrtFungi:Census10	0.23602	0.38520	0.613	0.2702
TrtInsects:Census10	0.08293	0.38520	0.215	0.4183
TrtSmall Mammals:Census10	0.23721	0.39209	0.605	0.2753
TrtLarge Mammals:Census10	-0.06968	0.48767	-0.143	0.4431
TrtFungi:Census11	1.05605	0.38516	2.742	0.0027
TrtInsects:Census11	-0.01793	0.38516	-0.047	0.4768
TrtSmall Mammals:Census11	1.04472	0.39205	2.665	0.0039
TrtLarge Mammals:Census11	0.03173	0.48735	0.065	0.4653
R <sup>2</sup> : Marginal	0.0575303	Conditional:	0.883338	

**Figure 4B**

diversity.shannon.exp ~ Insecticide\_trt\*Fungicide\_trt\*Census + (1|location)

AIC	BIC	logLik	deviance	df.resid
875.1	920.6	-423.5	847.1	177

Random effects:

Groups	Name	Variance	Std.Dev.
location	(Intercept)	11.070	3.327
	Residual	1.845	1.358

Number of obs: 191, groups: location, 64

Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	6.235201	0.904667	6.892	0.0000
Insecticide_trtInsecticide	-0.608326	1.274994	-0.477	0.3166
Fungicide_trtFungicide	0.763998	1.274994	0.599	0.0168
Census10	0.607132	0.491753	1.235	0.0105
Census11	0.280561	0.491753	0.571	0.2793
Insecticide_trtInsecticide:Fungicide_trtFungicide	-0.391900	1.799996	-0.218	0.1573
Insecticide_trtInsecticide:Census10	0.147724	0.687319	0.215	0.3393
Insecticide_trtInsecticide:Census11	1.056099	0.687319	1.537	0.0358
Fungicide_trtFungicide:Census10	-0.005367	0.687319	-0.008	0.4612
Fungicide_trtFungicide:Census11	-0.017885	0.687319	-0.026	0.4900
Insecticide_trtInsecticide:Fungicide_trtFungicide:Census10	-0.195037	0.966220	-0.202	0.3725
Insecticide_trtInsecticide:Fungicide_trtFungicide:Census11	-0.236325	0.966220	-0.245	0.3820

R<sup>2</sup>: Marginal 0.02197799, Conditional: 0.8603057

**Figure 4C**

**diversity.shannon.exp ~ Mammal\_trt\*Fungicide\_trt\*Census + (1|location)**

AIC	BIC	logLik	deviance	df.resid
866.7	912.0	-419.3	838.7	174

Random effects:

Groups	Name	Variance	Std.Dev.
location	(Intercept)	9.904	3.147
	Residual	2.017	1.420

Number of obs: 188, groups: location, 63

Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	4.67784	0.89148	5.247	0.0000
Mammal_trtControl	1.54807	1.24580	1.243	0.2564
Fungicide_trtFungicide	1.37454	1.24088	1.108	0.1137
Census10	1.08705	0.51858	2.096	0.0130
Census11	2.38857	0.51858	4.606	0.0000
Mammal_trtControl:Fungicide_trtFungicide	-0.60125	1.74417	-0.345	0.3789
Mammal_trtControl:Census10	-0.47063	0.73025	-0.644	0.1586
Mammal_trtControl:Census11	-2.09872	0.73025	-2.874	0.0007
Fungicide_trtFungicide:Census10	-0.07365	0.72183	-0.102	0.4558
Fungicide_trtFungicide:Census11	0.47327	0.72183	0.656	0.2371
Mammal_trtControl:Fungicide_trtFungicide:Census10	0.05899	1.01858	0.058	0.4376
Mammal_trtControl:Fungicide_trtFungicide:Census11	-0.50045	1.01858	-0.491	0.3031

R<sup>2</sup>: Marginal 0.07725853, Conditional: 0.843878

**Figure 4D**

**diversity.shannon.exp ~ Insecticide\_trt\*Mammal\_trt\*Census + (1|location)**

AIC	BIC	logLik	deviance	df.resid
820.8	866.1	-396.4	792.8	174

Random effects:

Groups	Name	Variance	Std.Dev.
location	(Intercept)	8.969	2.995
	Residual	1.478	1.216

Number of obs: 188, groups: location, 63

Fixed effects:

	Estimate	Std. Error	t value	P value
(Intercept)	4.67784	0.83453	5.605	0.0000
Insecticide_trtInsecticide	1.19449	1.16162	1.028	0.1428
Mammal_trtControl	1.55783	1.16549	1.337	0.2618
Census10	1.08705	0.44393	2.449	0.0064
Census11	2.38857	0.44393	5.380	0.0000
Insecticide_trtInsecticide:Mammal_trtControl	-1.80328	1.63223	-1.105	0.2762
Insecticide_trtInsecticide:Census10	0.22410	0.61793	0.363	0.3448
Insecticide_trtInsecticide:Census11	0.41220	0.61793	0.667	0.2450
Mammal_trtControl:Census10	-0.48039	0.62517	-0.768	0.1358
Mammal_trtControl:Census11	-2.10848	0.62517	-3.373	0.0001
Insecticide_trtInsecticide:Mammal_trtControl:Census10	-0.07591	0.87198	-0.087	0.4864
Insecticide_trtInsecticide:Mammal_trtControl:Census11	0.64437	0.87198	0.739	0.2013

R<sup>2</sup>: Marginal 0.08041946, Conditional: 0.8698903