

1 **Anthropogenic Land-Use Legacies Underpin Climate Change-Related**
2 **Risks to Forest Ecosystems**

3 Albert Vilà-Cabrera^{1,2,3,‡,*}, Julen Astigarraga^{3,‡}, Alistair S. Jump², Miguel A. Zavala³,
4 Francisco Seijo⁴, Dominik Sperlich⁵ and Paloma Ruiz-Benito^{3,6,‡}

5

6 ¹CREAF, Bellaterra, Catalonia, Spain

7 ²Biological and Environmental Sciences, Faculty of Natural Sciences, University of
8 Stirling, Scotland, UK

9 ³Universidad de Alcalá, Grupo de Ecología y Restauración Forestal, Departamento de
10 Ciencias de la Vida, 28805 Alcalá de Henares (Madrid), Spain

11 ⁴Instituto de Empresa, School of Global and Public Affairs

12 ⁵Department of Forestry Economics and Forest Planning, Faculty of Environment and
13 Natural Resources, University of Freiburg, Germany

14 ⁶Universidad de Alcalá, Grupo de Investigación en Teledetección Ambiental,
15 Departamento de Geología, Geografía y Medio Ambiente, 28801 Alcalá de Henares
16 (Madrid), Spain

17

18 [‡]These authors led the manuscript

19

20 *Correspondence: albert.vilac@uah.es (A. Vilà-Cabrera)

21

22

23

24

25

26

27

28

29

30

31

32

33

34 **Abstract**

35 Forest ecosystems with long-lasting human imprints can emerge worldwide as outcome
36 of land-use cessation. However, the interaction of these anthropogenic legacies with
37 climate change impacts on forests is not well understood. Here, we set out how
38 anthropogenic land-use legacies that persist in forest properties following alterations in
39 forest distribution, structure, and composition can interact with climate change stressors.
40 We propose a risk-based framework to identify anthropogenic legacies of land-uses in
41 forest ecosystems and quantify the impact of their interaction with climate-related stress
42 on forest responses. Considering anthropogenic land-use legacies alongside
43 environmental drivers of forest ecosystem dynamics will improve our predictive capacity
44 of climate-related risks to forests and our ability to promote ecosystem resilience to
45 climate change.

46

47 **Anthropogenic land-use legacies can exacerbate the impacts of climate change on**
48 **forest ecosystems**

49 Humans have exploited forest ecosystems for millennia [1–3]. Given the magnitude of
50 land-cover changes, forest management practices and plantation establishment over the
51 last few centuries, recent humanity’s footprint today shapes more than 70% of the world’s
52 forests [4]. While high-intensity **anthropogenic land-uses** (*see glossary*) prevail in
53 tropical and boreal regions [5,6], **land-use cessation** also spreads unevenly worldwide
54 since the post-1950’s **great acceleration** [7], from large areas and longer time periods in
55 the Global North to scattered less-extensive patches and shorter time periods in the Global
56 South [8]. As a consequence, forest ecosystems with long-lasting human impacts (Fig. 1)
57 can emerge across multiple biomes [9,10]. Although broadly recognised to influence
58 forest dynamics [11], there is still a lack of understanding on how these human imprints
59 may exacerbate the vulnerability of forest ecosystems to climate change [12–14].
60 However, such understanding is essential to better predict **impacts of climate change on**
61 **forests** and define mitigation and adaptation priorities.

62 The response of forest ecosystems to climate change strongly depends on historical
63 factors [15]. Anthropogenic land-uses can affect the stability of forest dynamics to climate
64 by altering forest main patterns in forest distribution [16], structure [12], and composition
65 [17] (Fig. 1). These alterations may result in persistent impacts of land-uses on **forest**
66 **properties** at different levels (genetic, population, community, and ecosystem). Despite

67 their relevance, the effects of anthropogenic land-uses are rarely considered alongside
68 environmental drivers, often hindering our ability to predict climate change impacts on
69 forests [18]. Given that climate change stressors and other anthropogenic perturbations
70 interact and today strongly impact biological communities [19], quantifying the long-
71 lasting influence of anthropogenic land-uses is essential to better predict forest ecosystem
72 responses to climate change [12,18]. This understanding might thus support the
73 anticipation of climate-related disturbance impacts on forests, such as the occurrence of
74 large-scale forest die-off events, mega-fires and forest degradation [12,20–22].
75 Consequently, a research agenda focused on identifying hotspots of increased
76 vulnerability in areas dominated by **human-modified forests** is of critical importance if
77 we are to develop and adopt mitigation and adaptation strategies to improve forest
78 **resilience to climate change** [10,23,24]. A risk-based framework using the concept of
79 **anthropogenic land-use legacy** could provide such an approach.

80 Disturbance legacies are ecological properties that persist in biological communities
81 following disturbance, shaping the response capacity of ecosystems to subsequent
82 disturbance [25,26]. Disturbance legacies may become detrimental and increase
83 ecosystem vulnerability to environmental stressors, particularly when legacies result from
84 anthropogenic disturbance. For example, historical management can facilitate tree
85 biomass development such that forest stands might become structurally mismatched to
86 water availability. During periods of increased drought stress, this process can lead to
87 episodes of forest dieback [12]. Therefore, the effective study of anthropogenic land-use
88 legacies depends on a deep knowledge of the history, ecological dynamics, and
89 environmental factors that will ultimately determine forest ecosystem responses to
90 climate change [11]. However, the lack of guidelines to quantify how anthropogenic land-
91 use legacies increase climate change-related risks to forests makes it difficult to anticipate
92 the outcomes of this interaction.

93 Here, we set out how anthropogenic land-use legacies that persist in forest properties
94 following the alteration in forest distribution, structure, and composition can heighten the
95 impact of climatic stressors on forest ecosystems. Based on existing literature, we first
96 summarise hypotheses on how anthropogenic land-use legacies in forest ecosystems can
97 interact with climate-related stress and lead to detrimental impacts on forests. We then
98 introduce a risk-based framework to first identify anthropogenic land-use legacies and
99 then quantify their impacts on forests and their interaction with climate change. Finally,

100 we apply the framework in an example to show that detrimental anthropogenic land-use
101 legacies contribute to large-scale risk of tree mortality at the driest distribution edge of
102 Scots pine (*Pinus sylvestris* L.). We focus on detrimental anthropogenic land-use legacies
103 given the urgency of reducing climate change-related risks to forests. However, the
104 proposed framework is flexible and general enough to not only anticipate detrimental
105 impacts of anthropogenic land-use legacies on forests in any given context, but also
106 accommodate and assess potential beneficial legacy effects. Therefore, we provide
107 scientists, managers, and policymakers with an approach to identify risk and promote
108 ecosystem resilience to climate change in regions dominated by human-modified forests.

109 **Processes underlying detrimental effects of anthropogenic land-use legacies in forest** 110 **ecosystems**

111 To identify processes leading to legacy effects of anthropogenic land-uses that can
112 exacerbate climate change impacts on forests, we surveyed scientific studies that reported
113 detrimental effects of land-uses on current forest ecosystem responses to climate change.
114 Across studies, we identified processes driven by anthropogenic land-uses that, following
115 the cessation of the land-use, shape the current forest distribution, structure, and
116 composition, leading to persistent legacies in forest properties. Hypotheses on detrimental
117 effects of anthropogenic land-use legacies are framed considering the outcomes of their
118 interaction with climatic stressors (Table 1).

119 ***Forest distribution: Tree species re-expansion towards increased climatic stress***

120 Although the emergence of **secondary forests** contributes to the recovery of ecosystem
121 functionality to some extent [10,27], ongoing re-expansion of tree species distributions is
122 occurring under a changing climate relative to past conditions. If anthropogenic land-uses
123 (i.e., deforestation) are clustered in areas where species are climatically stressed,
124 subsequent secondary forest regrowth may occur at the limit of species climatic
125 tolerances. An increased climatic stress relative to species climatic niches can thus
126 constrain forest recovery and heighten the impact of climatic stressors on forests (Table
127 1) [16,28–32]. Similar responses may occur when management practices lead to partial
128 tree canopy loss. Following the cessation of management practices, species can have re-
129 established at the limit of their climatic tolerances [31,33] (Table 1). Shifts in canopy
130 cover can also exacerbate the climatic stress experienced by understory trees [34] (Table
131 1). Anthropogenic land-use legacies may emerge through the establishment of tree
132 plantations, which can result in a broadening of tree species distributions where planting

133 has occurred beyond the species' natural range. Planted tree populations growing at the
134 limit or outside of the range of species climatic tolerances may be particularly vulnerable
135 to increased environmental stress (Table 1) [24,35–38].

136 ***Forest structure: Exacerbated conspecific negative density-dependence and increased***
137 ***disturbance impact***

138 Altered forest structure is characteristic of human-modified forests. Such changes can
139 derive from the establishment of high-density, structurally homogeneous tree plantations,
140 forest management practices, or the regrowth of secondary forests. Structural alterations
141 may persist over time once land-use cessation has occurred (Fig. 1). Alteration of tree
142 size distribution may result in exacerbated **conspecific negative density-dependence**
143 and, therefore, increased impacts of climate change on forests (e.g., higher tree mortality
144 under drought stress) (Table 1) [12,32,36,37,39,40]. Similar forest structural shifts may
145 increase the impact of climate-related wildfires on forest ecosystems (Table 1)
146 [20,21,24,41].

147 ***Forest composition: Shifts in functional traits, genetics, and interspecific interactions***
148 ***mediate reduced population and community resilience***

149 Human-induced alterations in forest composition may lead to persistent anthropogenic
150 land-use legacies in population- and community-level functional traits, genetic
151 composition, and interspecific interactions, in such a way that forest resilience to climate
152 change can be compromised. Following the cessation of forest management practices
153 (e.g., clearcutting, thinning, pruning, pollarding), their imprint may persist over time,
154 becoming a source of intraspecific functional variation (Fig. 1). These human-modified
155 forests may hold trees with functional traits that display reduced resilience to cope with
156 climate change stressors (Table 1) [12,39,41–47]. Such phenotypic constraints may also
157 arise in forests growing on altered soils by former agriculture and pasture (e.g., altered
158 soil fertility) (Table 1) [48,49]. Anthropogenic land use-derived shifts in forest
159 composition may generate forest populations with altered genetic composition [50,51]
160 that may decrease their ability to cope with environmental stressors (Table 1) [24,52–60].
161 Land use-driven species replacements may result in human-modified forests holding
162 novel communities that display altered ability to respond to climate change due to altered
163 functional trait composition (Table 1) [33,39,61]. Loss of forest resilience to climatic
164 stressors may also occur if anthropogenic land-use legacies affect interspecific
165 interactions such as mutualisms (Table 1) [48,62–64]. Species replacements may also

166 lead to increased performance of antagonists under changing climatic conditions,
167 including tree heterospecific competitors, invasive species, herbivores, and pathogens
168 (Table 1) [42,54,65–73].

169 ***Limitations in the current state of knowledge of anthropogenic land-use legacies in***
170 ***forest ecosystems***

171 Current studies suggest that anthropogenic land-use legacies can increase the
172 vulnerability of forest ecosystems to climate change, but anthropogenic land-use legacies
173 are rarely considered in experimental designs in research (see Table 1). This knowledge
174 limitation is probably triggered by the scarcity of reliable data on anthropogenic land-
175 uses, making it difficult to identify and predict detrimental legacy effects. Furthermore,
176 as the effects of anthropogenic land-use legacies and climatic stressors interact, the
177 detrimental outcomes may vary between different ecological and geographic contexts
178 (Table 1), making comparisons among systems (e.g., regions, species, legacy types, etc.)
179 difficult. However, risk-based approaches can support the predictive understanding of
180 interaction outcomes between global change stressors across contexts [74]. To advance
181 towards forest adaptation strategies and support forest resilience to climate change, we
182 need to better understand where and to what extent anthropogenic land-use legacies
183 increase climate change-related risks to forests.

184 **A risk-based framework using the concept of anthropogenic land-use legacy**

185 To overcome the limitations in knowledge of potential detrimental effects of
186 anthropogenic land-use legacies due to the lack of experimental designs, data, and
187 contexts, we incorporate three major knowledge needs to build up a risk-based framework
188 (Fig. 2, Key Figure). First, we point out the need to characterise forest properties
189 considering human land-use impact and to identify high-impact anthropogenic land-use
190 legacies (Fig. 2A). For instance, current data on forest distribution, structure, and
191 composition can be used as proxies of land use-driven alteration of forest properties
192 together with available information on anthropogenic land-uses (Box 1). Second, we
193 recommend that alterations of anthropogenic land-use legacies are explicitly incorporated
194 the into empirical study designs, so that they are considered alongside environmental
195 drivers of contemporary forest dynamics. Therefore, hypotheses on detrimental
196 anthropogenic land-use legacies (Table 1) can be tested in the context of a particular study
197 system (Fig. 2B; Box 1). Finally, the potential outcomes of interactions between
198 anthropogenic land-use legacies and climatic stressors can be better quantified and

199 understood using the risk components and the main forest patterns that can be altered (i.e.,
200 distribution, structure, and composition), as this can improve our ability to identify and
201 anticipate detrimental ecological effects on current forest properties (Fig. 2B-C; Box 1).

202 To define the risk components summarised in *Fig. 2 (Key Figure)* we adapt the IPCC
203 framework [19] to forest ecosystems. When forest anthropogenic land-use legacies are
204 linked to the components of risk (exposure, sensitivity, and adaptive capacity), responses
205 of forest ecosystems to climatic stressors can be framed and understood through the
206 interaction between anthropogenic land use-driven changes in distribution, forest
207 structure, and composition.

208 A reduction in forest distribution followed by re-expansion of species ranges after land-
209 use cessation may result in increased climate-related exposure (Fig. 2C). For example,
210 secondary forests in the Brazilian Amazon are distributed more toward drier and more
211 seasonal climatic conditions than the biome average because past deforestation was
212 concentrated in these areas. Species re-expansions under dry climatic conditions are
213 constraining the potential contribution of secondary forests to carbon sequestration [28].
214 The alteration in forest structure may result in increased sensitivity to climatic stressors
215 (Fig. 2C). When management practices are ceased, their legacies can endure, for example
216 in the form of increased competition in overcrowded even-aged stands, which may
217 exacerbate processes of conspecific negative density-dependence and thus impacts of
218 climatic stressors on forests [12,41,43].

219 Alteration in forest distribution and structure can result in anthropogenic land-use
220 legacies that increase risk through exposure and sensitivity (Fig. 2C), such as in the case
221 of tree plantations. For example, in the Carpathians where pine plantations have been
222 established outside their natural distributions, pine species perform worse than the native
223 conifers in response to drought stress [35]. Decreased performance of plantations may
224 reduce the delivery of ecosystem services, especially in drier areas worldwide [38].
225 Anthropogenic land-use legacies derived from shifts in forest distribution and structure
226 can also result in increased megafire risk, as reported in Australian forests where past
227 logging generated extensive areas characterised by densely stocked forests that today
228 influence fire dynamics [21].

229 Increased climate change-related risk to forests may occur by the reduction of adaptive
230 capacity through alterations in forest composition. At the population level, the cessation
231 of anthropogenic land-uses may result in the emergence of anthropogenic land-use
232 legacies toward more vulnerable tree phenotypes, such as in form of high abundance of
233 slow-growing weak trees [42] or increased canopy dominance of younger trees which
234 have root traits less able to cope with climatic stressors [47]. Legacy-driven population-
235 level functional shifts may also lead to increased structural and physiological constraints
236 that exacerbate hydraulic failure and fire impact [39,44–46]. For example, past high-
237 intensity management practices might underly increased drought-induced mortality of
238 larger canopy trees in temperate forests [46]. Anthropogenic land-use legacies in the
239 population genetic properties can also lead to decreased adaptive capacity to climate
240 change. For example, widespread tree planting within a species' native range may result
241 in forest dieback if seed material is translocated among lineages with differing climatic
242 niches [58]. In addition, gene flow from planted non-local genotypes into native
243 populations can result in landscape-level genetic homogenisation [55]. Although still
244 highly uncertain, shifts in forest genetic properties have the potential of inducing long-
245 term effects on the adaptive capacity of forest ecosystems [52,59,60].

246 Anthropogenic land-use legacies affecting interspecific composition in functional traits
247 may increase risks to forests driven by shifts in their adaptive capacity. For instance, shifts
248 towards reduced bark thickness and wood density compromises trait-mediated capacity
249 of forests to withstand fire and drought in tropical and temperate regions [39,61].
250 Anthropogenic land-use legacies that affect interspecific interactions can also influence
251 the adaptive capacity of forests to climatic stressors. For example, secondary forests may
252 hold distinct and simpler ectomycorrhizal communities because of increased nitrogen and
253 phosphorous availability due to former arable and pasture activities. This alteration may
254 reduce the capacity of trees to cope with drought stress [48,63,64]. High occurrence of
255 tree stumps or increased stand density and structural homogeneity may alter interspecific
256 interactions and favour, for instance, the spread of pathogen and insect outbreaks
257 [65,70,73] that in interaction with climate-related stress may exacerbate negative impacts
258 on forests [42,72].

259 The proposed risk-based framework can improve our predictive understanding of forest
260 dynamics and climate-related risks to forests, thus supporting adaptation strategies to
261 foster long-term forest resilience to climate change in regions dominated by human-

262 modified forests. As witnessed in Scots pine (*Pinus sylvestris* L.) in the Iberian Peninsula,
263 anthropogenic land-use legacies derived from altered forest distribution, structure, and
264 composition contribute to large-scale patterns of climate-driven tree mortality risk (Box
265 1).

266 **Concluding remarks**

267 Increased climate change-related risk to forest ecosystems can be intense and extensive
268 in areas dominated by human-modified forests, particularly if anthropogenic land-use
269 legacies interact with climatic stressors in such a way that their detrimental ecological
270 effects act synergistically. We recommend that future research explicitly incorporates the
271 impact of anthropogenic land-use legacies into experimental designs and investigates how
272 forest exposure, sensitivity, and adaptive capacity to climatic stressors are modified by
273 anthropogenic land-use legacies (Fig. 2, Key Figure). Such an approach can be of key
274 importance to identify forest areas at high risk from climate change, and understand and
275 anticipate climate change impacts on forest dynamics, biodiversity and ecosystem
276 functioning, providing an opportunity to reduce risks to forests and improve forest
277 resilience [10,23,24,75] (Box 1) (see **Outstanding Questions**). To apply the proposed
278 framework and better predict the impacts of climate change on forests, we propose using
279 current forest properties as proxies of anthropogenic land-use legacies using available
280 data such as forest inventories and forest historical information (Box 1), and emerging
281 high-resolution data as remote sensing technologies continue to develop [76]. In addition,
282 understanding how anthropogenic land-use legacies shape forest genetics, functional
283 traits, and interspecific interactions deserves urgent attention as these forest properties
284 ultimately underpin responses of forest ecosystems to changing climatic conditions.

285 Importantly, although our approach here focusses on anthropogenic land-use legacies that
286 are predominantly detrimental, the proposed framework allows the incorporation of
287 (sometimes counteracting) beneficial legacy effects. For example, younger canopy-
288 dominant trees can show larger growth reductions during drought, but this response can
289 be counteracted by a quick recovery from impact [47]. As the proportion of younger trees
290 increases across many forested regions (as a legacy of anthropogenic land-uses), the
291 application of the framework has the potential to support adaptive management to reduce
292 short-term risk of forest loss and enhance long-term C stocks [47]. Improved forest
293 resilience can also be attained by considering the benefits of sustainable management
294 practices and local knowledge in the framework application, such as traditional fire use

295 and forest management [52,77]. Finally, while this framework centres on an ecological
296 approach, emerging anthropogenic land-use legacies are inextricably dependent on
297 previous socio-economic contexts [8,78]. Therefore, a key step towards mitigation and
298 adaptation goals will involve the collaboration of natural and social scientists with local
299 stakeholders and policymakers in decision-making processes from local to global scales.
300 This cooperation will be pivotal for the success of the post-2020 UN Biodiversity
301 Framework and the UN Decade on Ecosystem Restoration.

302 **Acknowledgements**

303 We acknowledge the insight comments and suggestions of three anonymous reviewers to
304 improve the manuscript. A.V-C was supported by a Juan de la Cierva-Incorporación
305 fellowship [IJC2018-038508-I] from the Ministry of Science and Innovation (Spain) and
306 the 50th Anniversary Fellowship program of the University of Stirling (Scotland, UK).
307 J.A was supported by a FPI fellowship of the Department of Education of the Basque
308 Government. P.R-B was supported by the Community of Madrid Region under the
309 framework of the multi-year Agreement with the University of Alcalá (Stimulus to
310 Excellence for Permanent University Professors, EPU-INV/2020/010). D.S
311 acknowledges funding of the Federal Ministry of Food and Agriculture (BMEL) based
312 on a decision of the Parliament of the Federal Republic of Germany. We also
313 acknowledge support from the grants VERDAT (Organismo Autónomo de Parques
314 Nacionales, Ref. 2794/2021) from the Ministry for Ecological Transition and
315 Demographic Challenge (Spain), and LARGE (PID2021-123675OB-C41) from the
316 Ministry of Science and Innovation (Spain).

317

318 **References**

- 319 1 Ellis, E.C. and Ramankutty, N. (2008) Putting people in the map: Anthropogenic
320 biomes of the world. *Front. Ecol. Environ.* 6, 439–447
- 321 2 Mottl, O. *et al.* (2021) Global acceleration in rates of vegetation change over the
322 past 18,000 years. *Science (80-.)*. 372, 860–864
- 323 3 Grantham, H.S. *et al.* (2020) Anthropogenic modification of forests means only
324 40% of remaining forests have high ecosystem integrity. *Nat. Commun.* 11, 5978
- 325 4 FAO (2020) *Global Forest Resources Assessment 2020 - Key Findings*,
- 326 5 Taubert, F. *et al.* (2018) Global patterns of tropical forest fragmentation. *Nature*
327 554, 519–522

- 328 6 Gauthier, S. *et al.* (2015) Boreal forest health and global change. *Science* (80-.).
329 349, 819–822
- 330 7 Steffen, W. *et al.* (2015) The trajectory of the Anthropocene: The Great
331 Acceleration. *Anthr. Rev.* 2, 81–98
- 332 8 Winkler, K. *et al.* (2021) Global land use changes are four times greater than
333 previously estimated. *Nat. Commun.* 12, 2501
- 334 9 Meyfroidt, P. and Lambin, E.F. (2011) Global Forest Transition: Prospects for an
335 End to Deforestation. *Annu. Rev. Environ. Resour.* 36, 343–371
- 336 10 García, C. *et al.* (2020) Managing forest regeneration and expansion at a time of
337 unprecedented global change. *J. Appl. Ecol.* 57, 2310–2315
- 338 11 Foster, D. *et al.* (2003) The importance of land-use legacies to ecology and
339 conservation. *Bioscience* 53, 77–88
- 340 12 Jump, A.S. *et al.* (2017) Structural overshoot of tree growth with climate
341 variability and the global spectrum of drought-induced forest dieback. *Glob.*
342 *Chang. Biol.* 23, 3742–3757
- 343 13 Bürgi, M. *et al.* (2017) Legacy Effects of Human Land Use: Ecosystems as
344 Time-Lagged Systems. *Ecosystems* 20, 94–103
- 345 14 Verheyen, K. (2022) Land-use legacies predispose the response of trees to
346 drought in restored forests. *Glob. Chang. Biol.* 28, 1204–1211
- 347 15 McDowell, N.G. *et al.* (2020) Pervasive shifts in forest dynamics in a changing
348 world. *Science* (80-.). 368, eaaz9463
- 349 16 Goring, S.J. and Williams, J.W. (2017) Effect of historical land-use and climate
350 change on tree-climate relationships in the upper Midwestern United States. *Ecol.*
351 *Lett.* 20, 461–470
- 352 17 Filgueiras, B.K.C. *et al.* (2021) Winner–Loser Species Replacements in Human-
353 Modified Landscapes. *Trends Ecol. Evol.* 36, 545–555
- 354 18 Vilà-Cabrera, A. *et al.* (2019) Refining predictions of population decline at
355 species’ rear edges. *Glob. Chang. Biol.* 25, 1549–1560
- 356 19 IPCC, 2022: *Climate Change 2022: Impacts, Adaptation, and Vulnerability.*
357 *Contribution of Working Group II to the Sixth Assessment Report of the*
358 *Intergovernmental Panel on Climate Change,*
- 359 20 Hermoso, V. *et al.* (2021) Tree planting: A double-edged sword to fight climate
360 change in an era of megafires. *Glob. Chang. Biol.* 27, 3001–3003
- 361 21 Lindenmayer, D.B. *et al.* (2022) Logging elevated the probability of high-

362 severity fire in the 2019–20 Australian forest fires. *Nat. Ecol. Evol.* 6, 533–535
 363 22 Silva Junior, C.H.L. *et al.* (2020) Persistent collapse of biomass in Amazonian
 364 forest edges following deforestation leads to unaccounted carbon losses. *Sci. Adv.*
 365 6,
 366 23 Henne, P.D. *et al.* (2015) Reviving extinct Mediterranean forest communities
 367 may improve ecosystem potential in a warmer future. *Front. Ecol. Environ.* 13,
 368 356–362
 369 24 Gómez-González, S. *et al.* (2022) Moving towards the ecological intensification
 370 of tree plantations. *Trends Plant Sci.* 27, 637–645
 371 25 Johnstone, J.F. *et al.* (2016) Changing disturbance regimes, ecological memory,
 372 and forest resilience. *Front. Ecol. Environ.* 14, 369–378
 373 26 Schurman, J.S. *et al.* (2018) Large-scale disturbance legacies and the climate
 374 sensitivity of primary *Picea abies* forests. *Glob. Chang. Biol.* 24, 2169–2181
 375 27 Poorter, L. *et al.* (2021) Multidimensional tropical forest recovery. *Science* (80-
 376). 374, 1370–1376
 377 28 Smith, C.C. *et al.* (2020) Secondary forests offset less than 10% of deforestation-
 378 mediated carbon emissions in the Brazilian Amazon. *Glob. Chang. Biol.* 26,
 379 7006–7020
 380 29 Zhang, X. *et al.* (2019) Ecological contingency in species shifts: Downslope
 381 shifts of woody species under warming climate and land-use change. *Environ.*
 382 *Res. Lett.* 14,
 383 30 Vilà-Cabrera, A. *et al.* (2017) “New Forests” from the Twentieth Century are a
 384 Relevant Contribution for C Storage in the Iberian Peninsula. *Ecosystems* 20,
 385 130–143
 386 31 Wason, J.W. and Dovciak, M. (2017) Tree demography suggests multiple
 387 directions and drivers for species range shifts in mountains of Northeastern
 388 United States. *Glob. Chang. Biol.* 23, 3335–3347
 389 32 Elias, F. *et al.* (2020) Assessing the growth and climate sensitivity of secondary
 390 forests in highly deforested Amazonian landscapes. *Ecology* 101, e02954
 391 33 Vayreda, J. *et al.* (2016) Anthropogenic-driven rapid shifts in tree distribution
 392 lead to increased dominance of broadleaf species. *Glob. Chang. Biol.* 22, 3984–
 393 3995
 394 34 Zellweger, F. *et al.* (2020) Forest microclimate dynamics drive plant responses to
 395 warming. *Science* (80-). 368, 772–775

- 396 35 Hereş, A.-M. *et al.* (2021) Legacies of past forest management determine current
397 responses to severe drought events of conifer species in the Romanian
398 Carpathians. *Sci. Total Environ.* 751, 141851
- 399 36 Ruiz-Benito, P. *et al.* (2012) Large-scale assessment of regeneration and diversity
400 in Mediterranean planted pine forests along ecological gradients. *Divers. Distrib.*
401 18, 1092–1106
- 402 37 Sánchez-Salguero, R. *et al.* (2013) Contrasting vulnerability and resilience to
403 drought-induced decline of densely planted vs. natural rear-edge *Pinus nigra*
404 forests. *For. Ecol. Manage.* 310, 956–967
- 405 38 Hua, F. *et al.* (2022) The biodiversity and ecosystem service contributions and
406 trade-offs of forest restoration approaches. *Science (80-.)*. 4649, 1–28
- 407 39 Berenguer, E. *et al.* (2021) Tracking the impacts of El Niño drought and fire in
408 human-modified Amazonian forests. *Proc. Natl. Acad. Sci.* 118, e2019377118
- 409 40 Camarero, J.J. *et al.* (2021) Differences in temperature sensitivity and drought
410 recovery between natural stands and plantations of conifers are species-specific.
411 *Sci. Total Environ.* 796, 148930
- 412 41 Dieleman, C.M. *et al.* (2020) Wildfire combustion and carbon stocks in the
413 southern Canadian boreal forest: Implications for a warming world. *Glob. Chang.*
414 *Biol.* 26, 6062–6079
- 415 42 Sangüesa-Barreda, G. *et al.* (2015) Past logging, drought and pathogens interact
416 and contribute to forest dieback. *Agric. For. Meteorol.* 208, 85–94
- 417 43 Marqués, L. *et al.* (2018) Last-century forest productivity in a managed dry-edge
418 Scots pine population: The two sides of climate warming: The. *Ecol. Appl.* 28,
419 95–105
- 420 44 Stojanović, M. *et al.* (2017) Forecasting tree growth in coppiced and high forests
421 in the Czech Republic. The legacy of management drives the coming *Quercus*
422 *petraea* climate responses. *For. Ecol. Manage.* 405, 56–68
- 423 45 Maes, S.L. *et al.* (2019) Environmental drivers interactively affect individual tree
424 growth across temperate European forests. *Glob. Chang. Biol.* 25, 201–217
- 425 46 Meyer, P. *et al.* (2022) Management alters drought-induced mortality patterns in
426 European beech (*Fagus sylvatica* L.) forests. *Plant Biol.* DOI: 10.1111/plb.13396
- 427 47 Au, T.F. *et al.* (2022) Younger trees in the upper canopy are more sensitive but
428 also more resilient to drought. *Nat. Clim. Chang.* 12, 1168–1174
- 429 48 Mausolf, K. *et al.* (2018) Legacy effects of land-use modulate tree growth

430 responses to climate extremes. *Oecologia* 187, 825–837

431 49 Alfaro-Sánchez, R. *et al.* (2019) Land use legacies drive higher growth, lower
432 wood density and enhanced climatic sensitivity in recently established forests.
433 *Agric. For. Meteorol.* 276–277,

434 50 Ledig, F.T. (1992) Human Impacts on Genetic Diversity in Forest Ecosystems.
435 *Oikos* 63, 87

436 51 Jump, A.S. *et al.* (2009) Environmental change and the option value of genetic
437 diversity. *Trends Plant Sci.* 14, 51–58

438 52 Sjölund, M.J. and Jump, A.S. (2015) Coppice management of forests impacts
439 spatial genetic structure but not genetic diversity in European beech (*Fagus*
440 *sylvatica* L.). *For. Ecol. Manage.* 336, 65–71

441 53 Boyden, S. *et al.* (2008) Competition among eucalyptus trees depends on genetic
442 variation and resource supply. *Ecology* 89, 2850–2859

443 54 Lindgren, D. (2016) The role of tree breeding in reforestation. *Reforesta* DOI:
444 10.21750/REFOR.1.11.11

445 55 Steinitz, O. *et al.* (2012) Effects of forest plantations on the genetic composition
446 of conspecific native Aleppo pine populations. *Mol. Ecol.* 21, 300–313

447 56 Lind, B.M. *et al.* (2019) Effect of fire and thinning on fine-scale genetic structure
448 and gene flow in fire-suppressed populations of sugar pine (*Pinus lambertiana*
449 Dougl.). *For. Ecol. Manage.* 447, 115–129

450 57 Rajora, O.P. *et al.* (2000) Microsatellite DNA analysis of genetic effects of
451 harvesting in old-growth eastern white pine (*Pinus strobus*) in Ontario, Canada.
452 *Mol. Ecol.* 9, 339–348

453 58 Jia, Y. *et al.* (2020) Evolutionary legacy of a forest plantation tree species (*Pinus*
454 *armandii*): Implications for widespread afforestation. *Evol. Appl.* 13, 2646–2662

455 59 Unger, G.M. *et al.* (2016) Assessing early fitness consequences of exotic gene
456 flow in the wild: a field study with Iberian pine relicts. *Evol. Appl.* 9, 367–380

457 60 Jordan, R. *et al.* (2019) How well do revegetation plantings capture genetic
458 diversity? *Biol. Lett.* 15, 20190460

459 61 Strahan, R.T. *et al.* (2016) Shifts in community-level traits and functional
460 diversity in a mixed conifer forest: a legacy of land-use change. *J. Appl. Ecol.* 53,
461 1755–1765

462 62 Silva, C.A. da *et al.* (2020) Fine root-arbuscular mycorrhizal fungi interaction in
463 Tropical Montane Forests: Effects of cover modifications and season. *For. Ecol.*

464 *Manage.* 476, 118478

465 63 Guerrieri, R. *et al.* (2021) Land-use legacies influence tree water-use efficiency
466 and nitrogen availability in recently established European forests. *Funct. Ecol.*
467 DOI: 10.1111/1365-2435.13787

468 64 Correia, M. *et al.* (2021) Land-use history alters the diversity, community
469 composition and interaction networks of ectomycorrhizal fungi in beech forests.
470 *J. Ecol.* 109, 2856–2870

471 65 Netherer, S. *et al.* (2019) Acute Drought Is an Important Driver of Bark Beetle
472 Infestation in Austrian Norway Spruce Stands. *Front. For. Glob. Chang.* 2,
473 66 Umaña, M.N. *et al.* (2019) Dry conditions and disturbance promote liana
474 seedling survival and abundance. *Ecology* 100,

475 67 Robert, L.-E. *et al.* (2018) Landscape host abundance and configuration regulate
476 periodic outbreak behavior in spruce budworm *Choristoneura fumiferana*.
477 *Ecography (Cop.)*. 41, 1556–1571

478 68 Giuggiola, A. *et al.* (2018) Competition for water in a xeric forest ecosystem –
479 Effects of understory removal on soil micro-climate, growth and physiology of
480 dominant Scots pine trees. *For. Ecol. Manage.* 409, 241–249

481 69 Sangüesa-Barreda, G. *et al.* (2015) Reduced growth sensitivity to climate in bark-
482 beetle infested Aleppo pines: Connecting climatic and biotic drivers of forest
483 dieback. *For. Ecol. Manage.* 357, 126–137

484 70 Slaughter, G.W. and Rizzo, D.M. (1999) Past forest management promoted root
485 disease in Yosemite Valley. *Calif. Agric.* 53, 17–24

486 71 Meentemeyer, R.K. *et al.* (2008) Influence of land-cover change on the spread of
487 an invasive forest pathogen. *Ecol. Appl.* 18, 159–171

488 72 Burgess, T.I. *et al.* (2022) Anthropogenic Disturbances and the Emergence of
489 Native Diseases: a Threat to Forest Health. *Curr. For. Reports* 8, 111–123

490 73 Nakajima, H. (2019) Region-wide mass mortality of Japanese oak due to
491 ambrosia beetle infestation: Mortality factors and change in oak abundance. *For.*
492 *Ecol. Manage.* 449, 117468

493 74 Schulte to Bühne, H. *et al.* (2021) Improving Predictions of Climate Change–
494 Land Use Change Interactions. *Trends Ecol. Evol.* 36, 29–38

495 75 Perring, M.P. *et al.* (2016) Global environmental change effects on ecosystems:
496 the importance of land-use legacies. *Glob. Chang. Biol.* 22, 1361–1371

497 76 Lines, E.R. *et al.* (2022) The shape of trees: reimagining forest ecology in three

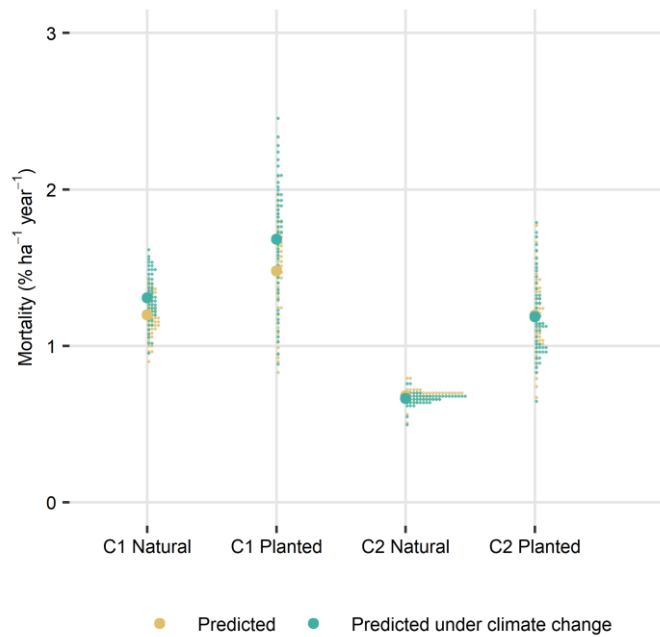
498 dimensions with remote sensing. *J. Ecol.* DOI: 10.1111/1365-2745.13944
499 77 Seijo, F. *et al.* (2018) Traditional fire use impact in the aboveground carbon stock
500 of the chestnut forests of Central Spain and its implications for prescribed
501 burning. *Sci. Total Environ.* 625, 1405–1414
502 78 Infante-Amate, J. *et al.* (2022) From woodfuel to industrial wood: A socio-
503 metabolic reading of the forest transition in Spain (1860–2010). *Ecol. Econ.* 201,
504 107548
505 79 Villar-Salvador, P. (2016) Restoration of Spanish pine plantations: A main
506 challenge for the 21st century. *Reforesta* DOI: 10.21750/REFOR.1.04.4
507 80 Vilà-Cabrera, A. *et al.* (2011) Structural and climatic determinants of
508 demographic rates of Scots pine forests across the Iberian Peninsula. *Ecol. Appl.*
509 21, 1162–1172
510 81 Vadell, E. *et al.* (2016) Large-scale reforestation and afforestation policy in
511 Spain: A historical review of its underlying ecological, socioeconomic and
512 political dynamics. *Land use policy* 55, 37–48
513 82 Neumann, M. *et al.* (2017) Climate variability drives recent tree mortality in
514 Europe. *Glob. Chang. Biol.* 23, 4788–4797
515 83 Ruiz-Benito, P. *et al.* (2017) Functional diversity underlies demographic
516 responses to environmental variation in European forests. *Glob. Ecol. Biogeogr.*
517 26, 128–141
518
519
520
521
522
523
524
525
526
527
528
529
530
531

532 **Box 1. Anthropogenic land-use legacies contribute to large-scale tree mortality risk**
533 **in Iberian Scots pine forests**

534 Scots pine (*Pinus sylvestris* L.) is a widespread species that reaches its driest range edge
535 in the Iberian Peninsula. In this region, the distribution, structure, and composition of
536 Scots pine forests have been historically altered by management practices and tree
537 planting for wood extraction and erosion control. The cessation of these land-uses
538 occurred at fast rates over the 20th century [79]. Previous studies have shown how stand-
539 level structural characteristics influence current forest demography under drought stress
540 [36,80], but the prevalence of anthropogenic land-use legacies exacerbating drought-
541 related Scots pine mortality has not been explicitly assessed. We used large-scale datasets
542 on historical anthropogenic land-use and forest inventory to apply the proposed risk-
543 based framework (Fig. 2, Key Figure) in Scots pine-dominated forests. We first
544 characterised forest inventory plots according to their natural or planted origin (i.e., stand
545 origin, see Supplementary Information). We considered stand origin as an indicator of
546 anthropogenic land-use legacies derived from alterations in the species' distribution and
547 stand composition. We then identified two main forest structural typologies across
548 natural- and planted-origin stands, one associated with high tree height diversity and
549 sapling abundance ('C1'; Fig. S1-S2), and the other one associated with high tree density,
550 basal area, height, height-to-diameter ratio, and height homogeneity ('C2'; Fig. S1-S2).
551 Tree diameter distribution was similar between both structural typologies (Fig. S2).
552 Within forest structural typologies, planted-origin stands (C1-planted and C2-planted)
553 displayed higher tree density, lower tree size, higher structural homogeneity, and greater
554 tree growth, but lower abundance of saplings than forests with natural origin (C1-natural,
555 C2-natural) (Fig. S3-S4), in accordance with previous studies [79,81]. We considered
556 these structural characteristics in plantations as indicator of anthropogenic land-use
557 legacies derived from alterations in forest structure and stand age reduction. We
558 hypothesised that anthropogenic land-use legacies in plantations, emerged following the
559 cessation of high-intensity land-uses (Fig. 2A, main text), would exacerbate drought-
560 related tree mortality risk (Fig. 2B-C, main text, Table 1). Tree mortality was overall
561 higher in planted- than in natural-origin stands, suggesting that persisting anthropogenic
562 land-use legacies following alterations in forest distribution, structure, and composition
563 can increase overall mortality risk to forests (Fig. I, Fig. 2B-C). This pattern remained
564 consistent within each forest structural typology (Fig. I). Under a climate change scenario
565 that assumed 20% reduction in water availability and 20% increase in drought severity,

566 mortality risk increased in C1 forest structural typology, being this increase 44% higher
567 in plantations (Fig. I; SI). The higher mortality in plantations in C1 structural typology is
568 consistent with increased climate change-related overall risk to forests. Increased
569 exposure may derive from the altered species distribution towards the expansion of the
570 species range, which may result in a greater chance to encounter drought stress out of the
571 species' climatic tolerance [36]. Higher sensitivity may derive from the altered forest
572 structure towards structurally homogeneous, denser, and younger stands, exacerbating
573 conspecific negative density-dependence in response to drought stress [37]. Decreased
574 adaptive capacity may derive from the alteration in composition towards increased
575 abundance of younger trees which have functional traits with lower capacity to deal with
576 drought stress [47]. On the contrary, climate change-type drought did not result in
577 increased tree mortality in C2 forest structural typology (Fig. I), suggesting that other
578 drivers than drought stress, such as long-term effects of altered stand genetic properties
579 [60] or warmer temperatures in combination with tree antagonist organisms (e.g.,
580 pathogens)[72], might explain the larger mortality trends observed in plantations in this
581 structural typology. This study case demonstrates that anthropogenic land-use legacies
582 persisting in plantations can elevate the risk of climate change-related stressors to forests.
583 At the same time, it also highlights that the proposed framework is a useful approach to
584 identify hotspot forest types at critical risk and its potential drivers, providing thus
585 opportunities to refine our predictive understanding of large-scale forest dynamics, and
586 develop and implement mitigation and adaptation strategies during coming decades.
587 *Note: data and code used can be found in Zenodo repository*
588 *(<https://doi.org/10.5281/zenodo.7120609>).* *Data is also available at*
589 *[https://www.miteco.gob.es/es/biodiversidad/temas/inventarios-nacionales/inventario-](https://www.miteco.gob.es/es/biodiversidad/temas/inventarios-nacionales/inventario-forestal-nacional/default.aspx)*
590 *[forestal-nacional/default.aspx](https://www.miteco.gob.es/es/biodiversidad/temas/inventarios-nacionales/inventario-forestal-nacional/default.aspx).*

591
592
593
594
595
596
597



598

599

600 **Figure I (Box 1). Planted-origin stands have higher tree mortality risk than natural-**
 601 **origin stands in Iberian Scots pine forests.** Quantile dotplot representing model-
 602 predicted tree mortality (yellow) and model-predicted tree mortality under a climate
 603 change scenario (green) assuming a 20% reduction in water availability (WAI) and 20%
 604 increase in drought conditions (SPEI) for each forest type defined by stand origin (planted
 605 or natural) and forest structural typology (C1 and C2). *Note: Mortality was estimated in*
 606 *terms of percentage of dead basal area at the plot level. Large dots represent the median*
 607 *values and small dots represents approximately a 2% (1/50) chance. The climate change*
 608 *scenario is based on the IPCC predictions for the Mediterranean basin [19].*

609

610

611

612

613

614

615

616

617

618

619

620 **Glossary**

621

622 **Anthropogenic land-use:** human-driven exploitation of land with impacts on forest
623 ecosystems, including complete tree cover loss due to land-cover changes, partial tree
624 cover loss due to forest management practices, and tree cover increase due to tree planting
625 (reforestation and afforestation).

626 **Anthropogenic land-use legacy:** persistent effect on forest properties following the
627 cessation of anthropogenic land-uses.

628 **Impacts of climate change on forests:** abrupt or progressive climate-related shifts in
629 forest dynamics and functioning, including forest die-off events, megafires, and forest
630 degradation.

631 **Conspecific negative density-dependence:** population-level processes that occur when
632 population growth is negatively influenced (e.g., tree mortality) by population density.

633 **Forest properties:** Genetic-, population-, community-, or ecosystem-level factors and
634 processes that underlie forest dynamics.

635 **Great acceleration:** continuous and exponential growth rate of human activities since
636 mid-20th century that are accompanied by substantial changes in ecosystems.

637 **Human-modified forests:** forest ecosystems shaped by past and current anthropogenic
638 land-uses.

639 **Land-use cessation:** temporary or permanent stopping of a given land-use.

640 **Resilience to climate change:** the ability of forest ecosystems to cope with- and adapt to
641 climate change stressors.

642 **Secondary forests:** forests that have re-established after the complete or partial loss of
643 the original tree cover.

644

645

646

647

648

649

650

651

652

Table 1. Hypotheses regarding detrimental anthropogenic land-use legacies. Each specific hypothesis is framed considering the interaction outcomes between climatic stressors and contemporary forest properties resulting from anthropogenic land use-driven alteration in forest distribution, structure, and composition.

Altered forest pattern	Affected forest property	Cause	Hypotheses on detrimental anthropogenic land-use legacies	Refs
Distribution	Species range re-expansion	Land-cover change	Secondary forest expansion towards stressful climatic conditions relative to species climatic niches	[16,28–32]
		Management practice	Forest recovery occurs under an unfavourable climatic context relative to species niches	[31,33,34]
		Plantation	Forest plantations decline and are highly impacted at the edge or out of the species natural range	[24,35–38]
Structure	Stand-level tree size distribution	Land-cover change, management practice	Climatic stressors exacerbate conspecific negative density-dependence processes and fire impact in secondary forests	[32,39]
		Plantation	Climatic stressors exacerbate conspecific negative density-dependence processes and fire impact	[20,24,36,37,40]
Composition	Within-species phenotypic, functional, and genetic properties	Land-cover change	Soil legacies result in within-species functional shifts towards tree phenotypes with reduced resilience capacity to increased climatic stressors	[48,49]
		Management practice, plantation	Shifts toward tree phenotypes with reduced resilience capacity to increased climatic stressors	[12,21,39,41–47]
		Management practice, plantation	Reduced genetic diversity and genetic-induced functional changes reduce the resilience capacity of populations	[24,52–60]
	Interspecific functional traits and interactions	Land-cover change, management practice, plantation	Novel tree functional assemblages with reduced resilience capacity to increased climatic stressors	[33,39,61]
			Increased forest vulnerability as a result of shifts in species interactions involving mutualisms and antagonisms	[42,48,54,62–73]

Figure 1. Examples of anthropogenic land-use legacies emerging from alterations in forest distribution, structure, and composition in Iberian forests. Description of photos: (A) Forest landscape with stands of reforested and afforested Stone pines (*Pinus pinea*). Pines are today declining in response to drought and biotic stressors (photo: F. Vilà-Carbonell. Maresme, Catalonia, Spain). (B) Over-dense and structurally homogeneous Scots pine (*Pinus sylvestris*) abandoned plantation in a drought stressed area (photo: A. Vilà-Cabrera. Parque Natural de la Sierra Norte de Guadalajara, Spain). (C) Pollarded beech trees (*Fagus sylvatica*) today overgrown and with altered phenotype because of the cessation of the traditional management practice (photo: J. Astigarraga. Oñati, Gipuzkoa, Spain). (D) Secondary beech (*Fagus sylvatica*) forest established following the cessation of agricultural use (photo: J. Astigarraga, Oñati, Gipuzkoa, Spain). (E) *Pinus radiata* plantation surrounding a native stand of sessile oak (*Quercus petraea*; photo: J. Astigarraga. Oñati, Gipuzkoa, Spain). (F) Abandoned plantation of *Eucalyptus nitens* with regenerating native oaks (photo: J. Astigarraga. Oñati, Gipuzkoa, Spain).

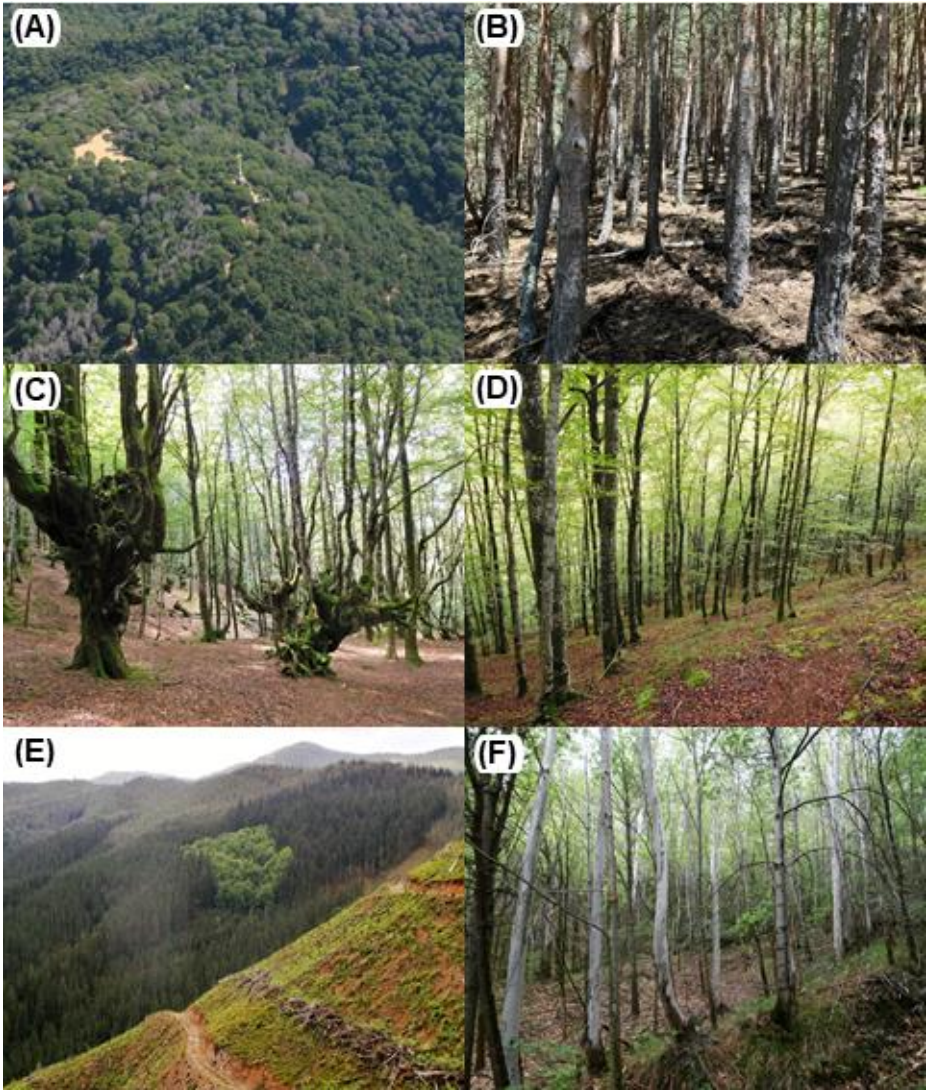


Figure 2 (Key figure). Risk-based framework using the concept of anthropogenic land-use legacy. We adapt the IPCC framework [19] to forest ecosystems considering the impact risk to forests as the potential response to a climate-related hazard (e.g., increased drought-induced mortality [82]). Risks to forests from climate-related hazards can be exacerbated by anthropogenic land-use legacies due to the interaction between exposure (ecological settings that expose forests to hazards, e.g., drought stress at the driest edge of species distributions [18]), sensitivity (the degree to which a forest ecosystem is affected, e.g., increased tree mortality due to exacerbated conspecific negative density-dependence [12]), and adaptive capacity (the capacity of forest trees, populations, and communities to counteract and cope with climatic stressors due to functional and genetic properties and interspecific interactions, e.g., where reduced functional diversity reduces the capacity of forest ecosystems to cope with climate change [83]). In this conceptual figure, a hypothetical region of pine-dominated forests is represented. Four- (in A) and two forest types (in B and C) are represented for simplification: intensively managed pine plantation, pine plantation following cessation of intensive land-use, sustainable managed natural pine forest, natural forest following sustainable land-use cessation. (A) Anthropogenic land-uses shape the main patterns in forest distribution, structure, and composition (i.e., human-modified forested areas). High-impact anthropogenic land-use legacies may derive after the cessation of high-intensity land-uses, e.g., high-density, and structurally homogeneous pine plantations for wood extraction. The characterisation, identification, and contextualisation of high impact anthropogenic land-use legacies in regions dominated by human-modified forests is of key urgency as these forests may be highly vulnerable to climatic stressors. (B) Testing interaction effects between anthropogenic land-use legacies and climatic stressors on forest ecosystem responses. The prevalence of increased detrimental anthropogenic land-use legacies leads to increased climate-related impacts on forests. (C) Linking detrimental anthropogenic land-use legacies in forest properties (derived from changes in forest distribution, structure, and composition) with climate-related risk components can support the predictive understanding of forest dynamics in regions dominated by human-modified forests. *Note: vector icons are from <https://www.flaticon.com>*

